



Journal of the Mexican Chemical Society

ISSN: 1870-249X

editor.jmcs@gmail.com

Sociedad Química de México

México

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Journal of the Mexican Chemical Society, vol. 54, núm. 1, 2010, pp. 59-68

Sociedad Química de México

Distrito Federal, México

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Sorption and Desorption of Remazol Yellow by a Fe-Zeolitic Tuff

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Received December 8, 2009; Accepted March 29, 2010

Abstract. The adsorption of remazol yellow from aqueous solution was evaluated using a Fe-zeolitic tuff. The adsorbent was characterized by scanning electron microscopy, IR spectroscopy and X-Ray diffraction. Sorption kinetic and isotherms were determined and the adsorption behavior was analyzed. Kinetic pseudo-second order and Langmuir-Freundlich models were successfully applied to the experimental results, indicating chemisorption on a heterogeneous material. The regeneration of the material was best accomplished by using a H₂O₂ solution. The sorption capacity of the Fe-zeolitic tuff increased when the saturated samples were treated with a H₂O₂ or FeCl₃ solution.

Keywords: Remazol, sorption, desorption, Fe-zeolitic tuff.

Resumen. Se estudió la adsorción de amarillo remazol utilizando una roca zeolítica acondicionada con hierro. El adsorbente se caracterizó por microscopía electrónica de barrido, espectroscopia IR y difracción de rayos-X. Se determinaron la cinética y las isotermas de sorción y el comportamiento fue analizado. Los resultados experimentales se ajustaron a los modelos de cinética de pseudo-segundo orden y Langmuir-Freundlich, indicando sorción química sobre un material heterogéneo. La regeneración del material se logró mejor con una solución de H₂O₂. La capacidad de sorción del material zeolítico aumentó cuando las muestras se trataron con soluciones de H₂O₂ o FeCl₃.

Palabras clave: Remazol, sorción, desorción, roca zeolítica.

Introduction

Color is one of the greatest contaminants in wastewater because it is highly visible and undesirable, causing considerable environmental pollution because of its absorption and reflection of sunlight entering the water; this interferes with the growth of bacteria and impedes the photosynthesis in aquatic plants [1] and also raises the chemical oxygen demand (COD) of the effluents [2]. It has been reported that 2% of dyes produced annually are discharged in effluent from manufacturing operations while it has been estimated that about 9% of the total amount (450 000 tons) of dyestuff produced in the world is discharged with textile wastewater [3].

In general, dyestuffs have complex chemical structures which are resistant to biological degradation. There are various methods available for the removal of dyes including membrane-separation, electrochemical, flocculation-coagulation, reverse osmosis, ozone oxidation, biological treatments, sorption, etc. [1]. Adsorption processes, on the other hand can be alternative methods for the removal of low concentrations of dyes, being economically feasible when compared with other methods.

Natural zeolites, usually regarded as low-cost materials are abundant and have been used as adsorbents for water purification [4]. The regeneration of adsorbents is an important issue in sorption processes. Fenton-driven oxidation, for example, has been proposed for regenerating spent organic-loaded carbons [5, 6]. The regeneration of zeolites is also an important issue in zeolite utilization. Currently, one of the main regeneration techniques is ion exchange for the restoration of the exchange capacity.

Both thermal and chemical methods have been employed to regenerate adsorbents saturated with organic compounds [7-9]. During thermal regeneration and wet oxidation organic contaminants are destroyed concomitantly with the regeneration of the adsorbents. In the former method the organics decompose thermally; in wet oxidation, oxidation in aqueous solution and in air via a free radical mechanism occurs [10]. Fenton's reaction has been studied, showing high efficiency in the destruction of most classes of organics in aqueous solutions [11]. This system is based on hydrogen peroxide, a clean oxidizer and a ferrous salt that generates hydroxyl radicals HO[•], which are very active for oxidizing organic molecules in an aqueous medium [12, 13]. Fenton's reagent was used to destroy adsorbed organochloro and remazol yellow contaminants on carbonaceous materials and a surfactant modified zeolite [11, 14]. Remazol yellow was removed using Fenton's reagent from a surfactant modified zeolite, but the surfactant was removed together with the dye and the adsorbent could not be regenerated, however, it was observed that the regenerated clinoptilolite-rich tuff had a higher sorption capacity than the original surfactant modified zeolite [14]. Some studies have reported the degradation of phenol in aqueous solutions using Fe-exchanged zeolite. Orange II was removed by using a bentonite clay-based Fe nanocomposite film as Fenton-type catalyst [15, 16].

