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Marine habitats map of "Isla del Caño", Costa Rica, comparing Quickbird and Hymap images
classification results

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Maps are considered a great research, planning, management and monitoring tool because they can provide the extent and distribution of benthic habitats and can improve conservation efforts (Mumby et al. 1997, Guzmán et al. 2004). The use of different spatial resolution satellite images and aerial photography has proven to be relatively effective in mapping (Mumby et al. 1997). Hyperspectral sensors, like Hymap (18 optical bands, 16m pixel size), increase our capability to detect narrow spectral bands that can be used for discriminating benthic communities of low and moderate mapping complexity (Kutser et al. 2003). However, the high spatial resolution of sensors like Quickbird (2m pixel size), which is comparable to IKONOS (4m pixel size), has proven to be important for high mapping complexity independently of the spectral resolution (Capolsini et al. 2003). Remote sensing has been applied already with success in the tropical eastern Pacific region for medium-high resolution mapping of coral reefs in Panama (Guzman et al. 2004, Benfield et al. 2007).

Coral reefs are the most diverse marine ecosystems (Reaka-Kudla 1997) and they are being threatened by natural and human impacts (Coté & Reynolds 2006). Costa Rica has coral communities and reefs on the Caribbean coast and on the Pacific side (coast and off-shore islands) (Cortés & Jiménez 2003a, 2003b). Corals are in better condition on the protected...
off-shore islands, like “Isla del Caño”, where there is less impact from human activities (Cortés & Jiménez 2003b). However, everywhere corals are also affected by global climate change (Wilkinson 2008), with little chance of recovery (Richmond 1993, Nyström et al. 2000, Nyström & Folke 2001), with few exceptions in the tropical eastern Pacific, included “Isla del Caño” (Guzmán & Cortés 2001).

“Isla del Caño” is surrounded by fringing reefs (Guzman & Cortés 1989), sandy (Cortés et al. 1996) and rocky bottoms. It is protected by the category of Biological Reserve since 1976 for terrestrial habitats and since 1984 for marine habitats, and it is an important step in for the eastern Pacific marine corridor. Its coral reefs have been studied for more than 20 years (Guzman & Cortés 2001). Cortés et al. (1996) produced deep bottom profiles and sediment analysis on surrounding waters from 30 to 110m deep. Fonseca et al. (in prep.) completed the bathymetric model of “Isla del Caño” with information from shallow waters. Coral reefs of Costa Rica have not been mapped so far with high resolution, high accuracy and field validation.

It is very important for an adequate management of the marine environments of the island, that receives daily a large number of visitors for diving and beach recreation, to have an accurate marine habitats map. This map will help analyze the adequate size and zoning of the marine protected area. The objective of this study was to compare the performance of Hymap (16m) and Quickbird (2m) in creating a marine habitats map around “Isla del Caño”, and to give recommendations for the management of these marine environments, aiming to increase the size of the protected area.

MATERIAL AND METHODS

Site description: “Isla del Caño” is located 15km west from the Península de Osa, south Pacific of Costa Rica, eastern tropical Pacific (ETP), and it is protected as a Biological Reserve (Figs. 2 and 3). Mean water visibility is 20m. The island has five coral reef flats, mainly built by pocilloporid corals covered by crustose coralline algae, and isolated microatolls of Porites lobata. The reef slope and base is dominated by the massive coral Porites lobata, which is the predominant species of the island. The shallow sections of the reef are structured mainly by physical factors: wave action, temperature and salinity fluctuations, and low tide exposure. While the deeper sections are controlled by biological interactions: bioerosion, damselfish algae lawns, and corallivores (Guzman 1986, 1988, Guzman & Cortés 1989, 2001, Fonseca 1999, Fonseca et al. 2006).

Image processing: A Hymap imagery from the island was obtained from the Costa Rican Airborne Research and Technology Application (CARTA) mission, in March 29, 2005, at 15:30 and at altitude of 7820m; a project of the National Aeronautics and Space Administration (NASA). This image had a medium spatial resolution (16m pixel size). It was georeferenced to ground control points from the island coastline and had atmospheric correction performed with the HyCorr software that converts the radiance values into apparent surface reflectance values (level compatible with ATREM3 processing). An unsupervised classification (30 classes) was performed on this image before going to the field as a guide for collecting the ground control points. The near infrared band (NIR) was used to mask out the land. Areas of cloud and shade were removed from the Hymap image by using manual digitized areas of interest in ENVI 4.1. software.

