



Atmósfera

ISSN: 0187-6236

editor@atmosfera.unam.mx

Universidad Nacional Autónoma de México
México

Dey, Sharadia; Gupta, Srimanta; Chakraborty, Arun; Sibanda, Precious
Influences of boundary layer phenomena and meteorology
on ambient air quality status of an urban area in eastern India
Atmósfera, vol. 31, no. 1, 2018, January-March, pp. 68-86
Universidad Nacional Autónoma de México
México

DOI: <https://doi.org/10.20937/ATM.2018.31.01.05>

Available in: <https://www.redalyc.org/articulo.oa?id=56555623009>

- How to cite
- Complete issue
- More information about this article
- Journal's webpage in redalyc.org

UNAM
redalyc.org

Scientific Information System Redalyc
Network of Scientific Journals from Latin America and the Caribbean, Spain and
Portugal

Project academic non-profit, developed under the open access initiative

Influences of boundary layer phenomena and meteorology on ambient air quality status of an urban area in eastern India

Sharadia DEY^{1,2*}, Srimanta GUPTA³, Arun CHAKRABORTY⁴ and Precious SIBANDA¹

¹ School of Mathematics, Statistics and Computer Science, University of KwaZulu-Natal, Private Bag X01 Scottsville, Pietermaritzburg, South Africa

² DST-NRF Centre of Excellence in Mathematical and Statistical Sciences (CoE-MaSS), Private Bag 3, Wits, 2050 Johannesburg, South Africa

³ The Department of Environmental Science, The University of Burdwan, Burdwan - 713104, West Bengal, India

⁴ Center for Oceans, Rivers, Atmosphere and Land Sciences, Indian Institute of Technology, Kharagpur Kharagpur-721302, West Bengal, India

*Corresponding author; email: sharadiadey1985@gmail.com

Received: February 13, 2017; accepted: December 6, 2017

RESUMEN

De abril a mayo de 2014 se llevó a cabo el monitoreo y análisis de varios contaminantes atmosféricos como material particulado (PM_{10}), dióxido de azufre (SO_2), dióxido de nitrógeno (NO_2), ozono (O_3), monóxido de carbono (CO) y amoníaco (NH_3), así como de parámetros meteorológicos (temperatura, humedad, y velocidad y dirección del viento), en tres sitios diferenciados (un área con alta densidad de tráfico vehicular, un área residencial y un área industrial) en una zona urbana del este de la India. El promedio máximo de la concentración de O_3 superficial por hora se encontró en el sitio industrial ($70.085 \mu g/m^3$), seguido por el residencial ($39.067 \mu g/m^3$) y el sitio con densidad vehicular alta ($25.936 \mu g/m^3$). Este último muestra la máxima concentración promedio de 24 h de NO_2 ($86.112 \mu g/m^3$) y NH_3 ($109.673 \mu g/m^3$). La máxima concentración promedio diaria (24 h) de PM_{10} ($392.127 \mu g/m^3$) y SO_2 ($174.069 \mu g/m^3$), así como la concentración promedio por hora más alta de CO ($3.846 mg/m^3$) en el sitio industrial pueden atribuirse a las emisiones de las industrias circundantes. El cálculo de los índices de calidad del aire (AQI, por sus siglas en inglés) muestra que tanto el sitio 1 (un área de tráfico congestionado con $AQI = 71.68$) corresponden a una categoría moderada de contaminación, en tanto que el sitio 2 (un área residencial con $AQI = 22.69$) se considera limpio. Este estudio detallado revela que la variación en las concentraciones de diversos contaminantes es una manifestación del efecto combinado de fuentes y patrones de emisión, la altura de la capa límite planetaria y la meteorología.

ABSTRACT

Monitoring and analyses of various air pollutants like particulate matter (PM_{10}), sulphur dioxide (SO_2), nitrogen dioxide (NO_2), ozone (O_3), carbon monoxide (CO) and ammonia (NH_3), as well as meteorological parameters (temperature, humidity, wind speed and direction) have been performed for the period between April and May, 2014 at three different sites (an area with high vehicular density, a residential area and an industrial area) of an urban area in eastern India. The hourly average concentration of surface ozone is found maximum at the industrial site ($70.085 \mu g/m^3$) followed by the residential site ($39.067 \mu g/m^3$) and the site with high vehicular density ($25.936 \mu g/m^3$). The site with high vehicular density shows the maximum 24 h average concentration of NO_2 ($86.112 \mu g/m^3$) and NH_3 ($109.673 \mu g/m^3$). Maximum daily (24 h) average concentration of PM_{10} ($392.127 \mu g/m^3$) and SO_2 ($174.069 \mu g/m^3$), and highest hourly average of CO ($3.846 mg/m^3$) at the industrial site can be attributed to emissions from the surrounding industries. Computation of air quality indices (AQI) show that both site I (a traffic congested area with $AQI = 53.85$) and site III (an industrial area with $AQI = 71.68$) fall under a moderately polluted category whereas site II (a residential area with $AQI = 22.69$) is found to be clean.

This detailed study reveals that variations in the concentration of different air pollutants manifest the combined effect of emission sources and patterns, planetary boundary layer height and meteorology.

Keywords: Air pollutants, planetary boundary layer (PBL), meteorology, air quality index (AQI).

1. Introduction

The planetary boundary layer (PBL) is the lowest layer of the troposphere that is directly influenced by its contact with the earth's surface. The height of the PBL is a crucial parameter for air quality analysis, pollutants dispersion and quantification of pollutant emissions and sources (Coen et al., 2014). The role of convective boundary layer is more pronounced over tropical regions (like India) in comparison to other parts of the world. The meteorological processes within the PBL influence the transport and chemistry of air pollutants (Reddy et al., 2012). Significant work has been done around the world on the influence of boundary layer phenomena and meteorology on the characteristics of air pollution and the status of air quality (Kolev et al., 2011; Quan et al., 2013).

Surface emissions of major air pollutants have been changing substantially over the tropical region in the past few decades. The global epidemiological study on the effect of air pollution revealed that particulate matter and gaseous pollutants could cause severe health effects (particularly in urban areas) like respiratory and cardiovascular diseases, resulting in cardiopulmonary mortality (Dockery et al., 1993; Koken et al., 2003). Concern about air pollution in urban areas is receiving increasing importance world-wide, especially pollution by gaseous and particulate trace metals (Azad and Kitada, 1998; Salam et al., 2003; Begum et al., 2004).

Systematic air quality monitoring programs have also been performed in different developing countries like Bangladesh and Pakistan (Ali and Athar, 2008; Salam et al., 2008). In India, air pollution has increased rapidly with population growth, increasing number of motor vehicles, use of fuels with poor environmental performance, badly maintained transportation systems, poor land use pattern, and above all, ineffective environmental regulations (Ghose et al., 2004; Gupta et al., 2008). Stationary and mobile sources emit a wide variety of pollutants, principally carbon monoxide (CO), nitrogen oxides (NO_x), sulphur oxides (SO_x), volatile organic compounds (VOCs) and particulate matter, which have an increasing impact on urban air quality. In addition,

photochemical reactions resulting from the action of sunlight on nitrogen dioxide (NO₂) and VOCs emitted by vehicles lead to the formation of ozone, a secondary long-range pollutant, which impacts rural areas often far from the original emission site. Systematic analysis has shown that boundary layer characteristics and meteorological parameters strongly influence the variation of surface ozone (Reddy et al., 2012; Ghosh et al., 2013, 2015). Long-range horizontal transport also plays a crucial role in the variation of ground level ozone concentration (Ghosh and Sarkar, 2015). Many researchers have focused on the monitoring and analysis of air pollutants and air quality in different parts of India like Kolkata, Delhi, Lucknow, Haryana, Chennai, Mumbai, Dhanbad-Jharia, Raniganj-Asansol, Durgapur and Burdwan (Lal and Patil, 2001; Jain and Saxena, 2002; Goyal and Sidhartha, 2003; Reddy and Ruj, 2003; Verma et al., 2003; Kaushik et al., 2006; Pulikesi et al., 2006; Gupta et al., 2008; Chattopadhyay et al., 2010; Dey et al., 2014, 2017).

Durgapur is one of the most polluted cities of the country with a cluster of industries. It is located in the eastern part of the country. So far no systematic air quality-monitoring program with a Geographic Information System (GIS) approach has been reported from this region. Therefore, systematic as well as scientific investigation of the ambient air quality status of different areas (with high vehicular density, residential and industrial) is essential to formulate a proper mitigation strategy for pollution.

The present study focuses on the variations in the concentration of particulate matter (PM₁₀) and various gaseous pollutants (SO₂, NO₂, CO, NH₃ and O₃) over an urban region of eastern India and attempts to find the influences of meteorology and planetary boundary layer characteristics on dispersion and dilution of pollutants near ground level. It also aims to explore the probable sources of air pollutants and the air quality status of the three different sites under consideration. Various statistical techniques are used for understanding the relationship between concentration of air pollutants, meteorological variables and PBL characteristics.

Computation of the air quality index (AQI) for each of the three chosen sites is done.

2. Materials and method

2.1 Site description

Data were collected at three different sites (Fig. 1) of an urban area in eastern India during April and May, 2014. These sites are located in Durgapur (Burdwan district) of West Bengal. Site I (Faridpur) is situated at $23^{\circ} 32' 27.61''$ N and $87^{\circ} 17' 7.28''$ E. The sampling site is adjacent to National High Way (NH-2) or the Grand Trunk Road. This site represents an area with high vehicular density.

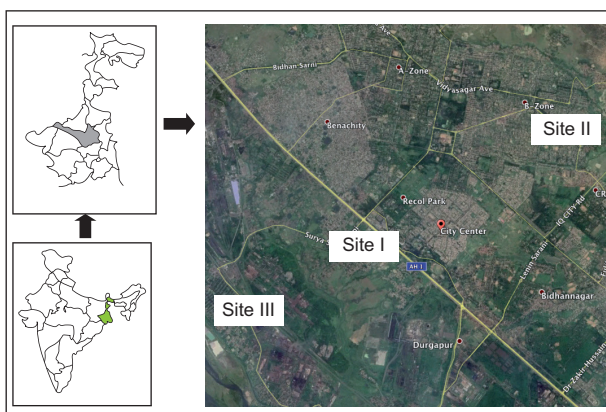


Fig. 1. Maps showing the location of the three sampling sites.