Indigo carmine removal from aqueous solution was evaluated using Fe-zeolitic tuff. Kinetic pseudo-second order and Langmuir-Freundlich models were successfully applied to the experimental results; this indicated that there were chemisorptions on heterogeneous materials [17].

Therefore, the aim of this work was twofold: first of all, to determine the efficiency of a Fe-zeolitic tuff for the removal of

remazol yellow dye and secondly, to explore the possibility of regenerating the material using different processes.

Results and Discussion

Scanning electron microscopy

Figures 1-4 shows the morphology of the natural zeolitic tuff, sodium and iron modified zeolitic tuffs and the iron zeolitic tuff saturated with remazol yellow, respectively. The typical morphology coffin and cubic-like crystals [18] were best observed in the modified materials. Trgo *et al.* [19] have shown that the amphoteric nature of hydroxyl surface groups $[(\text{Al/Si})\text{-OH}]$ can lead to the formation of sites with such

different energies that this mechanical effect increases the number of possible adsorption locations. The chemical compositions of the zeolitic tuff, the sodium and Fe-zeolitic tuff and the saturated Fe-zeolitic tuff with remazol yellow are presented in Table 1. The main elements, Si, Al and O, corresponding to aluminosilicates, were identified. Iron was found in the natural zeolitic tuff being identified by X-Ray diffraction as Fe_2O_3 . The Na content diminished when the material was treated with FeCl_3 solution; this behavior indicated that sodium was replaced in the material by iron, since the content of this element was higher in the Fe-zeolitic tuff than in the sodium-zeolitic tuff and as a result, it might have induced oxidation processes [20]. The external, internal and effective cation exchange capacities have been reported in a previous research [21].

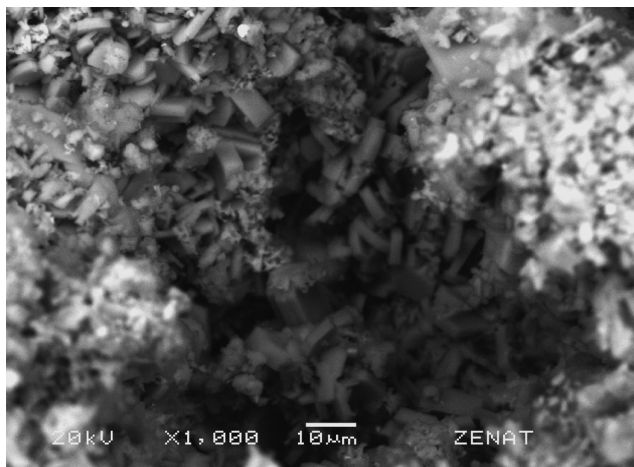


Fig. 1. Scanning electron micrograph of the zeolitic tuff.

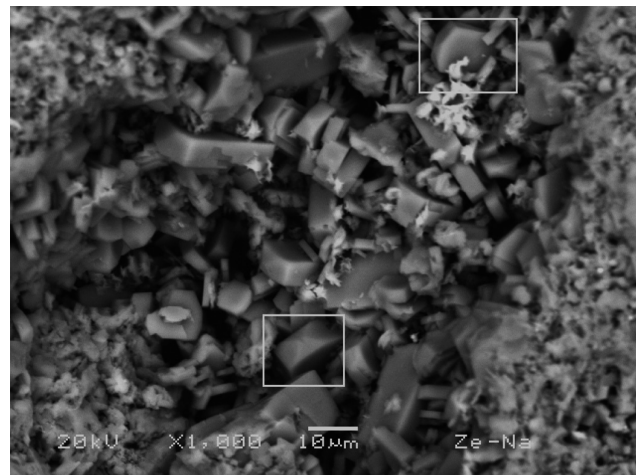


Fig. 2. Scanning electron micrograph of the Na-zeolitic tuff.

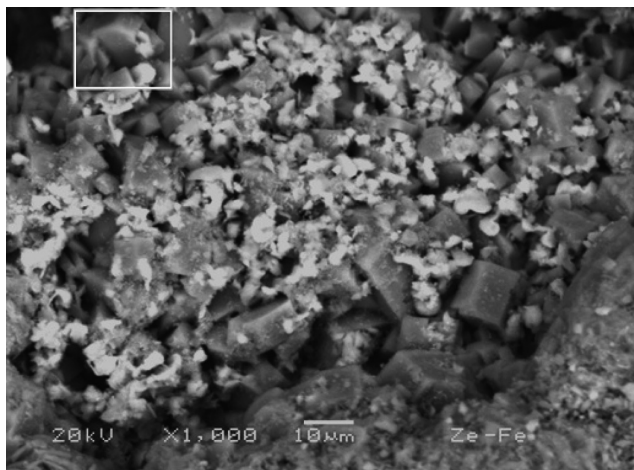


Fig. 3. Scanning electron micrograph of the Fe-zeolitic tuff.

