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Effect of indigo dye effluent on the growth, biomass production and phenotypic plasticity of *Scenedesmus quadricauda* (Chlorococcales)

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**ABSTRACT**

The effect of indigo dye effluent on the freshwater microalga *Scenedesmus quadricauda* ABU12 was investigated under controlled laboratory conditions. The microalga was exposed to different concentrations of the effluent obtained by diluting the dye effluent from 100 to 175 times in bold basal medium (BBM). The growth rate of the microalga decreased as indigo dye effluent concentration increased \((p < 0.05)\). The EC\(_{50}\) was found to be 166 dilution factor of the effluent. Chlorophyll \(a\), cell density and dry weight production as biomarkers were negatively affected by high indigo dye effluent concentration, their levels were higher at low effluent concentrations \((p < 0.05)\). Changes in coenobia size significantly correlated with the dye effluent concentration. A shift from large to small coenobia with increasing indigo dye effluent concentration was obtained. We conclude that even at low concentrations; effluents from textile industrial processes that use indigo dye are capable of significantly reducing the growth and biomass production, in addition to altering the morphological characteristics of the freshwater microalga *S. quadricauda*. The systematic reduction in the number of cells per coenobium observed in this study further confirms that environmental stress affects coenobium structure in the genus *Scenedesmus*, which means it can be considered an important biomarker for toxicity testing.

**Key words**: freshwater, microalgae, morphological variation, pollution, textile waste.

**INTRODUCTION**

Textile industrial processes consume large amounts of water and use toxic products. They contribute to the degradation of the environment by releasing untreated dye effluents into neighboring aquatic ecosystems in the form of sludge. The final disposal of this sludge remains a challenge, and its ecotoxicological assessment is important for minimizing its environmental impacts. Even when treated, conventional waste treatment techniques do not completely decolorize and detoxify the dye effluents (Puvaneswari et al. 2006).

Indigo dye is part of the numerous marketed organic colorants used for coloration of textiles, paper, leather, plastic and for specialized applications such as food, drug, cosmetic and photochemical productions (Zollinger 1987, Novotny et al. 2006). Textile effluents containing indigo dye and other dye types make water toxic (Robinson et al. 2001) and unfit for human and
animal consumption, and cause imbalance within different aquatic ecosystem food chains (Rocha 1992). In recent years various researchers have identified mutagenic effects of textile samples and waste water of the textile industry (Knasmüller et al. 1993, Jäger 1998, Mathur et al. 2012). Their findings show that dyes used for textile finishing are mainly responsible for the mutagenic effects observed. Although information on textile dye toxicity to microalgae remain scanty, Rannug et al. (1992), Mathur et al. (2005a, b), and Mathur and Bhatnagar (2007) were able to show that indigo dye had mutagenic effect on the bacterium Salmonella typhimurium.

The effluents of textile dyeing and printing industries are complex in nature, rich in dissolved and suspended solids, organic compounds, heavy metals and have high pH. The composition of textile effluents determines the extent to which they pollute rivers (Srivastava et al. 1994), affect soil characteristics (Raj et al. 1997), and affect plants (Khandelwal 1996, Raj et al. 1997). Hence, we hypothesize that the complex nature of indigo dye effluent would be responsible for the overall toxicity observed and not just the dye component alone. This is because the presence of metals like cadmium, chromium and copper in textile effluents can exhibit various interactive toxic effects on aquatic biota (Mathur et al. 2012). The final toxic effect of the indigo dye effluent may be synergistic, additive or antagonistic as a function of the different effluent components that make up the effluent and not the dye in isolation.

Scenedesmus chosen for this study is cosmopolitan and can be found in a wide array of aquatic ecosystems ranging from oligo-, meso-, eutrophic and metal polluted waters. Scenedesmus quadricauda forms miniature colonies termed autocolonies or coenobia, in which groups of 4 to 8 cells are connected together in a row. The morphological form i.e. the number of cells per coenobium of Scenedesmus varies depending on environmental conditions and the physiological state. Under environmental stress, it may fail to produce such colonies (van Den Hoek et al. 1995, Lurling and van Donk 1999, Lombardi et al. 2007). These changes in phenotypic plasticity are considered a response to environmental stress (Lurling and Beekman 2002, Peña-Castro et al. 2004).