A Quickbird imagery with a resolution of 2m and a radius of 6km around the island was taken on February 24, 2007, at 16:28, after the fieldtrip. An atmospheric correction was performed to this image using the dark pixel subtraction (Lyzenga 1978, 1981; Armstrong 1993).

The images and the maps were processed using ENVI 4.1 and ArcGIS 9.1 and enhanced with a 2% linear stretching. The image processing procedure is summarized in Fig. 1.
Field work: The scientific expedition to “Isla del Caño” took place from January 25 to February 5, 2007, on the M/V Phoenix, as part of an initiative of The Nature Conservancy to improve the coastal and marine management of the Peninsula de Osa. Ground control points from the coastline and marine environments of the island were collected during manta tow surveys (Rogers et al. 1994, Guzman et al. 2004) using a GPS Garmin GPSMAP 76S with an accuracy of ±10m. Visual identification of marine habitats from the surface was surveyed at 994 points to a maximum depth of 20m using pre-identified mapping categories.

Image pre-processing: The level of geometric accuracy of the raw imagery was checked from the ground control points collected along the coastline. A Principal Component Analysis was performed to the 18 bands of the visible spectral range of Hymap in order to compute a component that includes the contributing effects of all these bands, to reduce redundancy in the datasets and to integrate radiometric variance associated with the multispectral bands (sensu Jensen 2004, Mishra et al. 2006). PCA bands 1, 2, 3 were selected since they accounted for 99% of the variance. The following band combinations were also processed for accuracy comparison, and the water column was corrected for these bands combination with the empirical method Depth Invariant Index (Lyzenga 1978, 1981): bands 15,7,3 equivalent to the ETM mean wavelength for each visible band (1,2,3); bands 13,5,1 equivalent to the ETM minimum wavelength for each visible band (1,2,3); bands 13,9,1 equivalent to the CASI bands 2,4,5; bands 13,6,1 equivalent to the CASI bands 2,3,5; and bands 13,7,1 equivalent to the default bands selected for the True Color Composition of Hymap.
The water column was corrected in the Quickbird image using the “Depth Invariant Index” (Lyzenga 1978, 1981). The band pairs used for the Depth Invariant Index and final classification of this image were 3/1, 3/2, 2/1.

**Marine Habitats Classification:** Half of the ground control points were chosen randomly for image classification training data and the other half for accuracy assessment. From both images, supervised spectral signatures were generated for each habitat class using each point as a seed pixel for a “growing region of interest (ROI)”, so the number of pixels for Hymap grew to 62 and for Quickbird to 5870. ROI’s were assessed with photo-interpretation. The maximum likelihood was chosen for the supervised classification of both images and a 3x3 filter was applied to smooth the borders between categories in the final map.


**Accuracy Assessment:** Overall accuracy and Kappa coefficient were used to compare the classification results from both images. Kappa accounts for the amount of agreement that could be expected due to chance alone: poor = less than 0.20; fair = 0.20 to 0.40; moderate = 0.40 to 0.60; good = 0.60 to 0.80; and very good = 0.80 to 1.00 (Juurlink & Detsky 2005). Z tests were performed to test for significant differences between the Kappa coefficients. In order to determine the distribution of live coral cover, the following coral cover categories were used during the manta tow survey: 1. High: live coral > 40%; 2. Moderate: live coral 20-40%; 3. Low: live coral < 20%.

The overall classification accuracy for the different band combinations of Hymap is very similar (Table 1), so the different classifications were compared visually. The PCA bands 1, 2, 3, were selected for final classification of Hymap because they yielded a marine habitats map closer to what was found in the field.