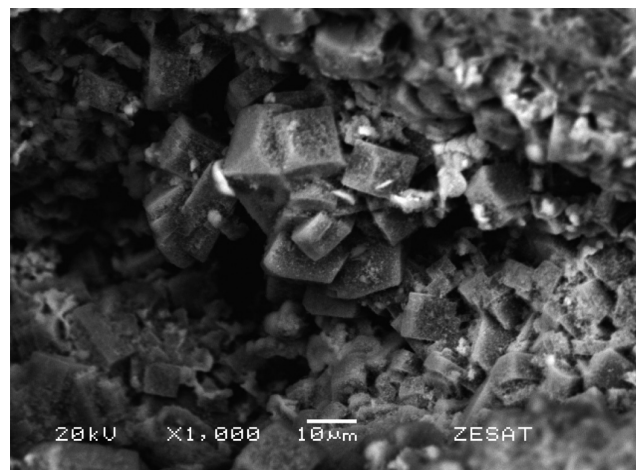


Fig. 4. Scanning electron micrograph of the yellow remazol saturated Fe-zeolitic tuff.

Table 1. Elemental analysis of the zeolitic tuff, Na-zeolitic tuff, Fe-zeolitic tuff and Fe-zeolitic tuff saturated with remazol yellow.

Element	Zeolitic tuff %	Na-zeolitic tuff %	Fe-zeolitic tuff %	Fe-zeolitic tuff saturated with remazol yellow %
C	12.83 ± 5.8	10.78 ± 1.92	8.79 ± 1.01	13.22 ± 5.87
O	44.01 ± 0.82	43.29 ± 1.80	42.62 ± 2.0	40.77 ± 3.68
Na	0.55 ± 0.12	1.65 ± 0.06	0.25 ± 0.06	0.35 ± 0.07
Mg	0.61 ± 0.06	0.46 ± 0.02	0.39 ± 0.05	0.35 ± 0.06
Al	5.98 ± 1.01	6.41 ± 0.14	4.90 ± 0.30	4.00 ± 1.51
Si	29.70 ± 4.90	33.12 ± 1.44	29.80 ± 1.50	27.47 ± 3.61
K	1.38 ± 0.16	1.51 ± 0.31	1.15 ± 0.31	1.19 ± 0.66
Ca	2.15 ± 0.52	1.48 ± 0.23	0.84 ± 0.05	0.76 ± 0.08
Fe	2.71 ± 0.85	1.31 ± 0.19	10.65 ± 1.95	11.42 ± 5.76
Cl	-	-	0.62 ± 0.11	0.37 ± 0.36

X-ray diffraction (XRD)

The X-ray powder diffraction patterns of the natural zeolitic tuff, Na-zeolitic tuff, Fe-zeolitic tuff and the Fe-zeolitic tuff saturated with remazol yellow were similar. The diffractograms were compared with clinoptilolite (JCPDS 039-1383) and quartz (JCPDS 33-1161), which were the principal components found in the zeolitic samples; other components found were sodium anortite (JCPDS 20-0528) and Fe_2O_3 (JCPDS 39-1346). No clay minerals were observed. These results suggest there were no notable changes in the structure of clinoptilolite after it was treated with sodium chloride and iron chloride solutions.

Infrared spectra of zeolitic materials

The natural sodium, iron and the Fe-zeolitic material saturated with remazol yellow showed similar IR spectra, the absorption peaks observed being assigned mainly to asymmetric and symmetric stretching, which are characteristics of this kind of material [22]. The characteristic infrared adsorption, which might be attributed to the pore opening aspect of the structure, was not observed since these frequencies generally occur below 400 cm^{-1} , beyond the region studied [22]. No absorption peaks due to remazol yellow were observed in the material saturated with this dye, probably because of its low concentration in the samples.