To the best of our knowledge, nothing is known about the effect of indigo dye effluents on freshwater microalgae. Due to the extent to which the dye is extensively used today in the textile and allied industries, understanding the effects of the indigo dye effluents is very important and requires investigations. Detection and quantification of the morphological response is easy and inexpensive, and thus very relevant in environmental monitoring (Lombardi et al. 2007). This study was aimed at determining the effect of indigo dye effluent on the growth, biomass production and phenotypic plasticity of S. quadricauda.

**MATERIALS AND METHODS**

The Scenedesmus quadricauda ABU12 strain used in this study was isolated from a freshwater pond in Zaria, Nigeria. It was maintained in the laboratory with Bold Basal Medium (BBM) at 23+2°C and under continuous illumination at 150 umol photons m⁻² s⁻¹ using cool white fluorescent lamps. Culture media sterilization was performed through autoclaving at 121°C for 15 min, and the media were left to stand and equilibrate for 24 h before use. Sterile techniques were maintained throughout the experimental period.

Dye effluent samples were collected from local textile industries located in Zaria, Nigeria. In order to confirm the identity of the dye used in the local dye industries, we purchased raw indigo dye from Emmybiz Dye and Chemicals International Limited, Sabon Gari, Zaria, Nigeria. This dye company supplies the local textile industries in Zaria with indigo dye. The identification of the raw dye and the dye in the effluent were done using spectroscopic methods by observing its absorption at different
wavelengths with the aid of a SpectrumLab 7525 UV-VIS spectrophotometer (B. Bran Scientific and Instrument Company, England). In addition, Fourier Transform Infrared Analysis (FTIR) was used to determine the different functional groups of the dye as a way of confirming its identification. About 1 mg was weighed in a small agate mortar with a drop of nujol and Kbr. The mull was then pressed between the flat plates of sodium chloride. Before running a scan on each sample type, a background scan was carried out. All scanning was done from 4000 to 625 cm⁻¹ (Williams and Fleming 2005) using a Shimadzu FT-IR Model 8400s spectrophotometer (Shimadzu, Japan).

The effluent was characterized for physical and chemical properties. pH and electrical conductivity levels were determined using HANNA pH/EC/Temp meter model 210, and turbidity was determined spectrophotometrically. Biochemical oxygen demand, dissolved oxygen and chemical oxygen demand were determined using the Azide Modification of the Winkler method as described in APHA (1998). Nitrate nitrogen, phosphate phosphorus and total sulphate were determined using the spectrophotometric procedures described in APHA (1998). Cadmium, chromium and copper concentrations were determined using flame atomic absorption spectroscopy (FAAS) (Shimadzu Model AA6800, Shimadzu, Japan). The physicochemical characteristics of the indigo dye effluent are presented in Table I.

Preliminary screening was done with the microalga for its ability to grow in the dye effluent. Exponential growing cultures were inoculated into different effluent concentrations (10 to 100%) in BBM in 200 mL using Erlenmeyer flasks. The cultures were grown for 7 days under controlled laboratory conditions listed above. Prior to inoculation, pH was adjusted to 7±0.3. This preliminary test showed no visible growth was observed under these conditions. Therefore, further dilutions of the original effluent were made to a factor range of 100 to 175 times. The control treatment with no dye effluent addition was represented as normal BBM. The microalgae was acclimated to the BBM using semi-continuous batch system. When the microalgae was at mid exponential growth phase, dilutions were made by re-inoculating at a 10% v/v until a steady state physiological growth status was obtained. This is the stage where the growth rate was statistically the same and reflecting the growth condition to which *S. quadricauda* was exposed. It is this exponential phase growing *S. quadricauda* that was exposed to the different dye dilution treatments to give an initial cell density of 10⁴ cells mL⁻¹ and monitored for a period of 4 days (0-96 h). All experiments were carried out in triplicates.

The growth of the cultures was monitored by taking aliquots from the cultures on a daily basis for absorbance and cell density measurements. Specific growth was determined using the natural logarithmic transformed regression growth curve. Cell counts were determined by using direct microscopic counts with a Naubauer Brightline Haemocytometer. Dry weight was determined gravimetrically by using pre-weighed Sartorius 0.22 µM pore size membrane filters. The extraction and analysis of chlorophyll *a* was carried out in line with the procedures of Shoaf and Lium.

