



Atmósfera

ISSN: 0187-6236

editor@atmosfera.unam.mx

Universidad Nacional Autónoma de México
México

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Atmósfera, vol. 31, no. 4, 2018, October-December, pp. 373-387

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DOI: <https://doi.org/10.20937/ATM.2018.31.04.05>

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Partitioning of the water soluble versus insoluble fraction of trace elements in the city of Santiago, Chile

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Received: December 18, 2017; accepted: August 17, 2018

RESUMEN

Se investigó la composición elemental total y la fracción soluble en agua de PM₁₀ en tres áreas urbanas diferentes de Santiago de Chile, del centro a los suburbios. Las muestras de PM₁₀ recogidas durante el mes de mayo (mediados del otoño en el hemisferio sur) en 2006, 2008, 2009 y 2010 fueron analizadas para metales mayores y trazas, y la partición entre fracciones insolubles y solubles se determinó para la mayoría de ellos. Las concentraciones medias de PM₁₀ variaron de 71 µg m⁻³ (Cerrillos) a 128 µg m⁻³ (La Pintana), valores que se encuentran dentro de los rangos estacionales observados en Santiago. Se cuantificaron 25 elementos principales y trazas (Fe, Al, Ca, K, Mg, P, Pb, S, Ti, Mn, Cu, Zn, Ba, Zr, Cr, As, Sn, Sb, Ni, V, Li, Co, Cd, La y Rb) en el presente estudio. Ba, Sb, Cd, As y Zn fueron los elementos más solubles, con proporciones que varían de 50% a 98% de solubilidad. Por el contrario, los oligoelementos menos solubles fueron Ti, Sn, Pb y Cr. La mayor parte de los metales traza de alta solubilidad no están solamente fuertemente ligados a la emisión de tubos de escape vehiculares, sino a emisiones vehiculares como el sistema de frenos o a ciertas fuentes industriales. Nuestros resultados evidencian la cantidad significativa de oligoelementos solubles en la atmósfera urbana de Santiago clasificados como tóxicos y/o carcinógenos, lo que sugiere un impacto no despreciable en la salud.

ABSTRACT

The total elemental composition and the water-soluble fraction of PM₁₀ from three different urban areas in Santiago, Chile, from downtown to the suburbs, were investigated. PM₁₀ samples collected during the month of May (mid-autumn in the Southern Hemisphere) in 2006, 2008, 2009, and 2010 were analyzed for major and trace metals, and the partitioning between the insoluble and soluble fractions was determined for most of them. PM₁₀ average concentrations ranged from 71 µg m⁻³ (Cerrillos) to 128 µg m⁻³ (La Pintana), which are within the seasonal ranges observed in Santiago. Twenty five major and trace elements (Fe, Al, Ca, K, Mg, P, Pb, S, Ti, Mn, Cu, Zn, Ba, Zr, Cr, As, Sn, Sb, Ni, V, Li, Co, Cd, La, and Rb) were determined in the present study. Ba, Sb, Cd, As, and Zn, with proportions in the soluble fraction varying from 50% to 98%, were the most soluble elements. On the contrary, the less soluble trace elements were Ti, Sn, Pb, and Cr. Most of the high-solubility trace metals are strongly linked to non-exhaust traffic emissions, as well as to certain industrial sources. Our results evidence the significant amount of soluble trace elements in Santiago's urban atmosphere classified as toxic and/or carcinogenic, thus suggesting a non-negligible health impact.

Keywords: Megacity, PM₁₀, trace metals, road traffic, atmospheric pollution.

1. Introduction

Santiago, the capital of Chile, is one of the cities with the highest atmospheric pollution levels in Latin America and among the most polluted worldwide, similar to Mexico City, Delhi, or Beijing (Molina and Molina, 2004; WHO, 2013). The city of Santiago concentrates 40% of the country's population, with around six million inhabitants and a population density of 450 people per km² (INE, 2017). The increasing urban population, together with the geographic location and prevailing weather conditions (unfavorable for ventilation) account for such high levels of atmospheric pollution (Toro et al., 2014). In spite of the adoption of certain abatement measures to reduce such high atmospheric pollution levels, which included the renewal of public transport, the improvement of the chemical quality of fuels by decreasing their proportion of sulfur, benzene, olefins, and aromatic compounds, the renewal of the truck fleet, the implementation of standards for light vehicles and industry, or the adoption of mitigation actions in dust production, PM₁₀ and PM_{2.5} concentrations are still above national air quality standards, especially during the autumn and winter months (www.sinca.mma.gob.cl) (Mena-Carrasco et al., 2014). Those measures have been progressively implemented since 1997.

Atmospheric particulate matter is a pollutant whose analysis can be addressed from different perspectives and/or attending to diverse characteristics, including mass concentrations, size distribution, particle shape or chemical composition. These characteristics are strongly related to the kinds and magnitudes of the effects on health (Pope and Dockery 2006; Franklin et al., 2008; Sarnat et al., 2008; Bell et al., 2009; Peng et al., 2009; Ostro et al., 2011). Particulate matter may arise from different natural and anthropogenic sources, and its chemical composition is usually well connected with its origin. A number of studies have been published on the chemical composition of particulate matter in Santiago (Prendez et al., 1984; Didyk et al., 2000; Gil et al., 2000; Jorquera 2002; Koutrakis et al., 2005; Gramsch et al., 2006, 2009), and specific ones on the elemental chemical composition of PM₁₀ (Artaxo et al., 1999; Richter et al., 2007; Sax et al., 2007). The fine particulate matter (PM_{2.5}) of Santiago has a predominance of secondary organic and inorganic aerosols, reaching up to 60% (Seguel

et al., 2009). During the last few years, as a result of the implementation of certain measures to reduce PM levels, the elemental composition of PM in the city of Santiago has substantially changed. There has been an important decrease of sulfur and lead, due to the substantial improvement of the quality of fuels (desulfurization) and the use of unleaded gasoline. A similar pattern is seen with bromine. Inorganic substances, including trace metals, have been determined (Jorquera and Barraza, 2012). In particular, a study over an industrial district in Santiago during 1997–2003 showed that the labile or water-soluble fraction of PM₁₀ has relatively high concentrations of copper, zinc, manganese, cadmium and arsenic, elements clearly associated to anthropogenic emissions (industry, road traffic). Overall, a decline in PM concentrations has taken place in Santiago since 2000. Toro et al. (2014) studied the variability of particulate matter from 2000 to 2012 at four locations in the metropolitan region of Santiago. While the PM₁₀ and PM_{2.5} levels have declined since 2000, they still exceed the environmental quality standards. Annual average PM₁₀ concentrations range from 52 to 76 µg m⁻³, with the highest concentrations in winter.