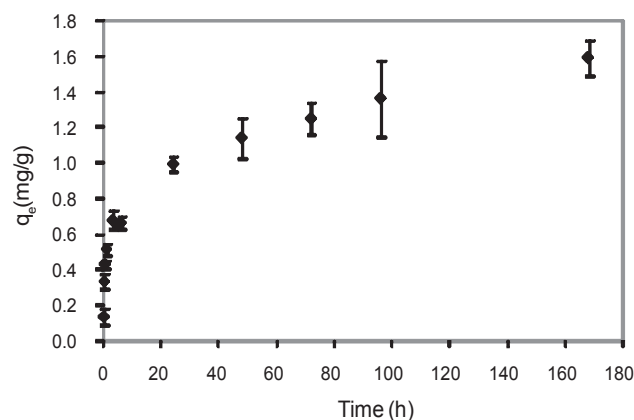
Surface Areas, BET (S_{BET})

The BET specific surface areas were 6.6, 5.2, 34.5, and $36.5\text{ m}^2/\text{g}$ for the natural zeolitic tuff, the Na-zeolitic tuff, Fe-zeolitic tuff and the dye saturated Fe-zeolitic tuff respectively.

As was observed, the specific surface area increased when the material was treated with FeCl_3 solutions. A similar behavior was reported for a zeolitic material treated with a nitrate solution [20]. This behavior could be attributed to pore openings during modifications of the zeolitic tuff.

Sorption kinetics

Sorption kinetics, expressed in terms of the rate of solute uptake (which governs the residence time), is one of the important considerations for economical wastewater treatment applications [23]. Figure 5 shows the relationship between contact time and the sorption capacities of the sorbent. According to figure 5, equilibrium was not completely reached even after 168 hours of contact time. One of the problems encountered was that the pH decreased in the solutions as contact time increased: the initial pH was 6.5 and after 168 hours it was 3.3;

**Fig. 5.** Sorption kinetics of remazol yellow by Fe-zeolitic tuff.

this is a disadvantage when adsorption material is employed in waste water treatments.

Several kinetic models were applied to the experimental data of the sorption of remazol yellow by Fe-zeolitic tuff. The feature constants of sorption were obtained by using a pseudo first order model, Elovich model and pseudo second order model. The sorption behavior of the dye was analyzed using linear regression.

Pseudo first order model (Lagergren)

This model is commonly used for homogeneous sorbents and physical sorption; the sorption rate is proportional to the solute concentration. If the sorption behavior is of the first order, then the experimental results could be adjusted to the following equation:

$$q_t = q_e (1 - e^{-Kt})$$

with the linear form being:

$$\ln (1 - (q_t/q_e)) = -Kt$$

where q_t and q_e are the amounts of adsorbed dye (mg/g) in the equilibrium and at time t (h), respectively and K (h^{-1}) is the sorption constant of Lagergren [24]. K , q_e ; their respective correlations are shown in Table 2.

Second order model (Elovich)

This model has been used suitably in chemisorptions on highly heterogeneous materials [25] and is represented by the following equation:

$$q_t = \beta \ln (\alpha \beta) + \beta \ln (t)$$

where q_t is the amount of adsorbed dye at time t , α is the sorption constant of the dye (mg/g) and β is the desorption constant (mg/g). Thus, the constant can be obtained from the slope

and intercept of the linear plot of q_t vs. $\ln(t)$. Table 2 shows the kinetic constants obtained with this model.

Pseudo-second order model

The pseudo second order model, proposed by Ho and McKay [26], is based on the assumption that the rate-limiting step may be chemisorption involving valence forces through the sharing or exchange of electrons between adsorbent and adsorbate. This model can be represented in the following form:

$$t/q_t = (1/kq_e^2) + (1/q_e)t$$

where q_t and q_e are the amount adsorbed at time t and at equilibrium (mg/g), respectively, and k is the pseudo second order rate constant for the sorption process (g/mg h). Thus, a plot of t/q_t vs. t should give a linear relationship with a slope of $1/q_e$ and an intercept $1/kq_e^2$. It was found that the adsorption system was best described by the pseudo-second order model. Figure 6 shows the adjustment of the experimental data to this model, the calculated k and q_e parameters being included in Table 2.

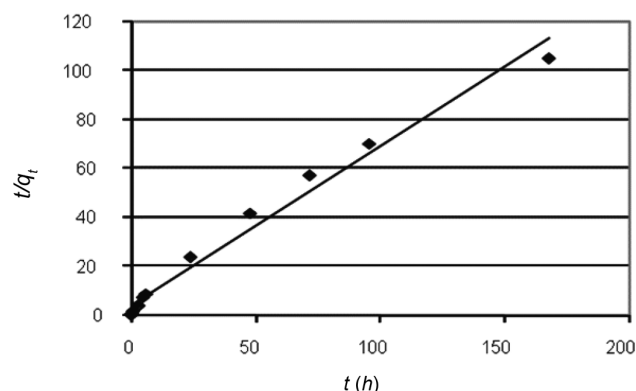


Fig. 6. Kinetic pseudo-second order model applied to the remazol yellow dye sorption from aqueous solution by Fe-zeolitic tuff.