### Table I

<table>
<thead>
<tr>
<th>Physicochemical parameters</th>
<th>Value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Nitrate nitrogen (mg L⁻¹)</td>
<td>0.001</td>
</tr>
<tr>
<td>Phosphate phosphorus (mg L⁻¹)</td>
<td>50.500</td>
</tr>
<tr>
<td>Dissolved oxygen (mg L⁻¹)</td>
<td>19.400</td>
</tr>
<tr>
<td>Chemical oxygen demand (mg L⁻¹)</td>
<td>24.000</td>
</tr>
<tr>
<td>Electrical conductivity</td>
<td>19.710</td>
</tr>
<tr>
<td>Temperature (°C)</td>
<td>28.200</td>
</tr>
<tr>
<td>pH</td>
<td>6.700</td>
</tr>
<tr>
<td>Biological oxygen demand (mg L⁻¹)</td>
<td>0.000</td>
</tr>
<tr>
<td>Cu (mg L⁻¹)</td>
<td>0.054</td>
</tr>
<tr>
<td>Cr (mg L⁻¹)</td>
<td>0.183</td>
</tr>
<tr>
<td>Cd (mg L⁻¹)</td>
<td>0.003</td>
</tr>
</tbody>
</table>
Pigment quantification was done using the equations provided by Németh (1998). Number of cells per coenobium was determined under the microscope to see if there was any relationship between the level of the effluent and the number of cells per coenobium.

The data obtained from this study was subjected to Levene’s homogeneity variance test. A two-way ANOVA was used to test for significant differences between response parameters (growth, biomass production and phenotypic plasticity) of the microalga under the different indigo dye effluent concentrations. Tukey’s HSD post hoc test was used to separate the significantly different means from the ANOVA computations. The relationship between the different parameters measured and the treatment conditions was determined using a correlations based principal components analysis (PCA). Homogeneity, ANOVA and Tukey’s HSD tests were performed using Statistica for windows, while PCA was obtained using PAST statistical software. Effective concentration 50 (EC50) was determined using the EPA Probit Analysis program version 1.5 for windows. All analyses were done at 5% significance level.

RESULTS

This research focused on the effect of indigo dye effluents on the growth, biomass production and morphological plasticity (number of cells per coenobium) of S. quadricauda. The growth rate of the microalga was highest in the control but decreased with increasing effluent concentration i.e. decreasing dilution rate (Table II). Furthermore, the EC50 concentration was 165 dilution of the original dye effluent. The control supported the highest chlorophyll a production and cell density among all treatments (Figs. 1 and 2). Fig. 3 shows that dry weight production in the control increased over time, while the response in the different treatment was somewhat different. The lowest dye concentration had similar dry weight production to that observed in the control, while at higher dye concentrations variable biomass production was observed over the study period. Differences in cell density, dry weight and pigment concentration over time with increasing concentration of the dye treatments were obtained (ANOVA, \( p < 0.05 \); Table III).

<table>
<thead>
<tr>
<th>Treatment Specific growth rate (d(^{-1}))</th>
<th>Control</th>
<th>0.93</th>
</tr>
</thead>
<tbody>
<tr>
<td>175</td>
<td>0.87</td>
<td></td>
</tr>
<tr>
<td>150</td>
<td>0.47</td>
<td></td>
</tr>
<tr>
<td>125</td>
<td>0.40</td>
<td></td>
</tr>
<tr>
<td>100</td>
<td>0.05</td>
<td></td>
</tr>
<tr>
<td>EC50</td>
<td>165.52</td>
<td></td>
</tr>
</tbody>
</table>

**TABLE III**

ANOVA summary results for the various parameters of Scenedesmus quadricauda analyzed after exposure to different dye effluent concentrations. N.B: where \( P \) value is less than 0.05, the ANOVA \( F \) value is significant.