It has been shown that fine particulate matter is responsible for increased mortality and morbidity of people in Santiago. Cakmak et al. (2007, 2009, 2011) and Valdés et al. (2012) concluded that the PM_{2.5} mass is not enough to explain the observed incidence of mortality, and they claimed for specific studies to identify the main sources of heavy metals such as Zn, Cr or Cu.

As discussed before, the chemical composition of PM is a better proxy than PM mass concentrations to assess health impacts. Furthermore, if we could introduce another parameter, such as the partitioning of trace elements between soluble and insoluble fractions, the epidemiological effects linked to particulate matter pollution could be better constrained.

The objective of this work was to determine the total elemental composition and the water-soluble fraction of so called heavy and trace elements present in PM₁₀ from three urban locations under different emission or source patterns in Santiago: downtown Santiago, a peripheral area, and an intermediate locations between the former. The choice of years for the study (2006, 2008, 2009 and 2010) is justified by a change in the type of public transport to one called

Transantiago (Gramsch et al., 2013), which replaced old buses with new ones fitted with nonpolluting technology, reduced circulation times and the number of vehicles, and established the use of exclusive lanes. However, the pattern of combustion sources has also varied in Santiago, especially over the last five years. Important sources of emission of fine PM are residential quarters, accounting for almost 30%, together with mobile sources. (Troncoso et al., 2012; MMA, 2014; Villalobos et al., 2015).

2. Experimental

2.1 Sampling zone

This study used samples from three monitoring stations in Santiago, located in Cerrillos (CR), La Pintana (LP) and Av. Ossa (OSS). The choice of these sites was due to the following: Cerrillos (lat. -33.491 , long. -70.719 , 21 km^2 and $83\,000$ inhabitants) is an industrial zone in the west side of the city; La Pintana (lat. -33.614 , long. -70.635 , 30.6 km^2 and $287\,000$ inhabitants) is a peripheral urban site, and Av. Ossa (lat. -33.449 , long. -70.571 , 16.9 km^2 and $170\,000$ inhabitants) is an intermediate area between the above. PM_{10} samples were obtained during the month of May (autumn) in 2006, 2008, 2009 and 2010. The choice of month is justified from previous studies (Jorquera, 2002; Gramsch et al., 2006), which found that May is a month with winter-like conditions (low temperatures, low wind speed and strong temperature inversions) but with little rain, so it is a time period in which environmental conditions do not change much. Figure 1 shows the location of these monitoring sites.

The wind pattern in Santiago de Chile has a valley-mountain behavior with generally low speeds. The predominant wind in the city occurs during the afternoon (11–19 LT) with a south-east direction, as shown in the wind roses of Figure 2. It can be seen that during the afternoon, the wind direction is the same for all stations and has little dispersion. During these hours, the speed is also higher than during the night or early morning. There is also a very clear difference between the day and night wind. Figure 3 shows the wind speed for all sites for the month of May 2009. The wind comes from the coast and enters Santiago from the south-west, therefore the highest speed is achieved in Cerrillos (see Fig. 3), and its velocity decreases as it enters the city. This wind brings

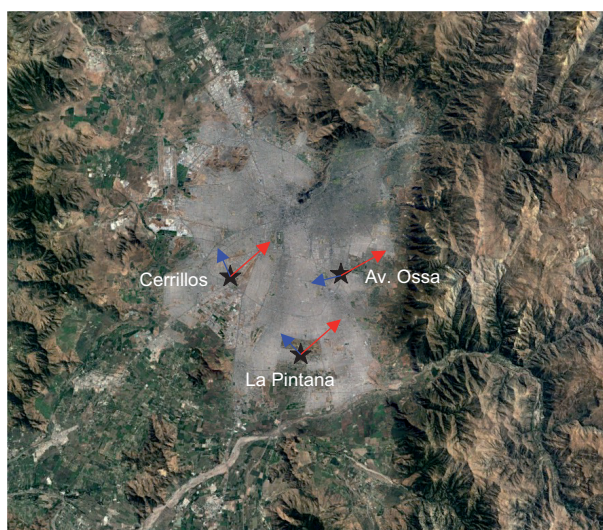


Fig.1. Map of Santiago, Chile, with the location of the three monitoring stations used for the study: Cerrillos (lat. -33.491 , long. -70.719 , located in the west of the city); La Pintana (lat. -33.614 , long. -70.635 , a peripheral urban site), and Av. Ossa (lat. -33.449 , long. -70.571 , an intermediate area). The red arrows indicate the predominant wind direction during the afternoon (11–19 LT) and the blue arrows show the wind direction at night and early morning.

clean air from the coast into the city (Gramsch et al., 2006) but it also carries pollutants from downtown Santiago towards the east. As shown in Figure 3, pollutants from La Pintana can be transported into Av. Ossa during the afternoon. At night and early morning, the wind direction reverses; however, the velocity is very low and the direction is not the opposite of the afternoon wind. Thus, there is probably no transport of pollutants from Av. Ossa towards the other sites at night.

2.2 Materials and methods

PM_{10} was collected on fiberglass filters ($20.32 \times 25.4 \text{ cm}$) during 24 h, using hi-vol. samplers (ECO-TECH model 3000) which absorbed 1600 m^3 of air in 24 h. All samples were provided and normalized by the Ministerio del Ambiente (Ministry of the Environment, www.mma.gob.cl) in collaboration with Chile's Ministerio de Salud (Ministry of Health).

Then, circular sections of each filter (1.53 cm^2) were obtained and merged to build-up a weekly composite sample. Overall, four samples were obtained