Table 2. Kinetic parameters of remazol yellow by Fe-zeolitic tuff.

Kinetic model	Equation	Parameters	r^2
First order	$q_t = 1.4(1 - e^{-0.03t})$	$q_e = 1.4 \text{ mg/g}$ $K = 0.3 \text{ h}^{-1}$	0.96
Elovich	$q_t = 0.17 \ln(17.71) + 0.17 \ln(t)$	$\alpha = 101.21 \text{ mg/g}$ $\beta = 0.17 \text{ mg/g}$	0.95
Pseudo second order	$\frac{t}{q_t} = \frac{1}{(0.09)(1.54)^2} + \frac{1}{1.54}t$	$k = 0.09 \text{ g/mg h}$ $q_e = 1.54 \text{ mg/g}$	0.98

Sorption isotherms

The maximum sorption capacity of sorbents was obtained from the sorption isotherms to optimize the use of these materials. The experimental results were analyzed by non-linear Langmuir, Freundlich and Langmuir-Freundlich sorption models. The experimental data were analyzed with the help of STATISTICA® version 6.0 software.

Langmuir model

This model considers that maximum sorption corresponds to a monolayer saturated with dye molecules on the sorbent surface [27]. The Langmuir isotherm is represented by the following equation:

$$q_e = (q_0 b C_e) / (1 + b C_e)$$

where q_0 is the amount of dye adsorbed per unit weight of the adsorbent in forming a complete monolayer on the surface (mg/g), q_e is amount of dye adsorbed (mg/g), C_e is the concentration of the dye in the solution at equilibrium (mg/L) and b is the constant related to the energy or net enthalpy of sorption. Table 3 shows the parameters obtained by applying this model to the experimental results; although the correlation was not perfect, it was still possible to calculate the sorption capacity of the zeolitic material for remazol yellow.

Freundlich model

The Freundlich model, which has been applied to adsorbents with heterogeneous surfaces, considering multilayer sorption [14], is given by the following formula:

$$q = K_f C_e^{1/n}$$

where q is the amount of dye adsorbed per unit weight of adsorbent (mg/g), C_e is the equilibrium concentration of the dye in the solution (mg/L), K_f is the equilibrium constant indicative of adsorption capacity and n is the adsorption equilibrium constant whose reciprocal is indicative of the heterogeneity of surface sorbent. K_f , $1/n$ and q_e parameters are reported in Table 3. According to the characteristics of the sorbent, it is better to apply this model to the results than the Langmuir model because it is a heterogeneous material.

The values of $1/n$ are similar and less than unity, implying a heterogeneous surface structure with minimum interaction between adsorbed atoms [28].

Langmuir-Freundlich model

The Langmuir-Freundlich model is a combination of Langmuir and Freundlich models [14] and can be represented by the following equation:

Table 3. Sorption isotherm parameters of remazol yellow by Fe-zeolitic tuff.

Model	Equation	Parameters	r ²
Langmuir	$q_e = \frac{(3.49)(0.37)C_e}{1 + 0.37C_e}$	$q_0 = 3.49 \text{ mg/g}$ $b = 0.37$	0.85
Freundlich	$q = 1.08 C_e^{1/2.30}$	$K_f = 1.08$ $n = 2.30$	0.95
Langmuir-Freundlich	$q_e = \frac{(0.12)C_e^{(1/23.30)}}{1 + (-0.87)C_e^{(1/23.06)}}$	$K = 0.12$ $n = 23.30$ $a = -0.87$	0.98

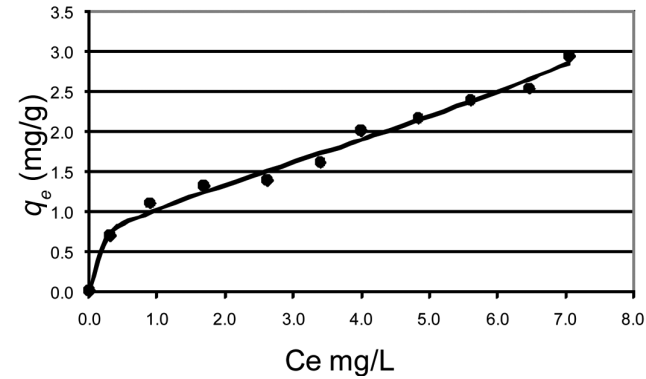


Fig. 7. Langmuir-Freundlich isotherm applied to the remazol yellow dye sorption from aqueous solution by Fe-zeolitic tuff.

$$q_e = (K C_e^{1/n}) / (1 + a C_e^{1/n})$$

where q_e is the amount of dye adsorbed per unit weight of adsorbent (mg/g), C_e is the equilibrium concentration of dye in solution, K and a are empirical constants.