Site II (B Zone) is located at $23^{\circ} 33' 53.79''$ N and $87^{\circ} 19' 17.26''$ E. This site is situated in a residential area which is far away from national highway and industrial area.

Site III (DVC-DTPS Colony) is situated in an industrial area at $23^{\circ} 31' 19.42''$ N and $87^{\circ} 15' 33.59''$ E. River Damodar flows to its south. The Durgapur Steel Plant (DSP) and the Durgapur Thermal Power Station of the Damodar Valley Corporation (DVC) are close to this site.

2.2 Meteorological background

This area undergoes hot and dry summers, lasting from March to the middle of June, with average daily temperatures near 32°C , followed by the monsoon season with heavy precipitation and somewhat lower temperatures. Durgapur receives most of its annual rainfall of around 32 mm during this season. The

monsoon is followed by a chilled dry winter from November to January. Temperatures are quite moderate, with average daily temperatures near 20°C . There is a short autumn at the end of October and a short spring in February, both of which have relatively moderate temperatures of around 25°C . This area with undulating topography is covered with red and yellow Ultisol soils, with an average elevation of 65 masl.

2.3 Measurement techniques

Various criteria pollutants and meteorological parameters were monitored at these three different sites from April to May 2014. Sampling was carried out at each site for 10 consecutive days. Monitoring of PM_{10} , SO_2 , NO_2 , O_3 , CO , NH_3 , and meteorological parameters (temperature, humidity, wind speed and direction) was performed for 24 h. Data were collected on the roof of single-story buildings in all the three sites.

The concentration of PM_{10} was obtained with the aid of a high volume sampler (Envirotech APM 460BL) from Envirotech. Air was drawn through a size-selective inlet and through a 20.3×25.4 cm filter paper (GF/A) at a flow rate which is typically 1132 L/min. The high volume sampler was run in the open air for 24 h and initial and final flow rate values were recorded. The collection interval of PM_{10} , SO_2 and NO_2 was 24 h. After the required time of sampling, the filter paper was taken out and stored carefully in an envelope within the desiccator for subsequent analysis. Particles with aerodynamic diameters less than a cut-point of the inlet were collected by the filter. The mass of these particles was determined by the difference in filter weights prior to and after sampling. The concentration of suspended particulate matter in the designated size range was calculated by dividing the weight gain of the filter by the volume of air sampled.

SO_2 , NO_2 and NH_3 were monitored with the same kind of high volume sampler. Twenty-five millimeters of potassium tetrachloromercurate (TCM) absorbing solution were placed in an impinger and sampled for 24 h. A calibrated flow-measuring device was used to control the airflow from 0.2 to 1.0 L/min. The volume of the sample was then measured and stored in a sample storage bottle. The chemical analyses of samples and blanks were performed in the laboratory.

(West and Gaeke, 1956) for estimation of ambient SO_2 ($\mu\text{g m}^{-3}$) concentration. The absorbances of the solutions were measured at 560 nm.

Ambient air was continuously drawn into 25 ml of a sodium hydroxide solution for analysis and estimation of NO_2 concentration. The flow rate of ambient air was 0.2–1.0 L/min and was controlled by a calibrated flow measuring device. The samples and blanks were chemically analyzed and the concentration of the nitrite ion was determined by colorimetric analysis (Jacobs and Hochheiser, 1958). The absorbances of the solutions were measured at 540 nm.

The concentration of ambient ammonia was determined by the indophenol blue method in the laboratory. Ambient air was continuously drawn into 25 ml of a dilute solution of sulphuric acid with a flow rate of 1–2 L/min and sampled for one hour. The ammonium sulphate thus formed in the sample was allowed to react with phenol and alkaline sodium hypochlorite to produce indophenol. Sodium nitroprusside acts as catalyst in this reaction. The estimation of NH_3 is done by measuring the absorbance of the samples and the blanks at 630 nm (Lodge, 1998). The absorbance of all the samples and blanks for the estimation of SO_2 , NO_2 and NH_3 was measured by means of a UV-VIS spectrophotometer (Lambda 35, Perkin Elmer).

The concentration of O_3 was continuously monitored with the help of an air quality sensor (Aeroqual series 500). The sampling was carried out for 24 h with a sampling interval of 1 h.

The concentration of CO was measured with a lightweight portable indicator (PPSMPL gaZguardTx). The sampling of CO was performed for 24 h with a sampling interval of 1 h.

Meteorological parameters such as humidity, temperature, and wind speed and direction were also recorded during the sampling time. Humidity and temperature were measured with portable hygrometer (Model-HTC-1), wind speed was measured with a digital anemometer (Lutron-AM-4201) and wind direction was recorded with a wind vane. The portable hygrometer, digital anemometer and wind vane were all placed on the roof of single story buildings (approximately 3.5 m above the ground level).

To ensure the quality of the laboratory analysis of air pollutant samples, pure chemicals of AR grade (Merck, Germany) were used. Low actinic glassware was used during the analytical analysis

of the samples. The quality of the titrimetric method was maintained by considering the mean titre value obtained by averaging the three concordant titre values. Different instruments used during sampling (i.e., the Aeroqual high volume sampler and the gaZguardTx lightweight portable indicator) were properly maintained and calibrated in order to fulfill the quality control requirements.

Data of parameters like PBL height, vertical mixing coefficient and cloud cover for the specific days of observation over the sites of interest were collected from the archived data of NOAA's Air Resources Laboratory (<https://ready.arl.noaa.gov>). The data of sunrise and sunset over the three sites during the study period were collected from the site of NOAA's Earth System Research Lab (<http://www.esrl.noaa.gov/gmd/grad/solcalc/sunrise.html>). Wind rose diagrams were constructed with the monitored data (wind direction and speed) for the study period using the WRPLOT View software (v. 7.0.0).

2.4 Statistical analysis

The primary data set was processed and analyzed with different statistical tools to obtain their descriptive statistics.

2.4.1 Multivariate statistical analysis

Principal components analysis (PCA) is a multivariate statistical procedure designed to classify variables based on their correlations. The goal of PCA and other factor analysis procedures is to consolidate a large number of observed variables into a smaller number of factors (components) that can be more readily interpreted as the underlying processes. The XLSTAT 2012 statistical software was used to perform PCA, specifying the principal components method with varimax rotation (Kaiser, 1958). The rotation of the component axis is performed so that components are clearly defined by high loadings for some variables and low loadings for others, facilitating the interpretation in terms of original variables.

2.5 Air quality index (AQI)

An AQI could be defined as a scheme that transforms the (weighted) values of individual air pollution related parameters into single number. Initially, the air quality rating of each parameter used for monitoring is determined in each site by the formula

$$q = 100 * V/V_s \quad (1)$$

where q is the quality rating, V the observed value of the parameter and V_s the value recommended for that parameter.

The geometric mean of 'n' number of quality ratings was determined by

$$g = \text{antilog} \{(\log a + \log b + \dots \log x)/n\}; \quad (2)$$

where g is geometric mean; a, b, c, d , and x are different values of air quality ratings; n is the number of values of air quality ratings, \log is the logarithm function and antilog is its inverse.

The air quality status of different sites is determined on the basis of the AQI categorization proposed by Mudri (1990) (Table I).

Table I. Air quality index (AQI) proposed by Mudri (1990).

Category	AQI of ambient air	Description of ambient air quality
I	< 10	Very clean
II	10-25	Clean
III	25-50	Fairly clean
IV	50-75	Moderately polluted
V	75-100	Polluted
VI	100-125	Heavily polluted
VII	> 125	Severely polluted

2.6 Backward trajectory simulation

An air mass backward trajectory analysis was carried out to identify the sources and pathways of air pollutants. The simulation was done using NOAA's HYbrid Single-Particle Lagrangian Integrated Trajectory (HYSPLIT) model (Draxler and Hess, 1998) with GDAS meteorological data. The 10-day back trajectories at three altitude levels (100, 500 and 1500 m above the ground level) are simulated for all three sites.

3. Results and discussion

3.1 Variation of the gaseous pollutants and particulate matter concentrations over the three sites

PM₁₀, O₃, SO₂, NO₂, CO, and NH₃ are some of the major air pollutants which require monitoring and analysis to understand the air quality status of an area.

3.1.1 Influences of PBL characteristics on pollutant concentrations

Gaseous NH₃ has long been known to play a key role in atmospheric chemical processes. NH₃ is a major contributor to secondary aerosol formation in the atmosphere. Anthropogenic sources of ambient ammonia include industrial processes, vehicular emissions and volatilization from soils and oceans (Behera et al., 2013). CO and NO₂ are the primary air pollutants, which are directly emitted into the air. The major anthropogenic source of CO and NO₂ is the burning of fossil fuels (combustion in motor vehicles). The primary sources of SO₂ are burning of fossil fuels and petroleum products, as sulfur is present in these products as impurities.

Variations in concentration of air pollutants (NO₂, SO₂, NH₃ and CO and PM₁₀) and average PBL heights over sites I, II, and III during the study period are plotted in Fig. 2a, b, c, respectively.