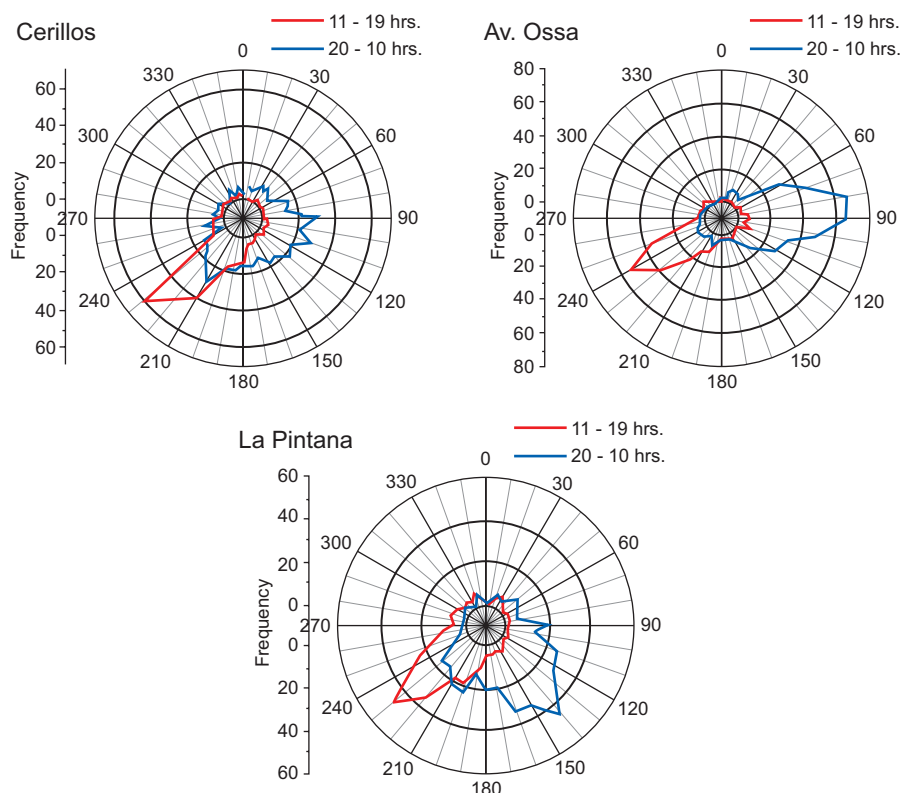


Fig. 2. Average wind roses in La Pintana, Cerrillos and Av. Ossa for May 2009. The red lines show the wind direction between 11 and 19 LT and the blue lines show the wind direction during the rest of the day.

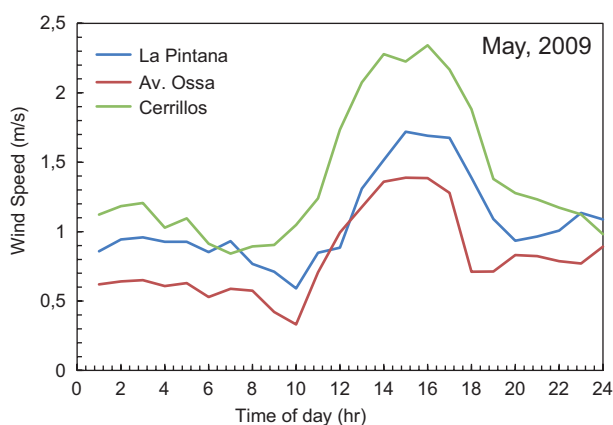


Fig. 3. Average wind speed in the three sites for May 2009.

per year at each urban observatory. The sampling was made in duplicate, analyzing 48 samples and 48 counter-samples.

2.3 Total metals extraction

Total metals extraction from the filters was carried out by the acid digestion in Teflon containers (EPA method). The samples were digested with 1.25 mL of nitric acid (Suprapur, Merck) and 2.5 mL of hydrochloric acid (Suprapur, Merck), and heated at 90 °C during 12 h. Then, 1.25 mL of HClO_4 was added and the mixture was heated for 3 to 6 h to dryness at 240 °C on a hotplate. The final residue was re-dissolved with 2.5 mL of nitric acid and the volume was completed to 25.0 mL with ultra-pure water. Finally, the solution was centrifuged for 20 min at 5000 rpm and stored at 4 °C until the chemical analysis was performed. Thereafter, the solutions were analyzed using an ICP-MS (X series II, THERMO).

In order to account for the filter contribution to individual concentrations of the elements and to ensure the analytical quality of our determinations, each set of samples was carried out together with a

number of blank filters and a blank filter with 0.01 g of a fly-ash reference material with certified concentrations (NIST-1633b).

The values in micrograms per liter were expressed in nanograms per cubic meter considering the volume of air absorbed during 24 h, the impact area of the filter (513 cm²), the particular area of the filters used for the analysis (1.53 cm²), and the number of filters analyzed.

2.4 Extraction of the soluble fraction of major trace elements

A duplicate of each composite sample was leached with 30 mL of ultrapure water inside PVC containers. The containers were then immersed in an ultrasonic bath during 10 min and placed in an oven at 60 °C for 6 h. The resultant solution was filtered through a 0.45 µm membrane filter, and finally analyzed by ionic chromatography (high performance liquid chromatography using a WATERS IC-pakTM anion column) for the determination of Cl⁻, NO₃⁻ and SO₄²⁻, and by a specific electrode (WATERS 432 conductivity detector) to obtain NH₄⁺ concentrations. Ten mL of the former solution were acidified with nitric acid (5%) and analyzed by ICP-AES to obtain the concentration of major and trace elements in the solution.

2.5 Statistical analysis

The number of samples was insufficient to run typical receptor models such as positive matrix factorization

(PMF). Hence, in this study the statistical analysis of the results was based on cluster analysis (CA), considering the total element concentrations. Data standardization of each variable was performed prior to statistical analysis because the concentration orders among the elements and the parameters differed greatly, and statistical results would be highly biased by elements with high concentration. The furthest neighbor method was applied for grouping of the different variables, and dissimilarity was defined by the square Euclidean distance. With this method we were not able to obtain sources and/or atmospheric processes; however, the results should be indicative of possible element associations when source apportionment studies are conducted.

3. Results and discussion

3.1 PM mass concentrations

In Santiago, Chile, PM₁₀ concentrations have decreased continuously since 1998, but the trend has slowed down since 2010. Figure 4a shows the PM₁₀ yearly average for the nine stations belonging to the air pollution network of the Sistema de Información Nacional de Calidad del Aire (National Air Quality Information System, SINCA) inside Santiago (SINCA, 2018). This network is used by the authority to monitor several air pollutants as well as meteorological variables. PM₁₀ is measured using TEOM monitors (Rupprecht & Patashnick, Albany, NY, and Thermo