The correlation coefficients were greater than 0.8 for the previous methods; the data correlated best with the Langmuir-Freundlich model, which had the highest correlation coefficients, as seen in Table 3 and Figure 7. Similar results were found by Dhaouadi and M'Henni [29] who used crude dehydrated sewage sludge.

Considering that the kinetic results obtained with the Fe-zeolitic tuff were best fitted to the pseudo-second order model and the isotherms to the Langmuir-Freundlich model, it can be suggested that the sorption mechanism of the dye is chemisorption on a heterogeneous material. The same behavior was reported by Torres-Pérez *et al.* [30] working with a modified zeolitic material with HDTMA (hexadecyltrimethylammonium) and remazol yellow.

Desorption of remazol yellow

The desorption results using different solutions are shown in Table 4; as observed, the dye in the modified zeolitic mate-

Table 4. Desorption of remazol yellow dye from the Fe-zeolitic tuff using different solutions.

Solution	Remazol yellow desorbed (%)	Sorption capacity of the dye saturated Fe-zeolitic tuff after being in contact with the solutions.
H ₂ O ₂	0.0	80.81
Fenton's reagent	54.96	92.35
FeCl ₃	45.82	121.21
Aqueous solution, pH = 1	0.00	-
Aqueous solution, pH = 3	1.05	-
Aqueous solution, pH = 6.7	1.29	-
Aqueous solution, pH = 9	4.27	-
Aqueous solution, pH = 11	63.42	54.33

rial is stable in pH values from 1 to 9; a similar behaviour was reported by Torres-Pérez *et al.* [30], who used a surfactant modified zeolite. The desorption of the dyes increases as the pH of the solutions increases and therefore the materials does not absorb any more dye. The saturated Fe-zeolitic tuff with remazol yellow, treated with an aqueous solution of pH 11 and then left in contact with the remazol yellow solution, showed a sorption capacity of 54.3 %.

H₂O₂ is one of the oxidation agents most used [31], its degradation producing water and oxygen. After being in contact dye saturated Fe-zeolitic tuff with the H₂O₂ solution, remazol yellow was not found in the solution; however, the sorption capacity of the resulting zeolitic material for the dye was 80.8 %. This behaviour shows that the dye is decomposed by the H₂O₂ solution, and it was therefore not detected in the remaining solutions.

When the saturated zeolitic material was treated with the Fenton's reagent as described above, the desorption percent was about 55 % and the sorption capacity of the resulting zeolitic material was 92 %; this behaviour is similar to the one observed by Torres-Pérez *et al.* [30] with a surfactant modified zeolite. FeCl₃ was the best reagent used for the desorption of remazol yellow since the adsorption capacity of the resulting zeolitic material increased 21 % with respect to the original Fe-zeolitic tuff.

Desorption kinetics

The desorption kinetics of yellow remazol from the dye saturated Fe-zeolitic tuff was determined with an H₂O₂ solution, an aqueous solution of pH 11 and Fenton's reagent solution.

When the dye saturated Fe-zeolitic tuff was treated with H₂O₂ solution, the remazol yellow content in the Fe-zeolitic

tuff did not change up to 72 hours of contact time. Remazol yellow was determined in the remaining aqueous solutions, this behaviour may be explained in two ways: the first one could indicate that remazol yellow is stable in the zeolitic tuff and is not desorbed; the second could indicate that the dye is decomposed by the H₂O₂ solution [32] and therefore was not observed in the remaining aqueous solutions. It is important to note that some bubbles were observed when the dye saturated Fe-zeolitic tuff was put into contact with the H₂O₂ solution; this behaviour may indicate the decomposition of the dye. However the dye did not decompose when it was treated only with the H₂O₂ solution, suggesting that iron plays an important role in the dye decomposition.

Figure 8 shows the desorption kinetics of yellow remazol from the dye saturated Fe-zeolitic tuff using a 0.02M FeCl₃ solution. A fast desorption occurs in the first minutes and then it slows down an almost complete desorption of the dye was observed in about 24 hours. Similar desorption behaviour was observed with an aqueous solution of pH 11, but in this last case only about 27 % of the dye was desorbed after 72 hours from the dye saturated Fe-zeolitic tuff.