The concentration of NO₂ is found to be high in sites I and III. Forty percent of the obtained values of NO₂ in site I exceeded the National Ambient Air Quality Standard (NAAQS) limit and 50% of the monitored values exceeded the NAAQS limit in site III. However, in site II the concentration of NO₂ is much lower than the permissible limit. The concentration of NH₃ is lower than the NAAQS permissible limit in all three sites. The 24-h average level of NH₃ is highest in site I (109.7 µg m⁻³ and a range of 17.1-231.8 µg m⁻³) and lowest in site II (with a 24-h average of 13.6 µg m⁻³ and a range of 4.2-29.2 µg m⁻³). The maximum SO₂ concentration was found at site III (industrial area) with all values exceeding the NAAQS permissible limit. A similar kind of result was also found by Chattopadhyay et al. (2010). In sites I and II, the SO₂ level is much lower than the permissible limit. The average hourly concentration of CO in the ambient air was lowest in site II (0.6 µg m⁻³) and highest in site III (3.8 µg m⁻³). Only 10% of the monitored data exceeded the NAAQS limits at site III whereas concentration values in sites I and II did not exceed the limits prescribed by NAAQS. High PM₁₀ levels were observed in all three sites. The highest average PM₁₀ level was observed at site III (392.1 µg m⁻³ with a range of 247.8-499 µg m⁻³) and all the monitored values exceeded the permissible limits. The lowest PM₁₀ level was found at site II with a mean of 85.7 µg m⁻³. Only 40% of the monitored

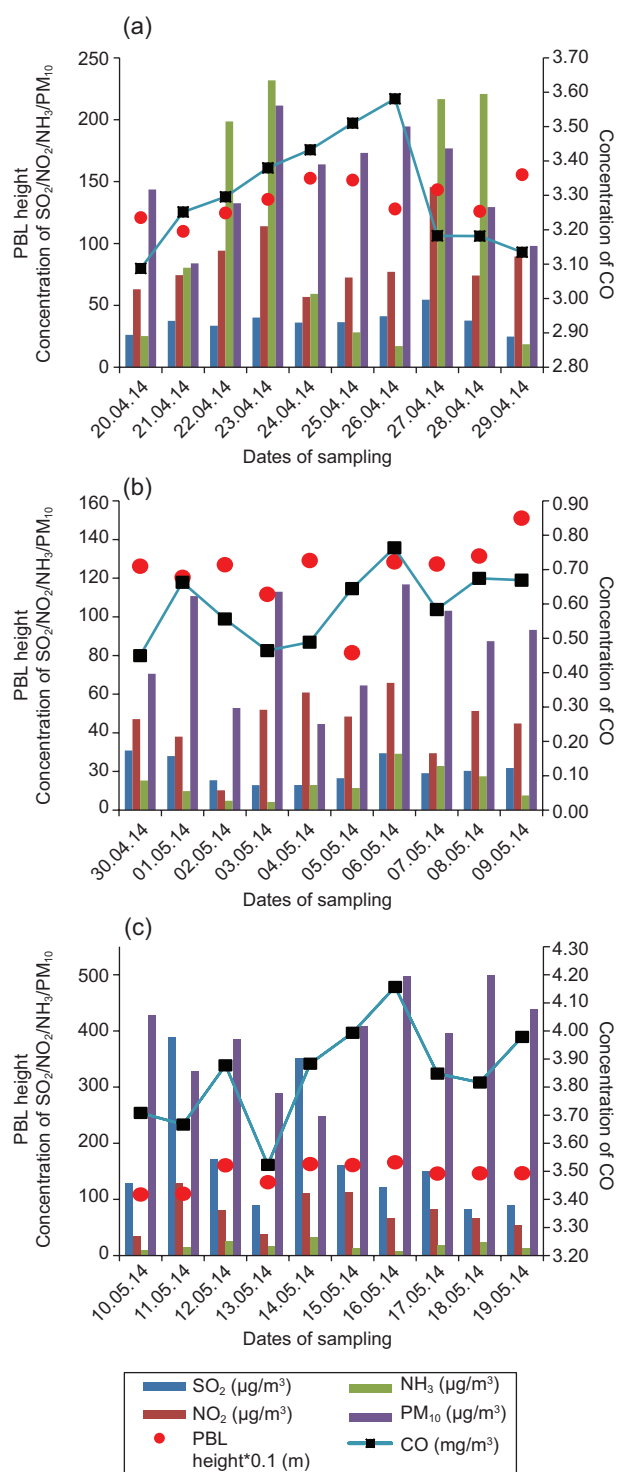


Fig. 2 Day-to-day variations of different air pollutants and PBL heights during the study period over (a) site I (Faridpur), (b) site II (B zone), and (c) site III (DVC-DTPS Colony).

values exceeded the NAAQS limits at this site. In site I, 80% of the monitored values exceeded the permissible limits. Although the estimation of ammonia emissions to the atmosphere is characterized by a high degree of uncertainty, the transportation sector is believed to be the largest contributor to urban emissions of this compound. As site I is adjacent to a national highway (NH-2), comparatively higher concentrations of ambient NH₃ are found over this site than in the other two sites. Ambient NH₃ contributes to increase PM concentrations in the urban atmosphere (Walker et al. 2004; Whitehead et al., 2007). The presence of high NH₃ levels at site I also contributes to a high PM₁₀ concentration at this site. High traffic density also enhances the concentration of CO and NO₂ over this site.

High concentrations of SO₂, NO₂, CO and PM₁₀ were recorded at site III (Fig. 2c). The burning of fossil fuels in various industrial processes produce huge amount of sulfur oxides, nitrogen and carbon. Therefore, industrial activities near site III account for the high concentration of SO₂, NO₂, CO and PM₁₀ in the ambient atmosphere of this site. Local biomass burning and domestic coal burning might be the cause of the occasional presence of very high values of these air pollutants.

Boundary layer characteristics also play a crucial role in the dispersion and dilution of air pollutants near the earth's surface. It is observed that days with higher average PBL height show lower pollutant concentrations near the ground (Fig. 2a, b, c). Higher average PBL heights and enhanced convective activities facilitate the dilution of pollutants, thereby decreasing their concentration near the earth's surface, whereas in days with lower PBL height pollutants get trapped near the ground.

3.1.2 Diurnal variation of ground level ozone concentrations over the three sites

The diurnal variation of ground level ozone at sites I (Faridpur), II (B zone), and III (DVC-DTPS residential area) are shown in Figs. 3-5, respectively.

In general, contour diagrams show that ozone concentration begins to increase gradually after sunrise, attains a maximum around noontime and starts decreasing after late afternoon or evening. Sunrise and sunset times during the study period over the three sites are reported in Table II.

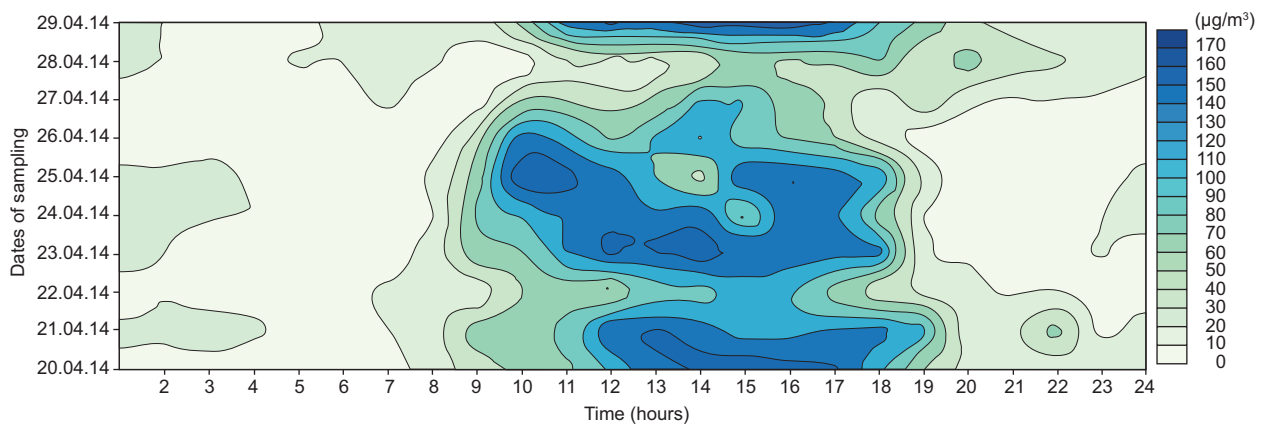


Fig. 3. Diurnal variation of ground level ozone concentration at site I (Faridpur) during the study period.

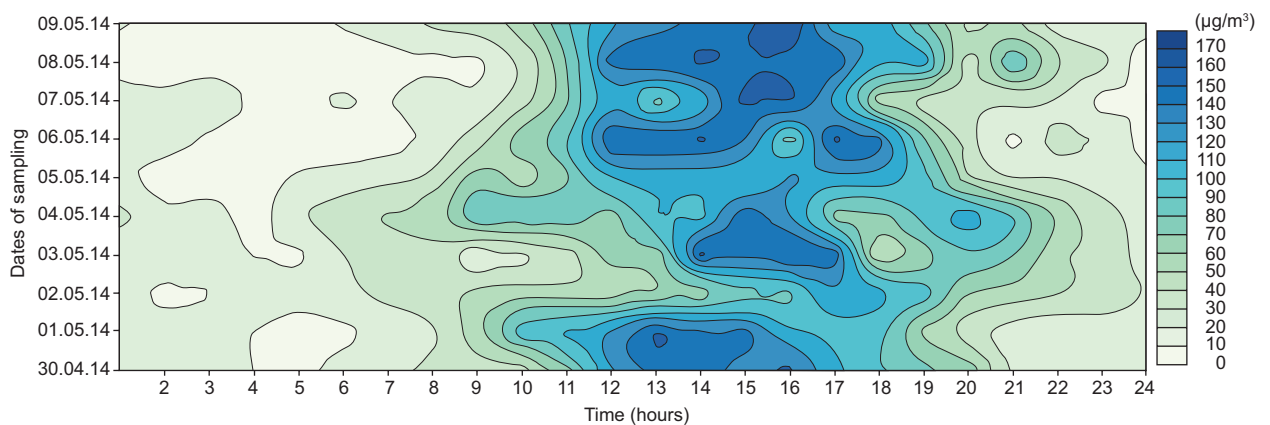


Fig. 4. Diurnal variation of ground level ozone concentration at site II (B zone) during the study period.