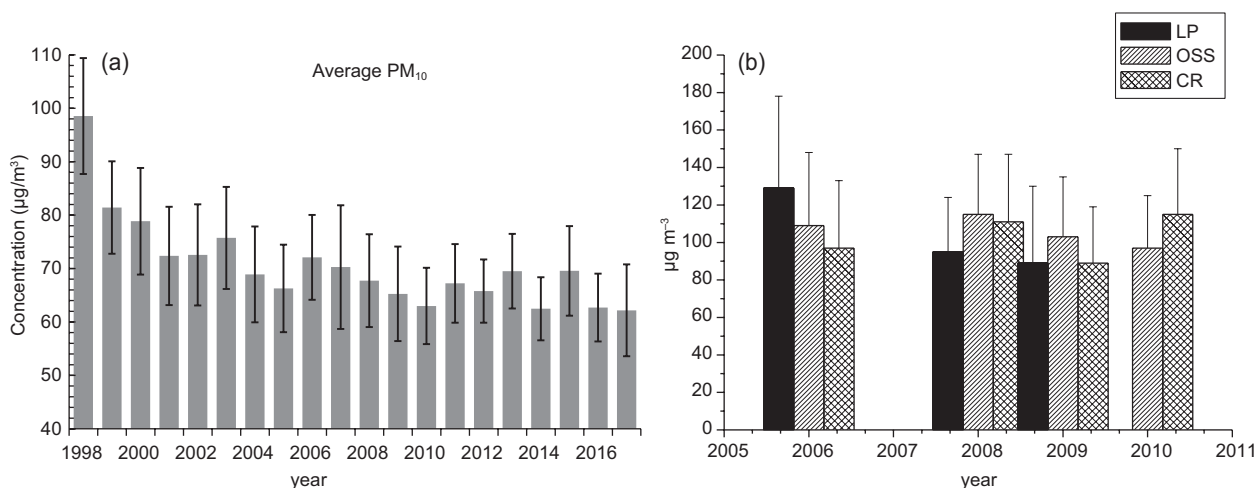


Fig. 4. (a) Evolution of the yearly PM₁₀ average at the nine stations of the SINCA network in Santiago. (b) Evolution of average PM₁₀ measured during May 2006, 2008, 2009, and 2010 for Cerrillos (CR), La Pintana (LP) and Av. Ossa (OSS).

Fisher Scientific, Waltham, MA). From 2007 ($70.3 \mu\text{g m}^{-3}$) to 2010 ($63.0 \mu\text{g m}^{-3}$) there was a slight decrease in PM_{10} , however this trend did not continue and PM_{10} has leveled since 2010, as shown in Figure 4a. In the three monitoring stations (Cerrillos [CR], La Pintana [LP] and Av. Ossa [OSS]) there has also been a slight decrease in concentrations from 2006 to 2010, as shown in Figure 4b. During these years important measures have been implemented to abate air pollution from public transport, which included the restructuring of the public transport system, the renewal of the vehicle fleet, and the shortening of circulation routes. These measures were somewhat unpopular, leading to a slight increase of private vehicles (around 6% more in 2009 with respect to 2003 [INE 2011]). This result agrees partly with the report of the Sociedad Chilena de Políticas Públicas (Chilean Society of Public Policies), which notes that “the annual evolution of PM_{10} between 1997 and 2010 shows a decreasing but moderate trend, indicating that air quality has improved steadily over the past 15 years”. However, as shown in Figure 4a, since 2010 the trend has decreased, and PM_{10} levels have been fairly constant (Moreno et al., 2010; Figuero et al., 2017).

3.2 Major and trace elements in PM_{10} : mean concentrations and spatio-temporal variation

The application of the analytical procedures described above has resulted in the determination and quantification of major and trace elements. Among the most abundant, Fe, Al, K, Ca, and Mg prevailed, with concentrations in the range $600\text{--}8000 \text{ ng m}^{-3}$ (Table I). These are mostly associated with mineral dust both from natural and anthropogenic origins (but essentially from the latter) (Paris et al., 2010). Sources such as biomass burning in the case of K (Cheng et al., 2013) or industrial emissions and brake abrasion in the case of Fe (Ulrich et al., 2012) may also contribute. Other abundant PM_{10} components are S, mostly as sulfate salts (ammonium sulfate predominantly but also forming Ca and Mg sulfate compounds), at concentrations between $2000\text{--}4000 \text{ ng m}^{-3}$ (Table I) and phosphorus, usually with concentrations in the range of $280\text{--}1600 \text{ ng m}^{-3}$ (Wang et al., 2015). As expected, lower abundances (usually under 100 ng m^{-3}) have been found for other elements, including heavy metals (Pb, Ti, Mn, Cu, Zn, Ba, Zr, Cr, As, Sn,

Sb, Ni, V, Li, Co, Cd, La, Rb). Details on year- and site-specific average concentrations can be found in Table I.

Regarding water-soluble PM_{10} components (sulfate, nitrate, ammonium and chloride), Table II shows that the most abundant is ammonium, with concentrations between 0.42 and $0.21 \mu\text{Eq m}^{-3}$. Nitrate is the second compound in abundance, in the range of 0.38 to $0.19 \mu\text{Eq m}^{-3}$; followed by chloride, showing concentrations in the range of 0.12 to $0.08 \mu\text{Eq m}^{-3}$, and finally sulfate in the range of 0.10 to $0.05 \mu\text{Eq m}^{-3}$. These results are not surprising: urban pollution episodes are generally characterized by huge quantities of secondary inorganic and organic components. Amongst the inorganic ones, ammonium nitrate is the foremost abundant component, mostly derived from the transformation in the atmosphere of combustion exhausts (NO_x to nitrates) from industry and vehicles.

Overall, the amount of major anions and cations, namely sulfate, nitrate, ammonium and chloride (Table II), accounts for 40–50% of PM_{10} mass concentrations. A good linear correlation between nitrates and ammonium was observed ($A = 0.0395 \pm 0.057$; $B = 1.015 \pm 0.199$; $R = 0.68707 \pm 0.10006$; $N = 31$; $p < 0.0001$). If we take into account the soluble metallic cations, the cation to anion ratio is close to 1, ensuring the neutrality of the particles.

In order to deal with the main characteristics of each monitoring site in terms of PM_{10} composition, in Figure 5 we have summarized some of the typical tracers for different sources. As seen in Figure 5a, a typical PM component such as sulfur was found to be homogeneously distributed among the sites during the specific annual campaigns, showing almost double concentrations in 2006 with respect to the period 2008–2010. Figure 5b displays year- and site-dependent concentrations of calcium, a typical mineral dust element. In this case, the highest concentrations were recurrently observed in Av. Ossa during the first two years, probably due to local influences such as construction activities. Thereafter, Ca concentrations remained comparable in all sites. Figure 5c shows year- and site-dependent concentrations of vanadium and nickel, as traces of fuel-oil combustion. This particular PM source seems to be of a regional nature, with a slight enrichment towards La Pintana in the first two years, probably due to the location of some industries in which heavy oil combustion

Table I. Monthly average concentrations (ng m^{-3}) of elements associated with atmospheric PM_{10} of Cerrillos (CR), Av. Ossa (OSS) and La Pintana (LP) in the total analysis of dissolved particulate matter.