**TABLE 1**

<table>
<thead>
<tr>
<th>Image</th>
<th>Band selection criteria</th>
<th>Band composition</th>
<th>Overall accuracy (%)</th>
<th>Kappa coefficient</th>
</tr>
</thead>
<tbody>
<tr>
<td>Quickbird</td>
<td>Visible bands</td>
<td>DII 3/1, 2/1, 3/2</td>
<td>87.33</td>
<td>0.79</td>
</tr>
<tr>
<td>Hymap</td>
<td>PCA bands (99%)</td>
<td>PCA 1, 2, 3</td>
<td>59.68</td>
<td>0.49</td>
</tr>
<tr>
<td></td>
<td>All processed selected band pairs</td>
<td>DII 12 pair bands</td>
<td>56.45</td>
<td>0.46</td>
</tr>
<tr>
<td></td>
<td>Bands 15,7,3 equivalent to the ETM mean wavelength for each visible band (1,2,3)</td>
<td>DII 15/7, 15/3, 7/3</td>
<td>53.22</td>
<td>0.42</td>
</tr>
<tr>
<td></td>
<td>Bands 13,5,1 equivalent to the ETM minimum wavelength for each visible band (1,2,3)</td>
<td>DII 13/5, 13/1, 5/1</td>
<td>58.06</td>
<td>0.48</td>
</tr>
<tr>
<td></td>
<td>Bands 13,9,1 equivalent to the CASI bands 2,4,5</td>
<td>DII 13/9, 13/1, 9/1</td>
<td>64.52</td>
<td>0.55</td>
</tr>
<tr>
<td></td>
<td>Bands 13,6,1 equivalent to the CASI bands 2,3,5</td>
<td>DII 13/6, 13/1, 6/1</td>
<td>66.13</td>
<td>0.57</td>
</tr>
<tr>
<td></td>
<td>Bands 13,7,1 equivalent to the default bands selected for the True Color Composition of Hymap</td>
<td>DII 13/7, 13/1, 7/1</td>
<td>64.52</td>
<td>0.56</td>
</tr>
</tbody>
</table>
RESULTS

The overall accuracy for classification (7 classes) was slightly higher using Quickbird (87%) than using Hymap (60%), but the difference was not significant (Z=4.33, df=61/5869, p=0.05); the Kappa coefficient for Quickbird is good (0.79) and for Hymap is moderate (0.49) (Table 1); deep water and sand category showed the best accuracies. Quickbird improved the user’s accuracy for 4 classes, and it was more trustable because the image was taken at a date closer to the time of the field trip, so the map produced from Quickbird was selected to show the marine habitats distribution from “Isla del Caño” (Fig. 2).

Coral reefs and coral communities in “Isla del Caño” account for 13% (325 pixels) and 14% (353 pixels) from shallow environments (2537 pixels) of the Quickbird image respectively (this proportion was calculated omitting the deep water category, > 20m). The resulting area was 1412m² (0.14ha) for coral communities and 1300m² (0.13ha) for coral reefs (Table 2).

During classification the sand with rodoliths was confused mainly with coral community, coral community was mainly misclassified as sand with boulders, all deep water pixels were classified correctly, bedrock was misinterpreted mainly as coral reef, sand with boulders was mostly mistaken as sand with rodoliths and coral reef as bedrock. The user and producer accuracies in the classification of coral reefs and coral communities are relatively high (Table 2).

Fig. 2. Marine habitats map of the marine protected area of “Reserva Biológica Isla del Caño”, and tourist diving sites. Results from the Quickbird image classification.
Live coral cover is higher in the northern and eastern coral reefs and communities, especially in the main coral flats of Bajo Glynn, Platanillo, Bajo Beltrán, Bajo Richmond and Bajo Cortés (Fig. 3).

DISCUSSION

Green et al. (2000) suggest that 60 to 80% is the recommended overall accuracy for coastal and marine resources inventory, so both sensors were considered good for medium to high resolution mapping and quantification of these habitats, as an input for coastal and marine management plans (Table 1). The higher spectral resolution of Hymap compensates to some extent for the lack of spatial resolution. Although the main limitation of Hymap is the loss of the habitats particular shape, currently Hymap is more cost effective than Quickbird in the case of Costa Rica, since 80% of the country was covered with Hymap in 2005, and images have a much lower user cost than Quickbird. This fact justifies the use of Hymap images to continue mapping shallow marine habitats from Costa Rica, with a medium to high resolution. If there will be possibility to collect airborne data in the future in Costa Rica we would recommend lower flights to obtain higher spatial resolutions of less than 1-2m, and planning the field work at the same time than the flight to improve the classification accuracy, although this would be more expensive. Habitat categories could have been combined in three broad classes to yield an even higher accuracy, but the 7 classes map was considered of more value for management purposes. Other studies using high resolution sensors like Quickbird or IKONOS, same classification method and an average of seven habitat classes report similar overall accuracies (Maeder et al. 2002, Mumby & Edwards 2002, Mishra et al. 2006, Benfield et al. 2007). However Benfield et al. (2007) improved the classification of Quickbird by 17% with the “Object Oriented Method”, which was not available for this study.