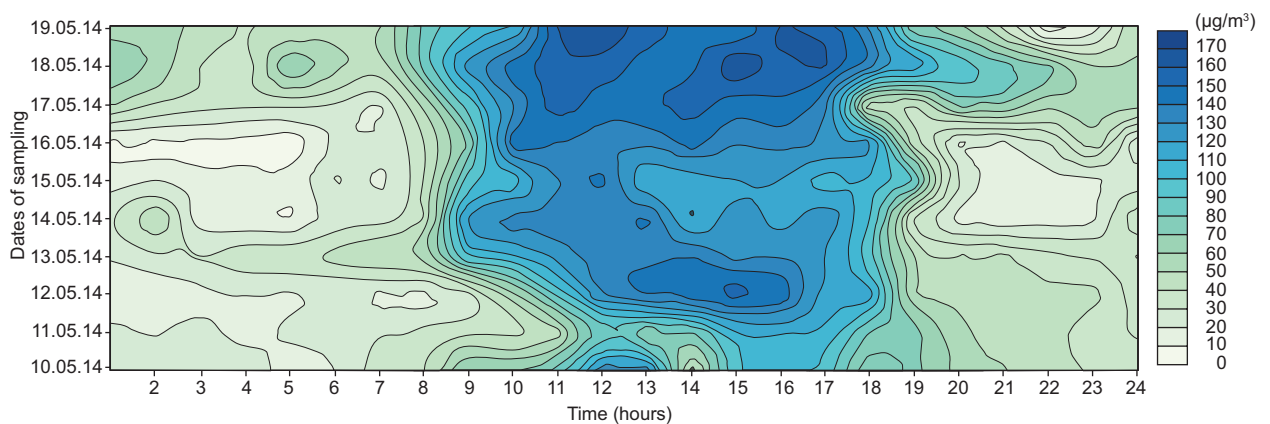


Fig. 5. Diurnal variation of ground level ozone concentration at site III (DVC-DTPS residential area) during the study period.

Table II. Sunrise and sunset times over the three sites during the study period.

Site I			Site II			Site III		
Date	Sunrise (am)	Sunset (pm)	Date	Sunrise (am)	Sunset (pm)	Date	Sunrise (am)	Sunset (pm)
20.04.14	5:16	6:04	30.04.14	5:08	6:08	10.05.14	5:02	6:13
21.04.14	5:15	6:04	01.05.14	5:07	6:09	11.05.14	5:01	6:14
22.04.14	5:14	6:05	02.05.14	5:07	6:09	12.05.14	5:01	6:14
23.04.14	5:14	6:05	03.05.14	5:06	6:10	13.05.14	5:00	6:15
24.04.14	5:13	6:06	04.05.14	5:05	6:10	14.05.14	5:00	6:15
25.04.14	5:12	6:06	05.05.14	5:05	6:11	15.05.14	4:59	6:16
26.04.14	5:11	6:07	06.05.14	5:04	6:11	16.05.14	4:59	6:16
27.04.14	5:10	6:07	07.05.14	5:03	6:12	17.05.14	4:58	6:17
28.04.14	5:10	6:08	08.05.14	5:03	6:12	18.05.14	4:58	6:17
29.04.14	5:09	6:08	09.05.14	5:02	6:12	19.05.14	4:58	6:17

The daytime increase in ozone concentrations, which is a pronounced feature of urban polluted areas, is basically due to the photo oxidation of precursor gases such as CO, CH₄, and non methane hydrocarbons (NMHCs) in the presence of a sufficient amount of NO_x (Lal et al., 2000).

Ozone concentration is highest around noon and in the afternoon (12:00-16:00 LT) in all three sites. Meteorological factors, boundary layer processes and regional/long range transport play vital roles in determining the ozone concentration depending on the topography of the observational sites. After sunrise, the boundary layer height gradually increases from 200-300 m to about 1500-2000 m during noon hours due to convective heating and the stratification of layers (Lal et al., 2000). During noon hours, air rich in ozone in higher altitudes mixes with the air in lower altitudes, which has a lesser amount of ozone (Naja and Lal, 1997). During nighttime, the photochemical production of ozone ceases and vertical transport of this element is inhibited by a nocturnal inversion layer. Loss of ozone by reaction with NO (titration) and dry deposition are the other causes of low ozone concentration during nighttime. Low ozone concentrations during early morning are due to the combined effects of suppressed boundary layer mixing processes and chemical loss by reaction with NO.

3.1.3 Site-specific variation of ground level ozone concentration

Site III (DVC-DTPS residential area) shows the highest hourly averaged ozone concentration

followed by sites II and I. Urban quenching (Schneider et al., 1988) appears to play a major role in the high ozone levels measured in site II. This seems to occur because there is so much NO in the atmosphere over sites I and III (from vehicular and industrial emissions, respectively) that ozone reacts immediately with this compound ($\text{NO} + \text{O}_3 \rightarrow \text{NO}_2 + \text{O}_2$). Table II shows that sunset time ranged from 18:08 to 18:12 LT over site II during the sampling period. The source of nocturnal ozone might be the long range of ozone transport. However, high concentrations of ozone over site II persist even after sunset (Fig. 4).

A ten days air-mass backward trajectory simulation was performed to ascertain the source and transport pathways of gaseous pollutants. The back trajectory analyses depicting the air mass trajectory at three different levels (100, 500 and 1500 m) over the three sites are shown in Fig. 6a, b, c, respectively.

Figure 6b shows that during the study period air masses at 100 and 500 m predominantly originated from the Bay of Bengal, entered the continental land mass through Andhra Pradesh and reached site II after travelling through Jharkhand, Chattisgarh and other provinces. These are highly industrialized areas with plenty of iron, steel and chemical industries, power plants, etc., which are known to be major sources of most of the precursors of daytime ozone. The air mass at 1500 m originated from Iran and traveled through arid regions of Pakistan and the desert of Rajasthan, entering site II. During its passage over the continental landmass, it was exposed to local pollution and accumulated precursors of daytime ozone, which resulted

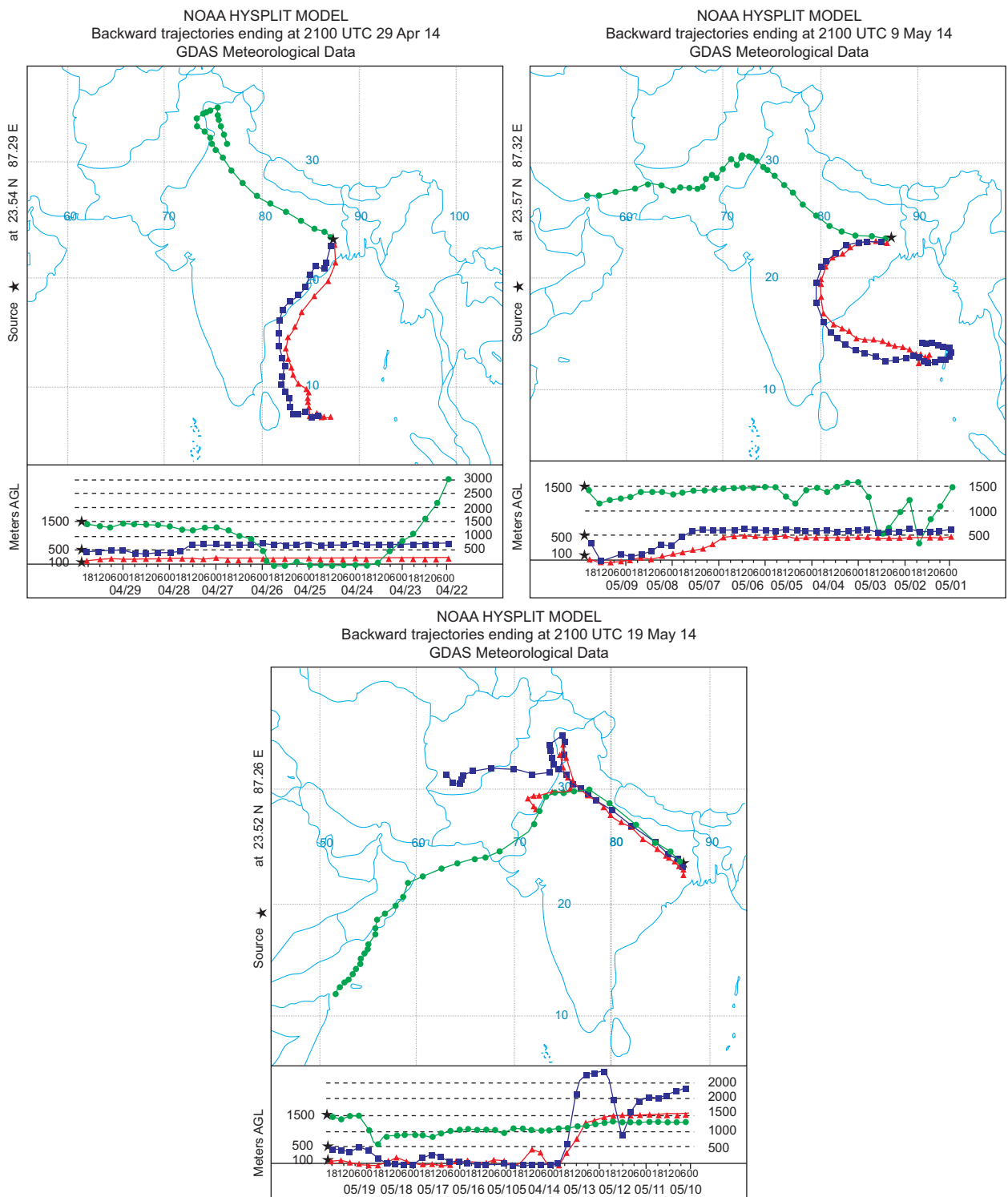


Fig. 6. Back trajectory analyses depicting the air mass trajectory at three different levels (100, 500 and 1500 m) over (a) site I, (b) site II and (c) site III.

in the formation of ozone under favorable meteorological conditions. This might be the reason of enhanced ground level ozone over site II. The wind rose diagram in Figure 7b shows that wind predominantly flowed in the W-SW direction (244°) over site II during the study period, which further confirms the transport of ozone from other regions. Site III is surrounded by a large number of industries (iron, steel, and alloy industries, among other industrial facilities), which are the major sources of daytime ozone precursors. The presence of high concentrations of precursors along with favorable meteorological conditions enhances the photochemical production of daytime ozone. Moreover, the back trajectory simulation depicts that the air mass at 100 and 500 m originated from arid regions in Rajasthan and Afghanistan, respectively, and reached site III after passing over Jammu, Kashmir, Uttar Pradesh, and Bihar. The air mass at 1500 m originated at the Arabian Sea near Somalia, crossed it and passed over Rajasthan, Uttar Pradesh and Bihar, arriving to site III. The high concentrations of ground level ozone at site III may be due to these air masses. The wind rose diagram of site III (Fig. 7c) shows that the dominant wind direction is W-NW (292°). The 10-day back trajectory air mass simulation clearly depicts that air masses at 100 and 500 m have mostly passed over the Bay of Bengal before reaching site I, and they have not circulated over the continental regions of India long-enough to be exposed to local pollution, which results in low ozone concentrations over site I. The wind rose diagram of site I (Fig. 7a) further confirms that wind predominantly flows in the S-SW direction (209°).