Element	Cerrillos				Av. Ossa				La Pintana			
	2006	2008	2009	2010	2006	2008	2009	2010	2006	2008	2009	2010
S	3802 ± 160	2422 ± 625	2333 ± 258	2840 ± 24	3514 ± 158	2029 ± 432	2411 ± 421	2024 ± 61	4104 ± 212	2115 ± 452	2415 ± 290	
Al	1869 ± 540	2539 ± 144	1965 ± 135	3885 ± 103	2658 ± 647	3347 ± 559	2168 ± 570	1973 ± 148	4319 ± 1472	3019 ± 779	2275 ± 997	
Ca	2533 ± 860	3122 ± 542	2461 ± 30	4457 ± 276	6035 ± 1200	7948 ± 928	2732 ± 565	3560 ± 276	4421 ± 1030	4017 ± 109	2787 ± 444	
Fe	2313 ± 600	2339 ± 490	2884 ± 107	4456 ± 203	3754 ± 630	4784 ± 600	3932 ± 710	4505 ± 247	4051 ± 1400	2012 ± 370	2409 ± 498	
K	1266 ± 405	1420 ± 465	1291 ± 30	1671 ± 25	1820 ± 334	1740 ± 271	1311 ± 302	1089 ± 53	1795 ± 570	1223 ± 295	3517 ± 1100	
Mg	68 ± 257	714 ± 95	632 ± 47	860 ± 14	986 ± 65	1435 ± 247	747 ± 153	630 ± 28	1055 ± 300	846 ± 86	664 ± 177	
P	358 ± 200	938 ± 311	1398 ± 143	277 ± 5	420 ± 244	808 ± 68	1607 ± 176	1597 ± 35	505 ± 70	953 ± 167	1521 ± 153	
Zn	291 ± 105	432 ± 160	219 ± 24	442 ± 13	678 ± 40	302 ± 51	181 ± 6	271 ± 64	179 ± 130	895 ± 79	123 ± 55	
Cu	141 ± 49	156 ± 26	104 ± 8	173 ± 5	199 ± 43	169 ± 42	133 ± 12	153 ± 6	117 ± 17	79 ± 10	66 ± 3	
Ti	134 ± 41	167 ± 36	178 ± 5	294 ± 5	19 ± 52	190 ± 73	208 ± 30	205 ± 0.2	289 ± 90	186 ± 40	199 ± 40	
Pb	125 ± 27	84 ± 18	70 ± 3	123 ± 0.9	87 ± 43	50 ± 17	46 ± 1	48 ± 2	67 ± 16	43 ± 3	42 ± 4	
Mn	86 ± 27	94 ± 12	90 ± 9	121 ± 0.8	108 ± 25	123 ± 32	100 ± 10	88 ± 1	113 ± 39	80 ± 10	78 ± 9	
Ba	83 ± 20	100 ± 13	59 ± 7	109 ± 3	178 ± 28	155 ± 29	88 ± 4	132 ± 6	63 ± 17	62 ± 4	55 ± 0.3	
As	28 ± 10	17 ± 4	19 ± 4	27 ± 0.4	24 ± 11	12 ± 5	17 ± 0.9	18 ± 0.5	43 ± 22	22 ± 6	27 ± 1	
Sb	19 ± 6	16 ± 4	11 ± 0.8	23 ± 0.5	20 ± 8	16 ± 4	19 ± 0.2	17 ± 0.2	15 ± 3	10 ± 1	9 ± 2	
Zr	19 ± 10	10 ± 3	10 ± 3	44 ± 2	23 ± 13	9 ± 3	9 ± 4	11 ± 2	25 ± 8	11 ± 1	3 ± 0.4	
Sn	25 ± 5	21 ± 5	17 ± 2	29 ± 0.04	20 ± 7	18 ± 5	16 ± 4	17 ± 0.7	9 ± 2	8 ± 0.6	5 ± 0.8	
Cr	14 ± 6	13 ± 2	13 ± 1	17 ± 5	18 ± 4	19 ± 4	14 ± 0.6	17 ± 0.7	9 ± 3	9 ± 0.4	10 ± 0.2	
V	8 ± 3	14 ± 3	7 ± 0.2	9 ± 0.7	9 ± 3	11 ± 2	6 ± 1	6 ± 0.5	14 ± 5	19 ± 6	7 ± 0.1	
Ni	8 ± 1	6 ± 2	14 ± 0.9	6 ± 0.4	8 ± 2	11 ± 1	1 ± 1	12 ± 0.02	8 ± 3	8 ± 4	13 ± 0.8	
Se	5 ± 0.8	6 ± 0.6	4 ± 0.4	8 ± 0.4	3 ± 0.1	3 ± 0.1	3 ± 0.05	4 ± 0.03	4 ± 2	3 ± 0.4	3 ± 0.9	
Rb	3 ± 0.9	4 ± 0.9	2 ± 0.1	5 ± 0.05	3 ± 0.9	3 ± 0.7	2 ± 0.7	2 ± 0.003	5 ± 2	3 ± 1	3 ± 0.2	
Co	1 ± 0.5	1 ± 0.2	1 ± 0.05	2 ± 0.02	1 ± 0.3	1 ± 0.3	1 ± 0.5	1 ± 0.02	2 ± 0.4	2 ± 0.6	1 ± 0.4	
Cd	2 ± 0.4	1 ± 0.2	1 ± 0.4	2 ± 0.03	1 ± 0.1	1 ± 0.2	1 ± 0.09	1 ± 0.04	1 ± 0.2	1 ± 0.2	1 ± 0.3	
Li	1 ± 0.6	1 ± 0.3	1 ± 0.03	2 ± 0.02	1 ± 0.5	1 ± 0.2	1 ± 0.2	1 ± 0.08	1 ± 0.7	1 ± 0.5	1 ± 0.3	

Table II. Concentrations of sulfate, nitrate, chloride and metallic cations; ammonium ($\mu\text{Eq m}^{-3}$), and cation to anion ratios in the water soluble fraction of PM_{10} .

Year	Area of Santiago	Sulfate	Chloride	Nitrate	Ammonium	Total anions	Total metallic cations	Total cations	Cation to anion ratio
2006	Cerrillos	0.10 ± 0.047	0.12 ± 0.031	0.28 ± 0.099	0.42 ± 0.197	0.5	0.145 ± 0.042	0.564	1.13
2008		0.06 ± 0.016	0.09 ± 0.020	0.28 ± 0.12	0.29 ± 0.116	0.43	0.184 ± 0.049	0.474	1.10
2009		0.06 ± 0.008	0.13 ± 0.014	0.23 ± 0.035	0.24 ± 0.038	0.42	0.156 ± 0.013	0.396	0.94
2010		0.07 ± 0.001	0.12 ± 0.014	0.29 ± 0.015	0.27 ± 0.0001	0.48	0.229 ± 0.003	0.499	1.04
2006	Av. Ossa	0.09 ± 0.048	0.09 ± 0.028	0.32 ± 0.124	0.39 ± 0.203	0.5	0.172 ± 0.054	0.562	1.12
2008		0.05 ± 0.010	0.08 ± 0.012	0.38 ± 0.029	0.36 ± 0.116	0.51	0.160 ± 0.033	0.520	1.02
2009		0.06 ± 0.012	0.09 ± 0.015	0.32 ± 0.005	0.29 ± 0.018	0.47	0.169 ± 0.018	0.459	0.98
2010		0.05 ± 0.032	0.09 ± 0.013	0.27 ± 0.088	0.21 ± 0.001	0.41	0.165 ± 0.001	0.375	0.91
2006	La Pintana	0.10 ± 0.050	0.09 ± 0.010	0.22 ± 0.060	0.35 ± 0.179	0.41	0.213 ± 0.061	0.563	1.37
2008		0.05 ± 0.024	0.10 ± 0.008	0.20 ± 0.078	0.24 ± 0.054	0.35	0.178 ± 0.052	0.418	1.19
2009		0.05 ± 0.005	0.09 ± 0.006	0.19 ± 0.049	0.24 ± 0.026	0.33	0.173 ± 0.034	0.413	1.25