<table>
<thead>
<tr>
<th>Parameters</th>
<th>Treatment F value</th>
<th>Treatment P value</th>
<th>Time F value</th>
<th>Time P value</th>
<th>Treatment vs. Time F value</th>
<th>Treatment vs. Time P value</th>
</tr>
</thead>
<tbody>
<tr>
<td>Cell count</td>
<td>7.1833</td>
<td>0.000186</td>
<td>24.1</td>
<td>0</td>
<td>2.153</td>
<td>0.035</td>
</tr>
<tr>
<td>Dry weight</td>
<td>2.743</td>
<td>0.0417</td>
<td>13.953</td>
<td>0</td>
<td>2.432</td>
<td>0.0176</td>
</tr>
<tr>
<td>Chlorophyll a</td>
<td>1.371</td>
<td>0.261</td>
<td>1.203</td>
<td>0.32</td>
<td>1.313</td>
<td>0.249</td>
</tr>
<tr>
<td>1 cell per coenobium</td>
<td>38.188</td>
<td>0</td>
<td>4.283</td>
<td>0.0103</td>
<td>1.0436</td>
<td>0.431</td>
</tr>
<tr>
<td>2 cell per coenobium</td>
<td>2.025</td>
<td>0.109</td>
<td>3.531</td>
<td>0.023</td>
<td>1.512</td>
<td>0.16</td>
</tr>
<tr>
<td>3 cell per coenobium</td>
<td>3.924</td>
<td>0.008</td>
<td>1.026</td>
<td>0.391</td>
<td>1.146</td>
<td>0.353</td>
</tr>
<tr>
<td>4 cell per coenobium</td>
<td>41.158</td>
<td>0</td>
<td>1.61</td>
<td>0.202</td>
<td>0.69</td>
<td>0.753</td>
</tr>
</tbody>
</table>

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Fig. 1 - Chlorophyll a concentration (mg L\(^{-1}\)) of *Scenedesmus quadricauda* exposed to different indigo dye effluent concentrations.

Fig. 2 - Cell density (cells mL\(^{-1}\)) of *Scenedesmus quadricauda* as a function of different indigo dye effluent concentrations.
Results related to the phenotypic plasticity of *S. quadricauda* can be seen in Figs. 4 and 5. The higher the concentration of dye effluent, the higher the numbers of single cells observed. However, more 4 celled organisms per coenobium were obtained at lower dye effluent concentrations. The changes in the number of single cells, 3 cells and 4 cells produced per coenobium were significantly different between the treatments ($p < 0.05$). The production of 2 cells per coenobium was only significantly different over time ($p < 0.05$), while no significant interaction between the exposure time and effluent concentration ($p > 0.05$) was observed.

Principal components analysis showed that lower dye concentrations (i.e. higher dilution rates) were positively ($p < 0.05$) associated with dry weight production, cell density and high (3 and 4) number of cells per coenobium (Fig. 6). Single cells and 2 cells per coenobium showed a negative association with low dye concentration. Exposure time was positively ($p < 0.05$) correlated with cell density, 4 cells per coenobium and dry weight production.

**DISCUSSION**

From the results of this study, it can be observed that indigo dye effluent is toxic to *S. quadricauda*, having negative effects on growth, biomass production and the number of cells per coenobium. This agrees with the findings of other studies (Mathur et al. 2012). The indigo dye effluent was rich in dissolved and suspended solids and heavy metals. These components including the dye itself can be implicated for the reduction in growth and biomass productivity of the microalga with increasing concentrations of the dye. Due to the fact that it is not always possible to analyze the ecotoxicological effects of individual compounds found in the dye effluent, the effect of the complex effluent mixture is what microalgae in aquatic ecosystems are exposed to. For example, the Cd like any of the other heavy metal reported in this study in the effluent has been reported to exhibit negative effects on the growth and biomass production of the freshwater microalga *Chlorella vulgaris* at environmentally relevant concentrations.
The toxicity of indigo dye effluent to green algae

Fig. 4 - Percentage of total production of single cells (a) and 2 cells per coenobium (b) of Scenedesmus quadricauda exposed to different indigo dye effluent concentrations.

Fig. 5 - Percentage of total of 3 cells per coenobium (a) and 4 cells per coenobium (b) of Scenedesmus quadricauda as a function of different indigo dye effluent concentrations.

(Chia et al. 2013a, b). These components are capable of interacting in an additive, synergistic or even antagonistic way to affect the growth, biomass production and biochemical composition of microalgae (Chia et al. 2013a).