3.1.4 Daily variation of ground level ozone and PBL characteristics

The daily variation of ground level ozone concentrations along with daily averages of PBL height and vertical mixing coefficients over sites I, II, and III are shown in Figure 8a, b, c, respectively. Apart from in situ photochemical production of tropospheric ozone, this element can be transported downward from the stratosphere (Reddy et al., 2012). Tropospheric photochemical production of ozone is the dominant source while boundary layer dynamics plays a crucial role in controlling the ground level ozone concentration. The vertical mixing coefficient is the measure of turbulent mixing within the boundary layer and gives an idea

of vertical movement of the pollutants. It is actually mixing length times the velocity of wind within the mixing length, which can be defined as the depth, measured upward from the earth's surface through which pollutants are vigorously mixed (Rao et al., 2003). The vertical mixing coefficient determines the air pollution potential of an area. The PBL starts evolving after sunrise, attains a maximum height during afternoon due to surface heating and again decays after sunset. During afternoon hours, the turbulent mixing in the convective boundary layer enables the vigorous mixing of the chemical species present within the boundary layer depth. This process significantly affects the ozone distribution near the surface and above. Day to day variations of surface ozone concentration, PBL height and vertical mixing coefficient are found to have reasonable agreement (Fig. 8a, b, c). Comparatively, tall box plots portray greater variation of surface ozone. Lower PBL heights lead to lesser vertical mixing which in turn leads to greater ozone loss by surface deposition and restrict the mixing of air masses poor in ozone at the surface with air masses rich in ozone at higher heights. Similar result were found by Reddy et al. (2012) at Gadanki, India.

3.2 Influence of meteorology on the variation of air pollutants

The variations of wind speed and wind direction over the three sites are shown in Figure 7a, b, c, respectively. The dominant wind direction is shown by the resultant vector in the respective diagram.

Computation of the Pearson correlation coefficient shows that O_3 concentration has positive and negative correlations with temperature and humidity, respectively. Thus, in presence of high humidity and low temperature, O_3 production remains low. The positive correlation between temperature and O_3 is due to the fact that radiation controls temperature, which in turn increases the photolysis efficiency. The major photochemical paths for removal of O_3 are enhanced in presence of high humidity. As higher humidity levels are associated with large cloud cover and atmospheric instability, the photochemical process slows down and surface O_3 is depleted by deposition on water droplets. Simple regression analysis between the concentration of surface O_3 and cloud cover at sites I, II, and III at 00, 03, 06, 09, 12, 15, 18 and 21 UTC are presented in Figures 9-11, respectively.

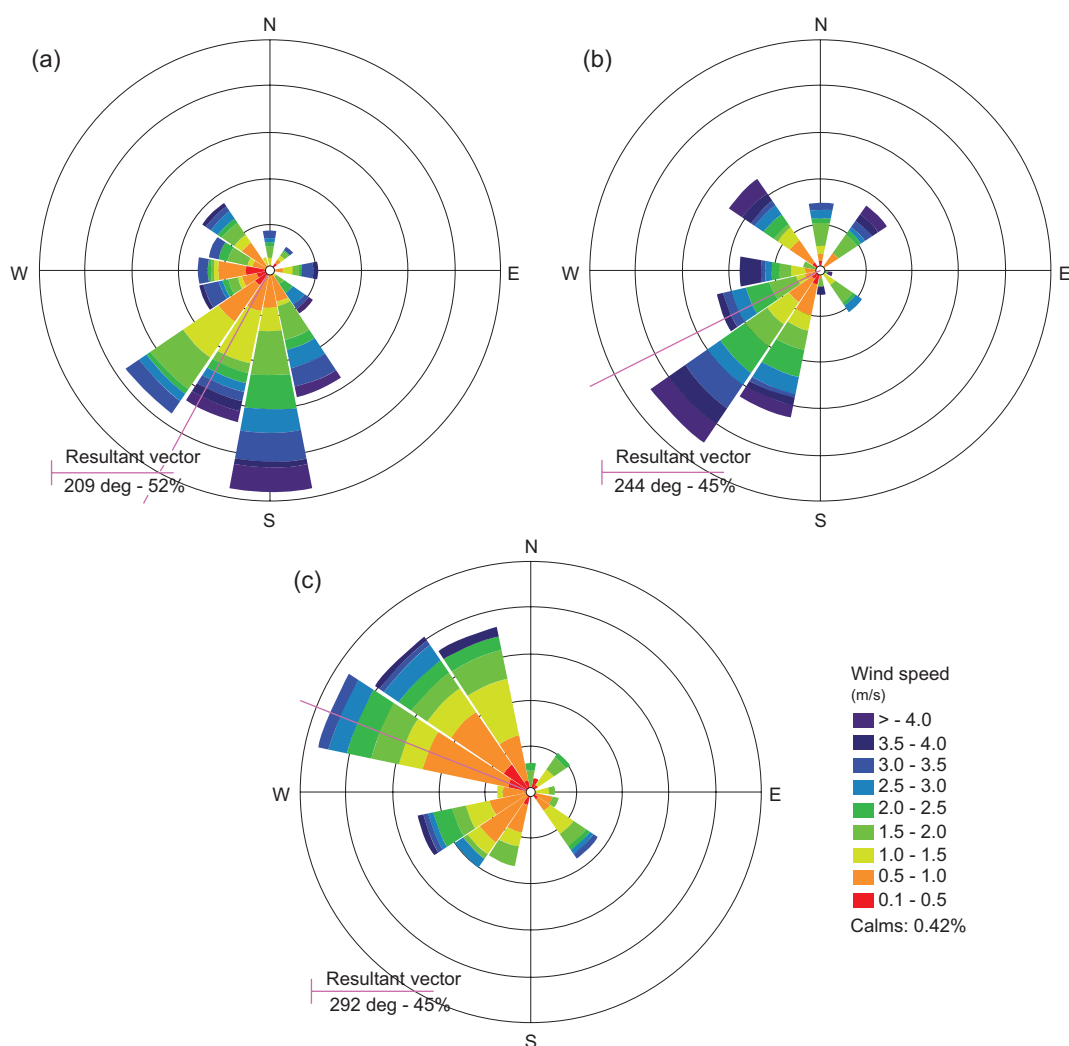


Fig. 7. Wind rose diagrams showing the variations of wind speed and wind direction over (a) site I (April 19-29, 2014), (b) site II (April 30-May 9, 2014), and (c) site III (May 10-19, 2014).

Since there is a time lag between the formation of O_3 and cloud coverage, ozone data from 30 min after cloud coverage are considered for computation.

It appears from the figures that the coefficient of determination (R^2) shows a very weak relationship between surface ozone and cloud cover. A similar result was found by Ghosh et al. (2015) over Kolkata, India. In the case of site I, negative relationships are observed at 03, 06 and 09 UTC, which indicates that after sunrise (daytime hours) the surface ozone concentration is regulated by the photo-oxidation of its precursors in presence of suitable meteorological factors. But during late afternoon and nighttime,

cloud coverage does not play a significant role in controlling the surface ozone concentration.

At site III, the correlation between ozone and cloud cover is found to be negative at 06, 09 and 12 UTC. This indicates that the amount of surface ozone is driven by photochemical production during daytime as higher cloud coverage leads to lesser amount of ozone at the surface. But at night the concentration of surface ozone is a complex function of concentration of precursors, boundary layer characteristics and other meteorological factors. Horizontal wind transport also plays a crucial role in either accumulation or dispersion of surface ozone at all the sites.

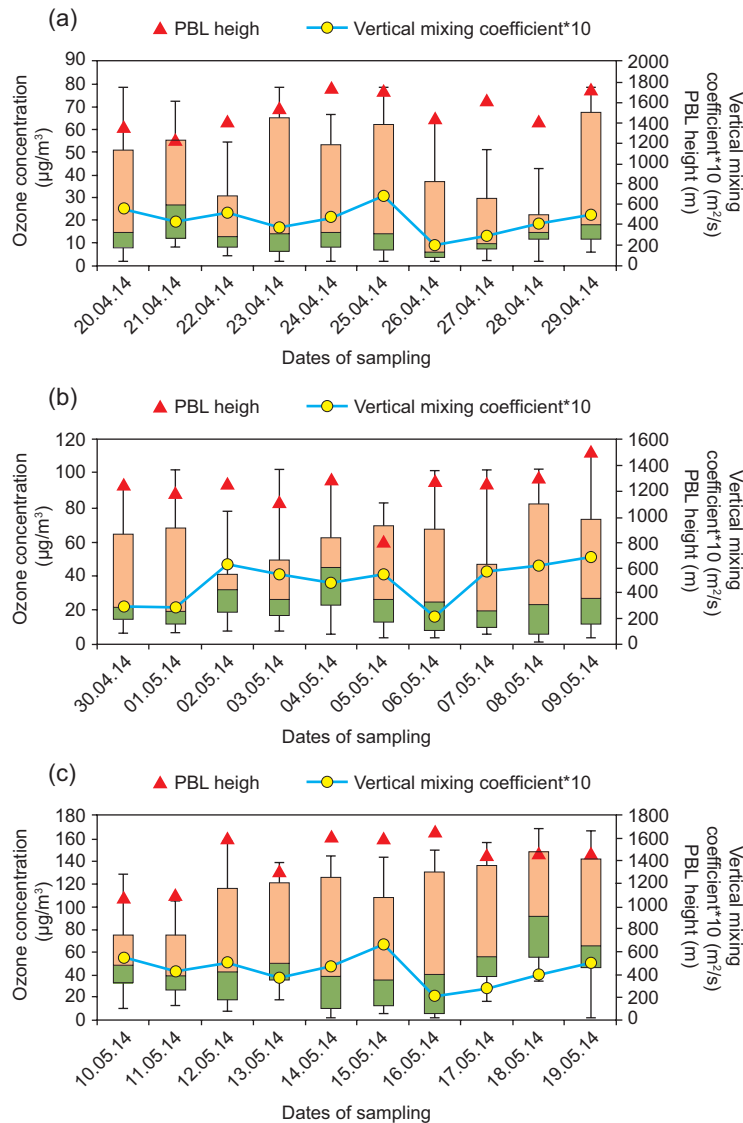


Fig. 8. Variation of ground level ozone concentration along with average PBL height and vertical mixing coefficient during the sampling period over (a) site I, (b) site II and (c) site III.