occurs. A common brake-wearing trace component, antimony, is presented in Figure 5d. Concentrations of this element were clearly shifted towards urban sites (CR and OSS) rather than the suburban area (LP), with rather stable values from 2006 to 2010, maybe slightly increasing in parallel to the vehicle fleet increase. In Figure 5e we plotted Zn concentrations as an abundant trace element with multiple origins. Industrial and non-exhaust traffic emissions contribute to ambient air concentrations of this trace metal. As seen in the graph, the highest concentrations occurred alternatively in La Pintana in 2008, probably connected to industrial pollution plumes, in Av. Ossa in 2006, or in Cerrillos in 2008 and 2010, mostly related to vehicular emissions. To conclude this section, Figure 5f displays arsenic concentrations, selected as a tracer of industrial emissions. In all sites, but especially in the suburban-industrial site of La Pintana, arsenic concentrations were high ($11\text{--}43 \text{ ng m}^{-3}$), and considering the European standard for this element (6 ng m^{-3}), control measures should be urgently adopted.

3.3 Cluster analysis: main associations of major and trace PM components

Our database was not large enough to perform a classical source apportionment analysis. However, alternative statistical treatments such as a cluster analysis may allow us to know the main associations

among the elements determined in this study. To this end, we merged all data in a single database and a dendrogram was obtained. The results are shown in Figure 6, where two big groups of elements are separated. The first one, subdivided into two groups, contains Ti and Sn on one hand, and Cr, Pb, Co and Ni on the other. These associations are related to anthropogenic activities, probably fuel-oil combustion emissions (Ni, Co, Pb, Cr, as seen by Pey et al., 2013) and certain industrial processes (Ti and Sn). The second group of elements is, at the same time, subdivided into two other associations, being the first one characterized by V, Mn, Cu, Se and Sb, which are rather coincident with non-exhaust vehicular emissions from brake and tire abrasion (Pacyna, 1998; Amato et al., 2009; Pey et al., 2010), and the second one by the presence of Zn, As, Ba, and Cd, which are mostly linked to industrial processes (Querol et al., 2007); however Ba, and to a lesser extent Zn, are also related to vehicular non-exhaust emissions (Amato et al., 2011). Overall, the dendrogram analysis revealed different anthropogenic emission sources in the Santiago metropolitan area, separating two main groups of particles: one from industrial-related processes and another one more related to vehicular emissions, but somewhat mixed with industrial influences. According to the 2014 emissions inventory, 20% of the PM_{10} in Santiago is due to industrial sources, 30% to wood burning,