Table 5 shows the sorption percent of remazol yellow dye with respect to the first sorption process by the Fe-zeolitic material treated with the H₂O₂ solution, the aqueous solution of pH 11, Fenton's reagent and the FeCl₃ solution for five sorption-desorption cycles. It was found that dye sorption was the highest when the Fe-zeolitic tuff was treated each time with the H₂O₂ solution: the sorption capacity increased up to 274.9 % in the fourth cycle, but opposite behaviour was reported with a surfactant modified zeolite [14]. The sorption was the lowest when the samples were treated with an aqueous solution of pH 11; the sorption capacity decreased with the cycle number, and in the fifth cycle the sorption capacity was only 10 %.

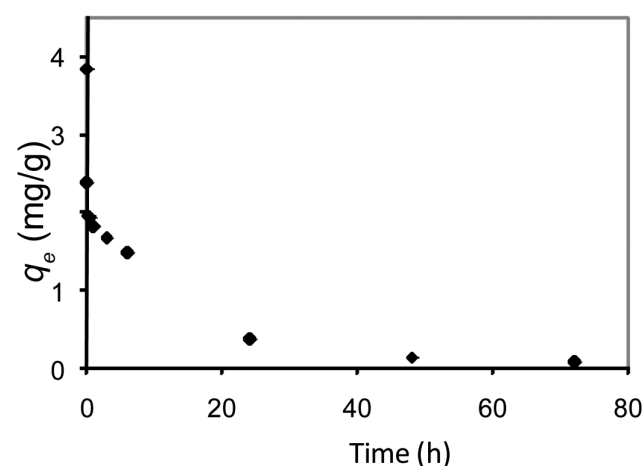
**Fig. 8.** Desorption kinetics of remazol yellow from dye saturated Fe-zeolitic tuff using 0.02M FeCl₃ solution.

Table 5. Sorption percent of remazol yellow by modified zeolitic samples in relation to the first sorption process during five sorption-desorption cycles.

Solution	Sorption cycles (sorption percent)				
	1	2	3	4	5
H ₂ O ₂	100.0 + 0.0	145.4 + 1.1	189.7 + 34.9	274.9 + 3.9	116.7 + 31.8
FeCl ₃	100.0 + 0.0	116.7 + 16.1	123.3 + 6.8	113.0 + 1.1	102.1 + 2.6
Fenton	100.0 + 0.0	81.8 + 2.5	73.7 + 0.6	69.8 + 3.7	83.1 + 8.0
Aqueous solution pH=11	100.0 + 0.0	50.8 + 4.5	20.5 + 0.6	10.0 + 1.4	10.3 + 0.2

These results show that Fe (III) and H₂O₂ are necessary for the degradation of remazol yellow. According to the results and those reported elsewhere [14], the regeneration of Fe-zeolitic tuff used in the sorption of remazol yellow is easier than the surfactant modified zeolitic tuff; because the former can be regenerated with an H₂O₂ solution. Gutiérrez *et al.* [17] reported that the chemical species of the dye has an important effect on the sorption of indigo carmine on the Fe-zeolitic tuff.

A possible interaction mechanism of the dye with the zeolitic material is catalytic degradation; recently, Kasiri *et al.* [33] used Fe-ZSM5 zeolite as a heterogeneous photo-Fenton catalyst for the degradation of acid blue 74.

Conclusions

Fe-zeolitic tuff is a suitable sorbent for the removal of dyes such as remazol yellow from aqueous solutions. A pseudo-second order model and the Langmuir-Freundlich model were successfully applied to the experimental data obtained with the Fe-zeolitic tuff, which may indicate that the sorption mechanism of the dye on this material is one of chemisorption on a heterogeneous material.

The regeneration of the material was best accomplished by using a H₂O₂ solution since the sorption capacity of the Fe-zeolitic tuff for remazol yellow increased when the saturated samples were treated with a H₂O₂ or FeCl₃ solution.

Experimental

Materials and Methods

Fe-zeolitic tuff

The clinoptilolite-rich tuff from Parral, Chihuahua, Mexico was milled and sieved. The grain size used in this work was between 1.19-1.68 mm. The zeolitic material was treated with

a solution of sodium chloride and refluxed (50 g of material with 250 mL of 0.125 M NaCl). This procedure was repeated 4 times. Afterwards, the zeolitic material was washed with distilled water until no presence of chloride ions was indicated in the washing solution using a AgNO₃ test. The sodium-treated zeolitic tuff was then dried at 333K for 2 h.

Fe-zeolitic tuff was prepared by mixing 40 g of the sodium zeolitic tuff with 500 mL of 0.1 M FeCl₃·6H₂O solution. This procedure was carried out twice in the same way as above. The Fe-zeolitic tuff was then dried at 333K for 2 h. The weight of the zeolitic material decreased after each conditioning; this behavior could be due to the removal of fine particles during the conditioning processes.