An inverse correlation with humidity is found in all the air pollutants under consideration. Significant negative correlation of humidity is observed with concentration of PM_{10} ($r = -0.514$ and $p = 0.004$) and CO ($r = -0.672$ and $p < 0.0001$). This inverse relationship between humidity and PM_{10} is due to the effect of humidity on coalescence and the settling of fine suspended particles. The lifetime of the PM in ambient air increases in low humidity, resulting in higher concentration of PM_{10} in the atmosphere.

Temperature is found to hold a positive correlation with SO_2 , NO_2 , NH_3 , CO and PM_{10} . A positive relationship of ambient NH_3 with temperature can be supported with by the fact that a rise in temperature increases evaporation of ammonia from PM and other sources present in the urban areas. This observation is further corroborated by the work of Perrino et al. (2002). All air pollutants hold an inverse relation with wind speed, which has a significant negative correlation with SO_2 ($r = -0.526$ and $p = 0.003$), NO_2

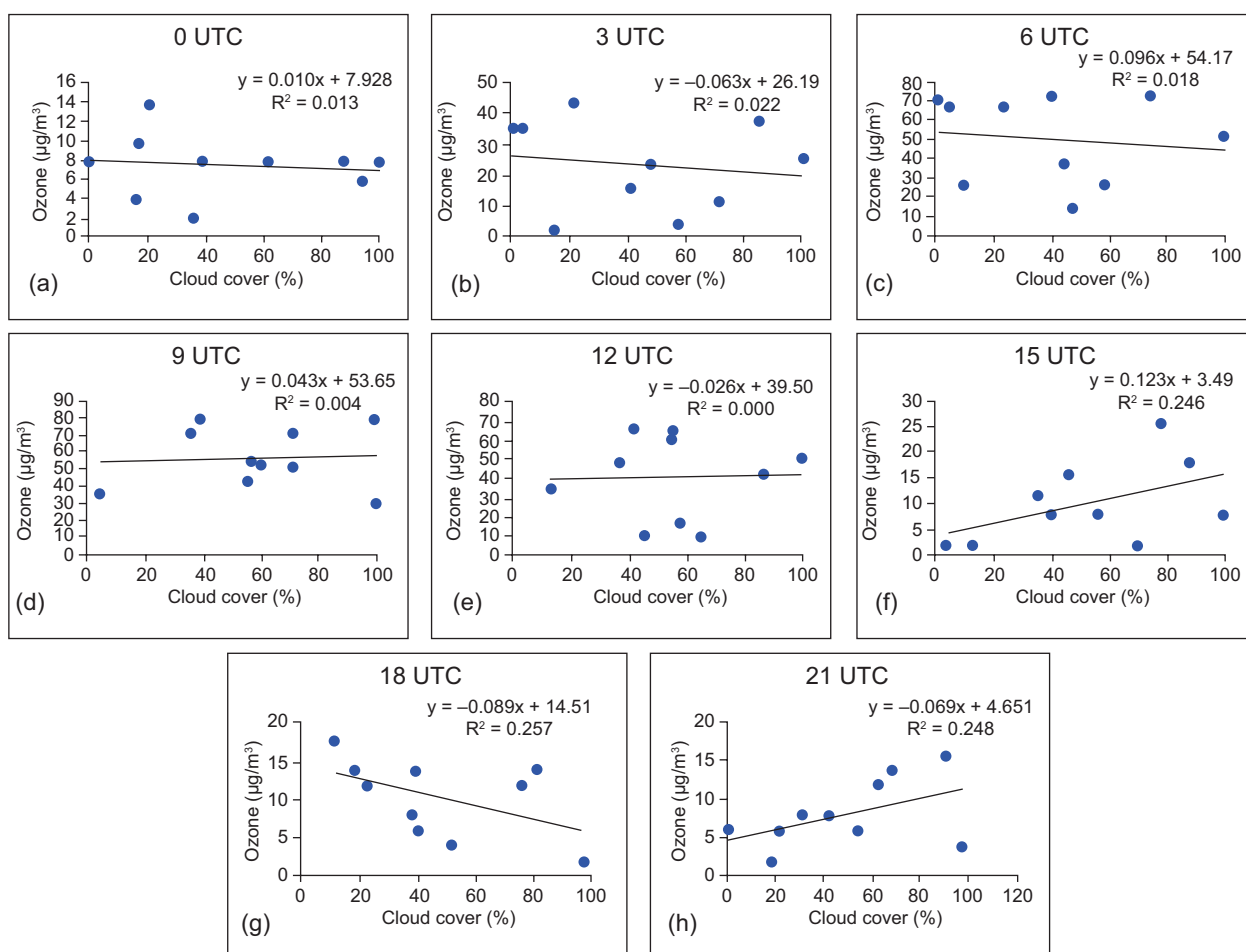


Fig. 9. Simple regression between surface ozone ($\mu\text{g m}^{-3}$) and cloud coverage (%) at (a) 00 UTC, (b) 03 UTC, (c) 06 UTC, (d) 09 UTC, (e) 12 UTC, (f) 15 UTC, (g) 18 UTC, and (h) 21 UTC over site I.

($r = -0.468$ and $p = 0.009$), PM_{10} ($r = -0.524$ and $p = 0.003$) and CO ($r = -0.524$ and $p = 0.003$). High wind speed helps in the dispersion and dilution of air pollutants thereby reducing the pollution load near the earth's surface.

3.3 Multivariate statistical analysis

Factors generated by multivariate statistical analysis give an insight into the main anthropogenic processes (industrial activities and vehicular emissions) that affect the concentration of different criteria pollutants in the ambient atmosphere. Principal component analysis on the composite data set has resulted in factors D1 and D2 (with eigen value > 1) that explain percentage of variability of data. Table III shows the

loading of each variable within two selected factors. Factor 1 (D_1) shows that loading of PM_{10} , SO_2 , CO , O_3 and NO_2 can be partially attributed to industrial emission.

Although O_3 is not directly emitted by nearby industries, the high concentrations of precursor gases emitted by these industries (along with suitable meteorological conditions) produce O_3 (a secondary pollutant). Factor 2 (D_2) shows a higher loading of NO_2 and NH_3 along with moderate loading of CO resulting from vehicular emissions.

3.4 Air quality index

Site-specific statistical summaries of air pollutants and AQIs are given in Table IV.