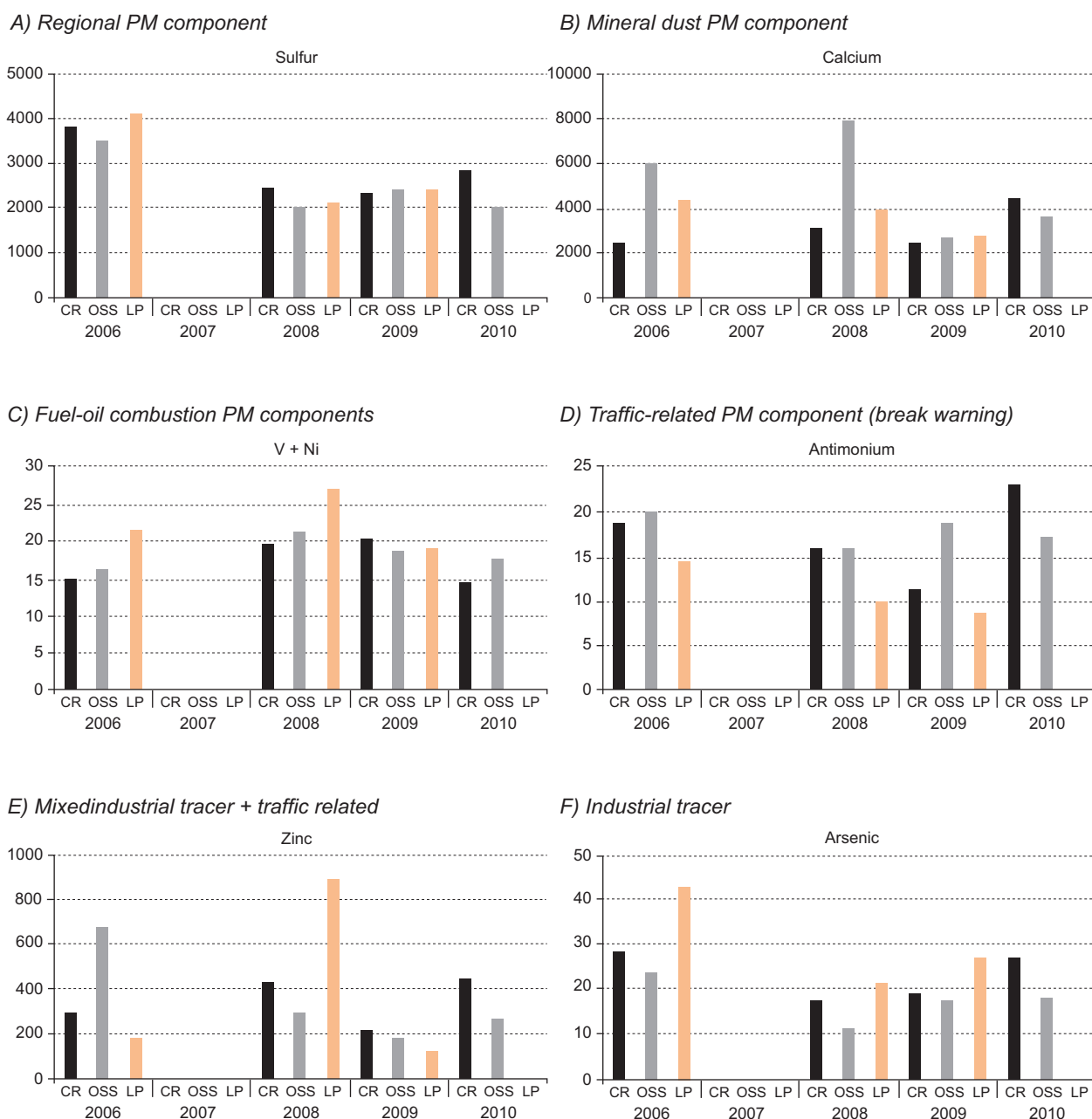


Fig. 5. Temporal and spatial variation of PM₁₀ components associated to important emission sources present in Santiago during May in the years 2006-2010: (a) regional PM component, (b) mineral dust PM component, (c) fuel-oil combustion PM components, (d) traffic-related PM component (brake wearing), (e) mixed industrial tracer + traffic related component, and (f) industrial tracer.

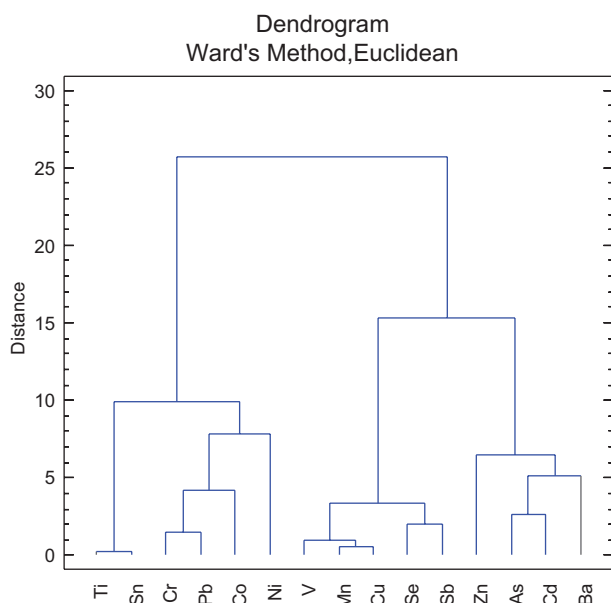


Fig. 6. Results of cluster analysis from the integrated metallic elements database (La Pintana + Av. Ossa + Cerrillos).

3% to agricultural burning, and 44% to mobile sources (MMA, 2014). An extended study covering all seasons and incorporating other chemical determinations would be desirable to fully investigate the real sources and their contributions, which is absolutely necessary to develop and implement adequate mitigation strategies.

3.4 Soluble elements

PM₁₀ samples were subjected to water extraction to determine the degree of solubility of major and trace metals in order to determine its lability in Santiago's atmosphere.

Table III shows monthly average concentrations for each site and year of major and trace elements. By combining the information presented in Tables I and III we were able to prepare Figure 5, in which the soluble/insoluble partitioning mass concentration of each of the elements determined in this study is shown. After sulfur, clearly shifted towards the soluble fraction (> 90% of the total mass was found as soluble compounds), specific trace elements such as Ba, Sb, Cd, As, and Zn displayed a significant water-soluble content, with 50% to 73%. Keeping in mind that some of these elements are considered toxic and/or carcinogenic (As, Cd and Zn) to human

beings (Hughes, 2002; Lafuente, 2013), our results are even more relevant. It has been found that the urban atmosphere of Santiago is highly enriched in trace elements, some of them highly soluble and therefore with a higher toxicological potential with respect to other less soluble or practically insoluble. On the other hand, Al, Mg, Zr, Ti, Sn, Cr, and Pb showed a preference for the non-soluble fraction, with relative abundances of 75-99%.

Comparing our results from the Cerrillos station with those obtained by Richter et al. (2007), the solubility percentages are quite comparable (As, 70% soluble in our study against 60% in the aforementioned article; Pb, about 15% in our work against 40% in the previous decade).

The results on solubility, besides providing information on the lability in the respiratory system, may give some indications about possible chemical states of the species, i.e., the kinds of salts present in PM. For example, water-soluble Zn is usually found in the form of zinc sulfate, while in the residual fraction it is commonly found as zinc oxide, frequently joined with Fe and Mn, carbonates, or aluminates (Dos Santos et al., 2009). Likewise, water-soluble Cu is usually encountered as copper sulfate, while the residual fraction is mostly linked to sulfides or silicates.

Overall, the atmosphere of Santiago is rich in soluble and insoluble major and trace metals. Most of them are associated with industrial and road-traffic emissions, especially non-exhaust emissions from brake and tire wearing and pavement abrasion. Santiago's inhabitants are exposed to high heavy metal concentrations, at least during the cold season.

The amount of soluble major and trace elements together with the amount of typical anions and cations (Tables II and III) is near 50% of the PM₁₀ mass.

The results from the cluster analysis in Figure 6 give an approximate composition of the sources responsible for the measured PM. Figure 8 gives the concentration of soluble elements according to these sources separated according to the sites. It is found in Figure 8a that La Pintana has a higher concentration of crustal elements (Ca, K, Mg and Al), which is due to the fact that this site is located towards the south of Santiago (see Fig. 1), far from downtown, thus it receives more influence from rural environments. Exhaust and fuel oil emissions do not show a preferred trend and all stations have similar

Table III. Monthly average of water soluble elements concentrations (ng m^{-3}) in atmospheric PM_{10} of Santiago.