The user and producer accuracies in the Quickbird classification of coral reefs and coral communities, which are the habitats of more interest in this study, are relatively high (Table 2), however, as pointed by Mishra et al. (2006), there is still a significant amount of intermixing between marine habitats even at a spatial resolution of 2m. Coral reefs and communities in “Isla del Caño” were impacted by the 1982-83 and the

<table>
<thead>
<tr>
<th>Class/ Ground truthing (%)</th>
<th>Sand with rodolyths</th>
<th>Coral community</th>
<th>Deep water</th>
<th>Bedrock</th>
<th>Sand with boulders</th>
<th>Sand</th>
<th>Coral reef</th>
<th>Total nº pixels</th>
<th>Commission (%)</th>
<th>User Acc. (%)</th>
</tr>
</thead>
<tbody>
<tr>
<td>Unclassified</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>Sand with rodolyths</td>
<td>36.81</td>
<td>19.41</td>
<td>0</td>
<td>1.23</td>
<td>0</td>
<td>4.32</td>
<td>2.76</td>
<td>190</td>
<td>68.42</td>
<td>31.58</td>
</tr>
<tr>
<td>Coral community</td>
<td>5.52</td>
<td>53.85</td>
<td>0</td>
<td>27.16</td>
<td>56.52</td>
<td>1.37</td>
<td>21.87</td>
<td>353</td>
<td>58.36</td>
<td>41.64</td>
</tr>
<tr>
<td>Deep water</td>
<td>0</td>
<td>0</td>
<td>100</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>3333</td>
<td>0</td>
<td>100</td>
</tr>
<tr>
<td>Bedrock</td>
<td>0</td>
<td>3.3</td>
<td>0</td>
<td>24.69</td>
<td>0</td>
<td>0</td>
<td>34.18</td>
<td>190</td>
<td>89.47</td>
<td>10.53</td>
</tr>
<tr>
<td>Sand with boulders</td>
<td>12.88</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>0.42</td>
<td>23</td>
<td>100</td>
<td>0</td>
</tr>
<tr>
<td>Sand</td>
<td>44.79</td>
<td>0</td>
<td>0</td>
<td>0</td>
<td>9.78</td>
<td>94.3</td>
<td>0</td>
<td>1456</td>
<td>5.63</td>
<td>94.37</td>
</tr>
<tr>
<td>Coral reef</td>
<td>0</td>
<td>23.44</td>
<td>0</td>
<td>46.91</td>
<td>33.70</td>
<td>0</td>
<td>40.76</td>
<td>325</td>
<td>40.92</td>
<td>59.08</td>
</tr>
<tr>
<td>Total nº pixels</td>
<td>163</td>
<td>273</td>
<td>3333</td>
<td>81</td>
<td>92</td>
<td>1457</td>
<td>325</td>
<td>471</td>
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<td></td>
</tr>
<tr>
<td>Omission (%)</td>
<td>63.19</td>
<td>46.15</td>
<td>75.31</td>
<td>100</td>
<td>5.7</td>
<td>59.24</td>
<td></td>
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<tr>
<td>Prod. Acc. (%)</td>
<td>36.81</td>
<td>53.85</td>
<td>24.69</td>
<td>94.3</td>
<td>40.76</td>
<td></td>
<td></td>
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<td></td>
<td></td>
</tr>
</tbody>
</table>

a. Overall Accuracy (%)=(5126/5870)*100=87.33; Kappa coefficient =0.79.
1997-1998 El Niño, with losses of up to 50% of the live coral coverage (Guzman et al. 1987, Guzman & Cortés 1989, 2001), and by phytoplankton blooms in 1985, maybe associated with La Niña, with loss of some coral species from shallow reef zones (Guzman et al. 1990). Currently they are recovering (Guzman & Cortés 2001, Guzman et al. in prep.).

MANAGEMENT RECOMMENDATIONS

“Isla del Caño” holds a great diversity of marine habitats and should be considered an important area within the Eastern Tropical Marine Corridor. Close to the island (4km to the northeast) there is a carbonated bank that is being used as a diving site called “Paraiso”, but it is not within the Biological Reserve border which is 3km offshore. We recommend that this carbonate bank be considered as an important feature that should be represented by expanding the marine protected area to 4km offshore. There is some illegal commercial fishing that should be regulated to improve current protection. Many fish and shrimp fishing vessels are anchoring inside the Biological Reserve where
they pollute the water with solid and liquid wastes. Patrolling is done mostly on the north side of the island and in day time. We strongly recommend reinforcing the vigilance all around the island and all day and night long. For this purpose the reserve needs more staff and navigation equipment. The protected area zoning, carrying capacity and behavior of vessels and visitors and other regulations according to the management plan should be respected in order to secure the long term protection of the island. Mainly, the number of diving sites should not be increased, the number of tourist vessels per buoy and visitors per day should be reduced, and fishing and fishing vessels should not be allowed at all within the reserve.

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