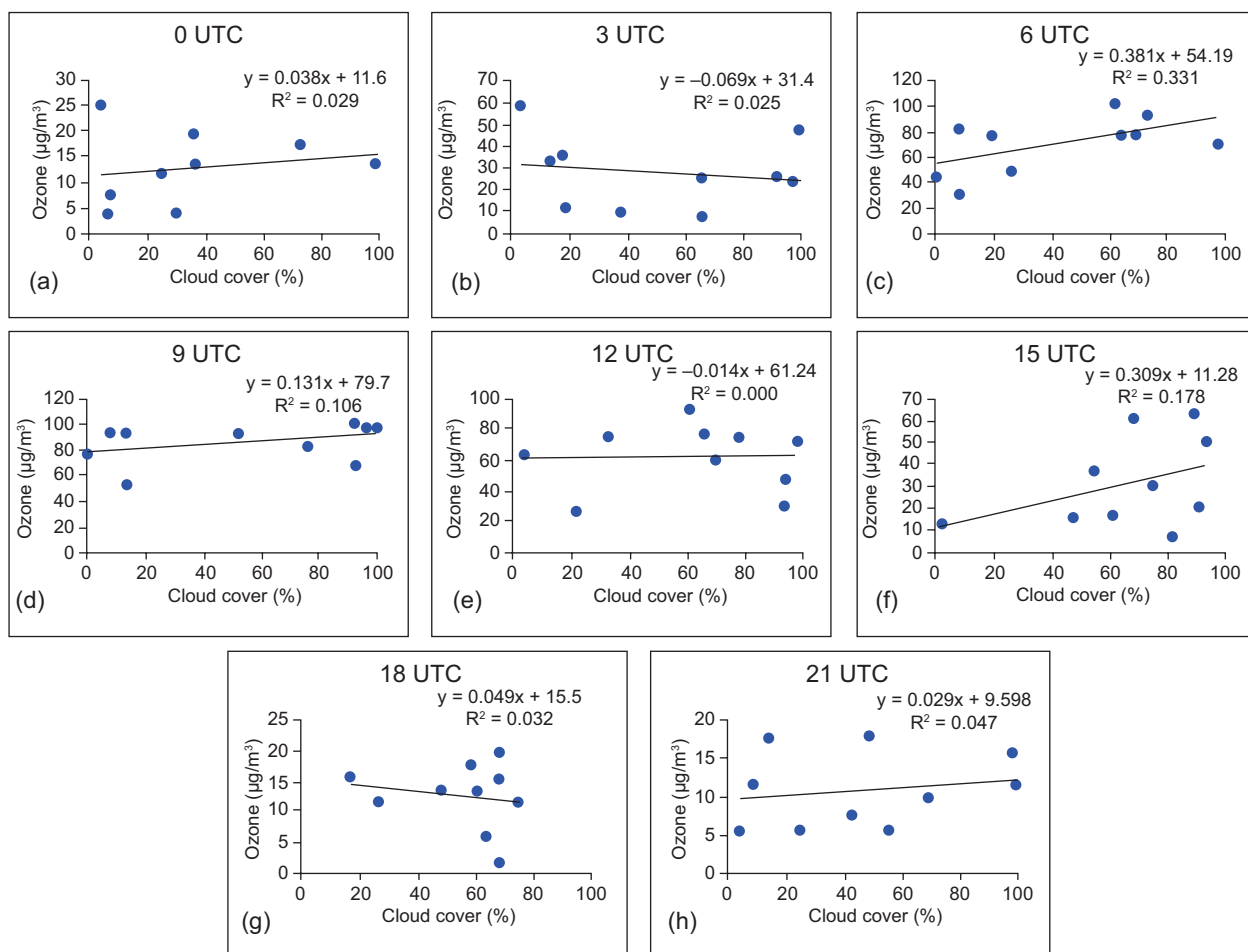


Fig. 10. Simple regression between surface ozone ($\mu\text{g m}^{-3}$) and cloud coverage (%) at (a) 00 UTC, (b) 03 UTC, (c) 06 UTC, (d) 09 UTC, (e) 12 UTC, (f) 15 UTC, (g) 18 UTC, and (h) 21 UTC over site II.

The site-specific statistical summary of air pollutants also supports the abundance of vehicular emissions (i.e., NO_2 , NH_3 and CO) and industrial emissions (i.e., PM_{10} , SO_2 , NO_2 and CO) in the ambient atmosphere of sites I and III, respectively. AQI calculations (Mudri, 1990) show that site II (residential area) is clean ($\text{AQI} = 22.69$). Both site I (traffic congested area) and site III (industrial area) are found to fall in a moderately polluted category, though the latter is more polluted than site I.

4. Conclusion

The present study focuses on the variation of different air pollutants (PM_{10} , SO_2 , CO , O_3 , NH_3 and NO_2), meteorological parameters (temperature, humidity, cloud cover, wind speed and wind direction) and PBL characteristics (height and vertical mixing coef-

ficient) in the ambient atmosphere of three different sites in Durgapur, an urban region in eastern India. The major conclusions are given below:

High ground level ozone is found in a residential area (site II). Backward trajectory simulations suggest the occurrence of long-range transport of ozone. Moreover, air masses containing precursor pollutants also produce ozone as they travel. So, systematic assessment and control strategies should be made for residential areas since ozone has many adverse effects on human health. The calculation of AQIs for the three different sites show that sites I (area with high vehicular density) and III (industrial site) are moderately polluted with AQIs = 53.85 and 71.68, respectively, while site II (a residential area) is clean ($\text{AQI} = 22.69$).

Meteorological variables and boundary layer characteristics strongly influence the chemistry and

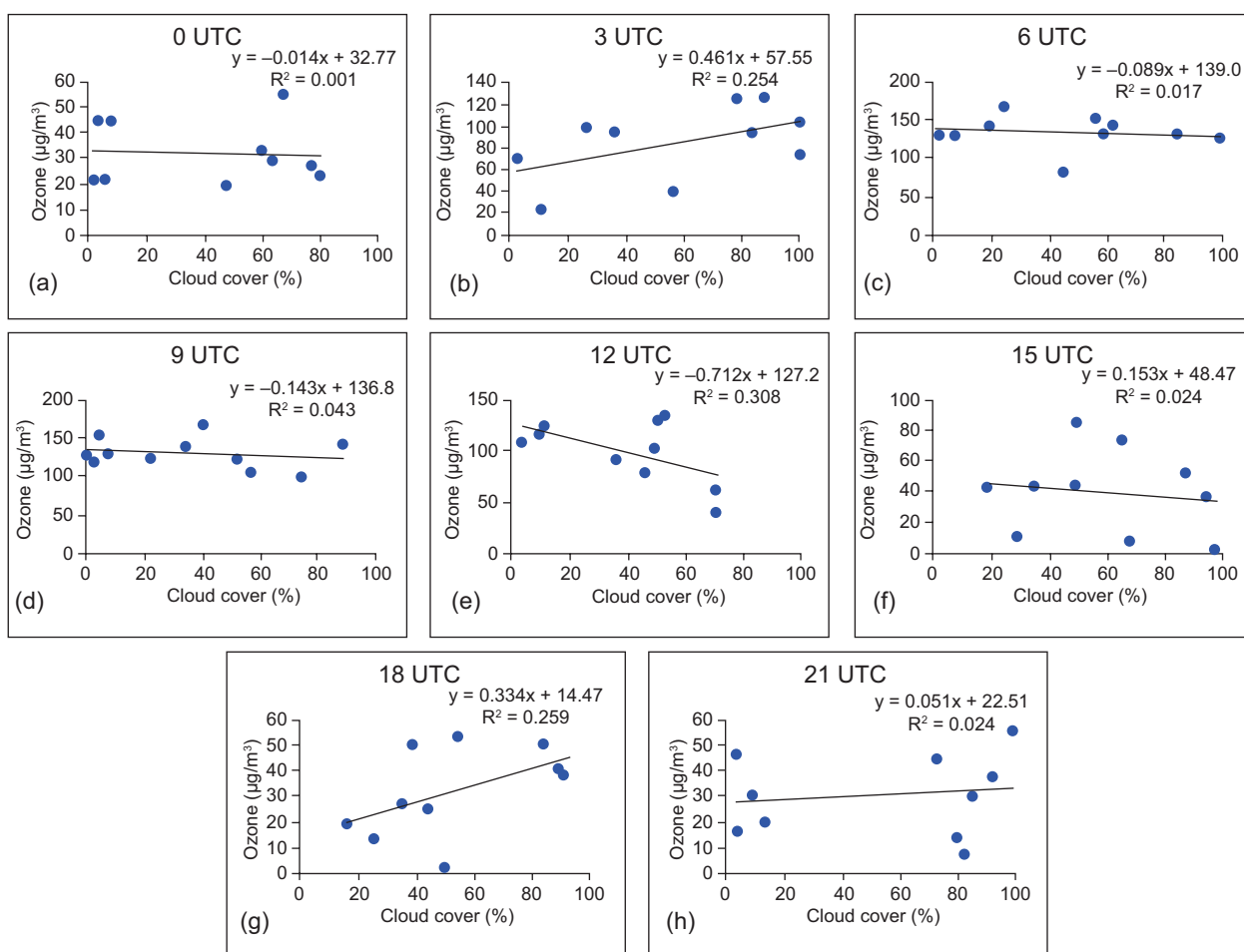


Fig. 11. Simple regression between surface ozone ($\mu\text{g}/\text{m}^3$) and cloud coverage (%) at (a) 00 UTC, (b) 03 UTC, (c) 06 UTC, (d) 09 UTC, (e) 12 UTC, (f) 15 UTC, (g) 18 UTC, and (h) 21 UTC over Site III.

Table III. Factor loadings after varimax rotation and percentage of variance after varimax rotation

Parameters	D1	D2
PM ₁₀	0.928	-0.069
SO ₂	0.793	0.136
NO ₂	0.402	0.814
O ₃	0.783	-0.494
CO	0.771	0.445
NH ₃	-0.196	0.875
Variability (%)	48.303	31.528
Cumulative (%)	48.303	79.831

Table IV. Statistical summaries and AQIs for sites I, II, and III.

Site I							
Statistics	O ₃ (μg m ⁻³)	SO ₂ (μg m ⁻³)	NO ₂ (μg m ⁻³)	NH ₃ (μg m ⁻³)	CO (μg m ⁻³)	PM ₁₀ (μg m ⁻³)	AQI
Mean	25.9	36.8	86.1	109.7	3.3	150.7	53.85 (moderately polluted)
Maximum	78.4	54.6	145.6	231.8	4.8	211.5	
Minimum	2.0	24.8	56.8	17.1	2.4	83.9	
Standard deviation	23.5	8.3	26.5	94.7	0.5	40.9	
Site II							
Mean	39.1	20.7	44.8	13.6	0.6	85.7	22.69 (clean)
Maximum	107.8	30.9	65.9	29.2	2.3	116.8	
Minimum	2.0	12.9	10.2	4.2	0.0	44.5	
Standard deviation	30.9	6.7	15.9	7.9	0.3	26.2	
Site III							
Mean	70.1	174.1	77.7	17.5	3.8	392.1	71.68 (moderately polluted)
Maximum	169.1	390.0	129.3	32.5	5.5	499.0	
Minimum	2.0	83.0	34.1	8.0	3.0	247.8	
Standard deviation	49.1	108.7	32.2	7.6	0.4	82.8	

dispersion of pollutants in the PBL. This work has identified the major pollutant sources over the three different sites and presents an idea of prevailing air quality status of different regions of this urban area.

The creation of comprehensive emission inventories of this urban area will further help to identify specific pollutant sources around the chosen sites, thereby helping in the formulation of more effective urban air quality management strategies.

Acknowledgments

The authors are thankful to the University of Kwa-Zulu Natal, the DST-NRF Centre of Excellence in Mathematical and Statistical Sciences (CoE-MaSS) and the University of Burdwan for supporting this research work. The authors also convey thanks to the anonymous reviewers for evaluating the manuscript and providing their valuable suggestions.