At very low concentrations the effect of the dye effluents was minimal, which means that with proper comprehensive pretreatment before discharge into aquatic ecosystems, their acute effect could be significantly reduced. Although not much is known about the effect of indigo dye effluent on microalgae, reports on higher plants show that reactive dyes can be carcinogenic and mutagenic (Puvaneswari et al. 2006). Our results agree with those of other authors that show that effluents from textile industries even after treatment can remain toxic, before their eventual release into the environment (Umbezeiro et al. 2004, Malachová et al. 2006, Alves de Lima et al. 2007, Bergsten-Torralba et al. 2009). Lim et al. (2010) observed that the growth of C. vulgaris decreased with increasing concentrations of textile wastewater. This agrees with the findings of our study where the growth and biomass accumulation of S. quadricauda decreased with increasing dye effluent concentration. However, in our study the toxicity of indigo dye effluent was much higher than that obtained by Lim et al. (2010). In addition, Rosa et al. (2007) were
Fig. 6 - Principal components analysis (PCA) biplot showing the correlation between different growth, biomass and morphological parameters of *Scenedesmus quadricauda* exposed to different indigo dye effluent concentrations.

...able to show that fresh textile sludge inhibited the growth and biomass production of *S. subspicatus*. Indigo as a dye which forms the principal component of the textile effluent used in this study was reported by Cheriaa et al. (2009) to inhibit the growth and biomass production of the freshwater microalga *Chlorella*. Similar to our findings, these authors showed that low dye concentrations were capable of negatively impacting the physiology of the microalga. Furthermore, weight reduction as another biomarker, has been observed to be a common response of freshwater microalgae to the toxic effect of textile effluents due to reduced photosynthesis and cell division (Roy et al. 2010).

The use of biomass production and growth are the usual biomarkers for phytotoxicity evaluation. However, morphological variations also present effects of pollutants on microalgal physiology and can serve as important biomarkers. Morphological adaptations of microalgae have potential applications in monitoring and in ecotoxicology related investigations (Lombardi et al. 2007). Here we observed that the changes in number of cells per coenobium also known as phenotypic plasticity demonstrated significant correlations with the indigo dye effluent concentrations to which the microalga was exposed. Significant physiological effects can be deduced from this observation as a shift from high to low number of cells per coenobium with increasing indigo dye effluent concentration. This finding implies that the changes in coenobium structure observed for *S. quadricauda* can be used as an important biomarker for ecotoxicological investigations of indigo dye effluents, heavy metals and other pollutants.

Several mechanisms have been suggested to be responsible for the formation of coenobium in microalgae (Pickett-Heaps 1975). Nevertheless, the mechanisms involved in coenobium size reduction are unclear. Although, the plasticity findings in this study are similar to results obtained for *S. acuminatus* and *S. incrassulatus* exposed to toxic concentrations of ionic copper by Lombardi et al. (2007) and Peña-Castro et al. (2004), respectively, a few exceptions were observed in this study. They did not report 3 cells per coenobium in *S. acuminatus* whereas in *S. quadricauda* under the present experimental conditions, this morphological variation was...
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observed. Furthermore, the highest number of cells per coenobium observed in this study was 4 cells, which is unlike what was obtained by Lombardi et al. (2007) with even 8 cells per coenobium occurring at low Cu++ concentrations. It is possible to relate our findings to Lombardi et al. (2007) because of the evidence of heavy metal contamination of the effluent we tested. However, the differences between our study and theirs may be due to the fact that the microalgae investigated in both studies are of two different species, and also due to the nature of the stress to which they were exposed.

Conclusions can be made that these results show that indigo dye effluents exhibit acute toxicity effects on the microalga *S. quadricauda* at the concentrations tested in this study. Therefore, there is the risk of disrupting the primary aquatic food chain when untreated indigo dye effluents are discharged into aquatic systems. Recommendations are made that long term studies be carried out to show the chronic effect of the dye on the growth, biomass production, biochemical composition and genotoxicity for freshwater microalgae because this information is currently lacking. The systematic reduction in the number of cells per coenobium observed in this study further confirms that environmental stress can affect coenobium structure in *Scenedesmus* and that this feature offers a sensitive biomarker for toxicity testing.

REFERENCES