Characterization

Scanning electron microscopy

For scanning electron microscopy (SEM) observations, the Fe-zeolitic tuff samples before and after remazol yellow sorption were mounted directly on the holders and then observed at 10 and 20 kV in a JEOL JSM-5900-LD electron microscope. The microanalysis was done with an EDS (Energy X-ray Dispersive Spectroscopy) system.

Surface Areas, BET (S_{BET})

The BET specific surface areas were determined by standard multipoint techniques of nitrogen adsorption using a Micromeritics Gemini 2360 instrument. The samples were heated at 373K for 2 h before specific surface areas were measured.

X Ray Diffraction (XRD)

Powder diffractograms of the zeolitic samples were obtained with a Siemens D500 diffractometer coupled to a copper anode X-ray tube. The conventional diffractograms were used to

identify the compounds and to verify crystalline structure and were compared with clinoptilolite patron (JCPDS 039-1383) and quartz (JCPDS 33-1161).

IR spectroscopy

The IR spectra in the 4000-400 cm^{-1} range were recorded for the dye and the adsorbents treated with it at room temperature using a Nicolet Magna IRTM 550 FTIR. Samples were prepared following the standard KBr pellets method.

Sorption kinetics

Kinetic removal of remazol yellow dye by the Fe-zeolitic tuff was performed as follows: 100 mg of the adsorbent and 10 mL aliquots of a 40 mg/L solution were placed in centrifuge tubes and shaken for different time periods (5, 15 and 30 min, 1, 3, 5, 6, 24, 48, 72 and 168 h) at 120 rpm in a shaker at 30 °C. Later, the samples were centrifuged and decanted; the experiments were carried out in duplicate. The remazol yellow dye concentrations in the solutions were determined using a UV/Vis Perking Elmer Lambda 10 ultraviolet-visible spectrophotometer analyzer, with $\lambda = 485.47$ nm. The pH of each solution was measured before and after the treatments.

Sorption isotherms

One hundred milligrams of samples of the Fe-zeolitic tuff were put into contact with 10 mL of different concentrations of remazol yellow dye solutions (10, 20, 30, 40, 50, 60, 70, 80, 90 and 100 mg/L) for 168 h at 30 °C. The experiments were performed in duplicate. Later, the samples were centrifuged and decanted. Dye concentrations were determined in the liquid phases as described above and the pH was measured in each solution.

Desorption of the remazol yellow from the saturated Fe-zeolitic material.

Fe- zeolitic material was saturated with remazol yellow as follows: The Fe-zeolitic material was left for 168 h in contact with a 100 mg/L solution of remazol yellow, the phases were separated and the concentration of dye in the aqueous phase was measured in order to determine the quantity of dye retained by the Fe-zeolitic material. The saturated zeolitic material was washed with distilled water to eliminate the excess of dye solution; finally the zeolitic material was dried at room temperature.

Samples of the Fe-zeolitic material saturated with remazol yellow (100 mg) were left for 24 h in contact with 10 mL of the following aqueous solutions: 30% H_2O_2 , 0.02 M FeCl_3 , the

Fenton reactive (0.01 M FeCl_3 , 3% H_2O_2), acid solutions pH 1 and 3, basic solutions pH 9 and 11. The phases were separated and the concentrations of remazol yellow were measured in the aqueous solutions and desorption percentages were calculated. The zeolitic materials were washed with distilled water, dried and left for 48 h with 10 mL of a 100 mg/L of remazol yellow solution and the regeneration percentages were calculated considering the sorption quantities of dye in the saturation process. This last process was carried out 5 times to determine efficiency of the adsorbent in five sorption-desorption processes.

Desorption kinetics

Desorption kinetics was carried out with the solutions that showed the highest regeneration of the sorption material, which were a 30% H_2O_2 solution, the Fenton reactive (0.01 M FeCl_3 , 3% H_2O_2) and the solution of pH 11.

100 mg of the zeolitic material saturated with remazol yellow samples and 10 mL aliquots of each solution were placed in centrifuge tubes and shaken for different time periods (5, 15 and 30 minutes, 1, 3, 6, 24, 48 and 72 h) at 120 rpm in a shaker at 30 °C. The samples were then centrifuged and decanted; the experiments were performed in duplicate. The remazol yellow dye concentrations in the solutions were determined using a UV/Vis Perking Elmer Lambda 10 ultraviolet-visible spectrophotometer analyzer, with $\lambda = 485.47$ nm. The pH of each solution was measured before and after the treatments.

Acknowledgements

We acknowledge financial support from CONACYT, project 46219-Q.

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