References

- Ali M. and Athar M., 2008. Air pollution due to traffic, air quality monitoring along three sections of national highway N-5, Pakistan. *Environ. Monitor. Assess.* 136, 219–226. doi: 10.1007/s10661-007-9677-3
- Azad A.K. and Kitada T., 1998. Characteristics of air pollution in the city of Dhaka, Bangladesh in winter. *Atmos. Environ.* 32, 1991–2005. doi: 10.1016/S1352-2310(97)00508-6
- Begum B.A., Kim E., Biswas S.K. and Hoopke P.K., 2004. Investigation of sources of atmospheric aerosol at urban and semi urban areas in Bangladesh. *Atmos. Environ.* 38, 3025–3038. doi: 10.1016/j.atmosenv.2004.02.042
- Behera S.N., Sharma M., Aneja V.P. and Balasubramanian R., 2013. Ammonia in the atmosphere: a review on emission sources, atmospheric chemistry and deposition on terrestrial bodies. *Environ. Sci. Pollut. R.* 20, 8092–8131. doi: 10.1007/s11356-013-2051-9
- Chattopadhyay S., Gupta S. and Saha R.N., 2010. Spatial and temporal variation of urban air quality: A GIS approach. *J. Environ. Prot.* 1, 264–277. doi: 10.4236/jep.2010.13032
- Coen M.C., Praz C., Haeefe A., Ruffieux D., Kaufmann P. and Calpini B., 2014. Determination and climatology of the planetary boundary layer height above the Swiss plateau by in situ and remote sensing measurements as well as by the COSMO-2 model. *Atmos. Chem. Phys.* 14, 13205–13221. doi: 10.5194/acp-14-13205-2014
- Draxler R.R. and Hess G.D., 1998. An overview of the HYSPLIT_4 modelling system for trajectories, dispersion, and deposition. *Aust. Meteorol. Mag.* 47, 295–308.
- Dey S., Gupta S. and Mahanty U., 2014. Study of particulate matters, heavy metals and gaseous pollutants at Gopalpur

- (23° 29' 52.67" N, 87° 23' 46.08" E), a tropical industrial site in eastern India. *IOSR Journal of Environmental Science, Toxicology and Food Technology* 8, 01-13.
- Dey S., Gupta S., Sibanda P. and Chakraborty A., 2017. Spatio-temporal variation and futuristic emission scenario of ambient nitrogen dioxide over an urban area of eastern India using GIS and coupled AERMOD–WRF model. *PLoS ONE* 12, e0170928. doi: 10.1371/journal.pone.0170928
- Dockery D.W., Pope C.A., Xu X., Splender J.D., Ware J.H., Fay M.E., Ferris B.G. and Speizer F.E., 1993. An association between air pollution and mortality in six US cities. *New Engl. J. Med.* 329, 1753-1759. doi: 10.1056/NEJM199312093292401
- Ghosh D., Lal S. and Sarkar U., 2013. High nocturnal ozone levels at a surface site in Kolkata, India: Trade-off between meteorology and specific nocturnal chemistry. *Urban Climates* 5, 82-103. doi: 10.1016/j.uclim.2013.07.002
- Ghosh D. and Sarkar U., 2015. Analysis of the photochemical production of ozone using tropospheric ultraviolet-visible (TUV) radiation model in an Asian megacity. *Air Qual. Atmos. Hlth.* doi: 10.1007/s11869-015-0346-3
- Ghosh D., Midya S.K., Sarkar U. and Mukherjee T., 2015. Variability of surface ozone with cloud coverage over Kolkata, India. *J. Earth Sys. Sci.* 124, 303-319. doi: 10.1007/s12040-015-0543-3
- Ghose M.K., Paul R. and Banerjee S.K., 2004. Assessment of the impacts of vehicular emissions on urban air quality and its management in Indian context: The case of Kolkata (Calcutta). *Environ. Sci. Policy* 7, 345-351. doi: 10.1016/j.envsci.2004.05.004
- Goyal P. and Sidhartha., 2003. Present scenario of air quality in Delhi: A case study of CNG implementation. *Atmos. Environ.* 37, 5423-5431. doi: 10.1016/j.atmosenv.2003.09.005
- Gupta A.K., Karar K., Ayoob S. and John K., 2008. Spatio-temporal characteristics of gaseous and particulate pollutants in an urban region of Kolkata, India. *Atmos. Res.* 87, 103-115. doi: 10.1016/j.atmosres.2007.07.008
- Jacobs M.B. and Hochheiser S., 1958. Continuous sampling and ultra microdetermination of nitrogen dioxide in air. *Anal. Chem.* 30, 426-428. doi: 10.1021/ac60135a032
- Jain M.K. and Saxena N.C., 2002. Air quality assessment along Dhanbad-Jharua road. *Environ. Monitor. Assess.* 79, 239 -250. doi: 10.1023/A:1020278610521
- Kaiser H.F., 1958. The varimax criterion for analytical rotation in factor analysis. *Psychometrika* 23, 187-200. doi: 10.1007/BF02289233
- Kaushik C.P., Ravindra K., Yadav K., Mehta S. and Hari-tash A.K., 2006. Assessment of ambient air quality in urban centers of Haryana (India) in relation to different anthropogenic activities and health risks. *Environ. Monitor. Assess.* 122, 27-40. doi: 10.1007/s10661-005-9161-x
- Koken P.J., Piver W.T., Ye F., Elixhauser A., Olsen L.M. and Portier C.J., 2003. Temperature, air pollution and hospitalization for cardiovascular diseases among elderly people in Denver. *Environ. Health Persp.* 111, 1312 -1317.
- Kolev N., Savov P., Donev E., Ivanov D., Evgenieva T., Grigorieva V. and Kolev I., 2011. Boundary layer development and meteorological parameters impact on the ground level ozone concentration over an urban area in a mountain valley (Sofia, Bulgaria). *Int. J. Remote Sens.* 32, 8915-8933. doi: 10.1080/01431161.2010.524679
- Lal S., Naja M. and Subbaraya B.H., 2000. Seasonal variations in surface ozone and its precursors over an urban site in India. *Atmos. Environ.* 34, 2713-2724. doi: 10.1016/S1352-2310(99)00510-5
- Lal S. and Patil R.S., 2001. Monitoring of atmospheric behaviour of NO_x from vehicular traffic. *Environ. Monitor. Assess.* 68, 37-50. doi: 10.1023/A:1010730821
- Lodge J.P. Jr., 1998. *Methods of air sampling and analysis*, 3rd ed. CRC Press, USA, 784 pp.
- Mudri S.S., 1990. Categorisation of ambient air quality index. *Indian J. Environ. Prot.* 10, 424-427.
- Naja M. and Lal S., 1997. Solar eclipse induced changes in surface ozone at Ahmedabad. *Indian J. Radio Space* 26, 312-318.
- Perrino C., Catrambone M., Di Menno Di Bucchianico A. and Allegrini I., 2002. Gaseous ammonia in the urban area of Rome, Italy and its relationship with traffic emissions. *Atmos. Environ.* 36, 5385-5394. doi: 10.1016/S1352-2310(02)00469-7
- Pulikesi M., Skaralingam B.P., Elango D., Rayudu V.N., Ramamurthi V. and Sivanesan S., 2006. Air quality monitoring in Chennai, India, in the summer of 2005. *J. Hazard. Mater.* 136, 589-596. doi: 10.1016/j.jhazmat.2005.12.039
- Quan J., Gao Y., Zhang Q., Tie X., Cao J., Han S., Meng J., Chen P. and Zhao D., 2013. Evolution of planetary boundary layer under different weather conditions, and

- its impact on aerosol concentrations. *Particuology* 11, 34-40. doi: 10.1016/j.partic.2012.04.005
- Rao S.T., Ku J.Y., Berman S., Zhang K. and Mao H., 2003. Summertime characteristics of atmospheric boundary layer and relationships to ozone level over eastern United States. *Pure Appl. Geophys.* 160, 21-55. doi: 10.1007/s00024-003-8764-9
- Reddy G.S. and Ruj B., 2003. Ambient air quality status in Raniganj-Asansol area, India. *Environ. Monitor. Assess.* 89, 153-163. doi: 10.1023/A:1026070506481
- Reddy K.K., Naja M., Ojha M., Mahesh P. and Lal P., 2012. Influences of the boundary layer evolution on surface ozone variations at a tropical rural site in India. *J. Earth Sys. Sci.* 121, 911-922. doi: 10.1007/s12040-012-0200-z
- Salam A., Bauer H., Kassin K., Ullah S.M. and Puxbaum H., 2003. Aerosol chemical characteristics of a mega-city in southeast Asia (Dhaka Bangladesh). *Atmos. Environ.* 37, 2517-2528. doi: 10.1016/S1352-2310(03)00135-3
- Salam A., Hossain T., Siddique M.N.A. and Shafiqul Alam M.A., 2008. Characteristics of atmospheric trace gases, particulate matter, and heavy metal pollution in Dhaka, Bangladesh. *Air Qual. Atmos.* 1, 101-109. doi: 10.1007/s11869-008-0017-8
- Lee S.D., Wolters G.J.R., Grant L.D. and Schneider T. (Eds.), 1988. *Atmospheric ozone research and its policy implications*. Elsevier Science Publishing Company, New York, 1044 pp. (Studies in Environmental Science, 35).
- Verma A., Singh S.N. and Shukla M.K., 2003. Air quality of the Trans-Gomati area of Lucknow city, India. *B. Environ. Contam. Tox.* 70, 166 -173. doi: 10.1007/s00128-002-0171-x
- Walker J.T., Whitall D.R., Robarge W. and Paerl H.W., 2004. Ambient ammonia and ammonium aerosol across a region of variable ammonia emission density. *Atmos. Environ.* 38 1235-1246. doi: 10.1016/j.atmosenv.2003.11.027
- West P.D. and Gaeke G.C., 1956. Fixation of sulphur dioxide as sulfitomercurate (II) and subsequent colorimetric estimation. *Anal. Chem.* 28, 1816-1819. doi: 10.1021/ac60120a005
- Whitehead J.D., Longley I.D. and Gallagher M.W., 2007. Seasonal and diurnal variation in atmospheric ammonia in an urban environment measured using a quantum cascade laser absorption spectrometer. *Water Air Soil Poll.* 183, 317-329. doi: 10.1007/s11270-007-9381-5