	Cerrillos				Av. Ossa				La Pintana			
	2006	2008	2009	2010	2006	2008	2009	2010	2006	2008	2009	2009
S	3760 \pm 1698	2240 \pm 522	2299 \pm 278	2536 \pm 8	3273 \pm 1542	1847 \pm 404	2315 \pm 475	1823 \pm 15	3048 \pm 1469	2036 \pm 464	2279 \pm 243	
Ca	1338 \pm 397	1923 \pm 470	1773 \pm 146	2747 \pm 33	1887 \pm 662	1851 \pm 424	1989 \pm 283	1901 \pm 7	2608 \pm 682	2225 \pm 852	1898 \pm 467	
K	763 \pm 236	785 \pm 239	521 \pm 56	767 \pm 11	687 \pm 173	567 \pm 127	507 \pm 0	425 \pm 3	889 \pm 259	654 \pm 53	684 \pm 0	
P	233 \pm 72	354 \pm 24	991 \pm 126	179 \pm 5	241 \pm 102	326 \pm 22	1108 \pm 103	1088 \pm 22	272 \pm 38	351 \pm 22	1019 \pm 104	
Zn	312 \pm 101	338 \pm 123	138 \pm 3	344 \pm 1	190 \pm 93	134 \pm 26	119 \pm 3	140 \pm 4	136 \pm 47	103 \pm 24	108 \pm 50	
Fe	151 \pm 41	160 \pm 52	170 \pm 11	134 \pm 2	163 \pm 70	153 \pm 24	243 \pm 26	256 \pm 5	102 \pm 32	97 \pm 35	146 \pm 34	
Mg	110 \pm 22	147 \pm 27	163 \pm 3	218 \pm 2	132 \pm 40	130 \pm 25	171 \pm 24	168 \pm 1	178 \pm 51	135 \pm 35	149 \pm 35	
Al	97 \pm 15	107 \pm 33	99 \pm 11	92 \pm 1	102 \pm 7	103 \pm 0	114 \pm 9	107.5 \pm 0.2	107 \pm 43	96 \pm 0	106 \pm 26	
Ba	61 \pm 16	54 \pm 10	35 \pm 11	73.1 \pm 0.7	102 \pm 31	63 \pm 13	58 \pm 2	61.2 \pm 0.2	31 \pm 7	22 \pm 6	20 \pm 3	
Cu	59 \pm 12	71 \pm 13	33.0 \pm 0.1	73 \pm 1	62 \pm 16	56 \pm 12	44 \pm 2	52.2 \pm 0.2	41 \pm 9	29 \pm 10	18.7 \pm 0.5	
Mn	43 \pm 13	50 \pm 8	39 \pm 5	57.4 \pm 0.6	46 \pm 16	47 \pm 11	43 \pm 3	38.0 \pm 0.3	51 \pm 17	36 \pm 7.0	35 \pm 6	
As	21 \pm 8	12 \pm 3	14 \pm 3	18.75 \pm 0.08	17 \pm 8	9 \pm 4	12.5 \pm 0.4	13.0 \pm 0.2	33.0 \pm 17.0	1.0 \pm 5	19.8 \pm 0.9	
Pb	28 \pm 5	13 \pm 6	7 \pm 1	15.4 \pm 0.4	12 \pm 6	5 \pm 1	2.82 \pm 0.04	2.53 \pm 0.06	7.0 \pm 3.0	6.0 \pm 1	4.1 \pm 0.4	
Sb	10 \pm 3	8 \pm 2	6.6 \pm 0.5	10.6 \pm 0.1	9.0 \pm 3.0	7.0 \pm 2.0	8.7 \pm 0.7	8.32 \pm 0.06	0.07 \pm 1.0	5.6 \pm 0.7	5.0 \pm 1.0	
V	3.7 \pm 1.0	7.0 \pm 2.0	2.6 \pm 0.4	2.055 \pm 0.007	3.0 \pm 0.1	4.9 \pm 1.0	1.9 \pm 0.5	1.41 \pm 0.01	4.8 \pm 1.7	11.0 \pm 3	2.0 \pm 0.4	
Cr	1.7 \pm 0.4	1.7 \pm 0.5	2.4 \pm 0.7	1.61 \pm 0.06	2.0 \pm 1	2.0 \pm 0.2	2.2 \pm 0.4	2.3 \pm 0.5	0.4 \pm 0.4	0.6 \pm 0.3	1.44 \pm 0.03	
Ni	1.5 \pm 0.4	2.1 \pm 0.6	1.0 \pm 0.2	1.17 \pm 0.01	1.5 \pm 0.5	1.5 \pm 0.2	0.8 \pm 0.1	0.82 \pm 0.01	2.0 \pm 1	2.4 \pm 0.8	0.6 \pm 0.1	
Se	2.0 \pm 0.3	2.2 \pm 0.3	1.9 \pm 0.2	3.33 \pm 0.04	1.4 \pm 0.3	1.6 \pm 0.3	1.5 \pm 0.3	1.8 \pm 0.1	1.5 \pm 0.5	1.2 \pm 0.1	1.2 \pm 0.3	
Rb	1.4 \pm 0.4	1.7 \pm 0.4	0.9 \pm 0.1	1.50 \pm 0	1.1 \pm 0.4	1.2 \pm 0.2	0.84 \pm 0.06	0.79 \pm 0.007	1.3 \pm 0.4	1.1 \pm 0.4	1.0 \pm 0.2	
Cd	1.2 \pm 0.2	1.0 \pm 0.2	0.7 \pm 0.2	0.995 \pm 0.007	0.8 \pm 0.4	0.44 \pm 0.09	0.48 \pm 0.03	0.41 \pm 0.03	0.9 \pm 0.2	0.66 \pm 0.08	0.7 \pm 0.2	
Ti	0.8 \pm 0.3	0.8 \pm 0.3	1.3 \pm 0.1	0.76 \pm 0.06	0.8 \pm 0.3	0.9 \pm 0.1	2.0 \pm 0.1	1.64 \pm 0.01	1.2 \pm 0.4	1.0 \pm 0.3	1.0 \pm 0.2	
Sn	0.9 \pm 0.3	0.6 \pm 0.3	1.17 \pm 0.03	0.39 \pm 0.04	0.6 \pm 0.3	0.7 \pm 0.2	0.90 \pm 0.07	0.91 \pm 0.04	0.31 \pm 0.09	0.5 \pm 0.2	0.36 \pm 0.05	
Co	0.4 \pm 0.2	0.33 \pm 0.08	0.31 \pm 0.01	0.56 \pm 0.01	0.3 \pm 0.1	0.31 \pm 0.06	0.31 \pm 0.04	0.33 \pm 0.01	0.42 \pm 0.08	0.31 \pm 0.06	0.26 \pm 0.04	

1 H 1.008																	2 He 4.003														
3 Li 6.941	4 Be 9.012													5 B 10.81	6 C 12.01	7 N 14.01	8 O 16.00	9 F 19.00	10 Ne 20.18												
11 Na 22.99	12 Mg 24.31													13 Al 26.98	14 Si 28.09	15 P 30.97	16 S 32.07	17 Cl 35.45	18 Ar 39.95												
19 K 39.10	20 Ca 40.08	21 Sc 44.96	22 Ti 47.88	23 V 50.94	24 Cr 52.00	25 Mn 54.94	26 Fe 55.85	27 Co 58.93	28 Ni 58.69	29 Cu 63.55	30 Zn 65.39	31 Ga 69.72	32 Ge 72.61	33 As 74.92	34 Se 78.96	35 Br 79.90	36 Kr 83.80														
37 Rb 85.47	38 Sr 87.62	39 Y 88.91	40 Zr 91.22	41 Nb 92.91	42 Mo 95.94	43 Tc (98)	44 Ru 101.1	45 Rh 102.9	46 Pd 106.4	47 Ag 107.9	48 Cd 112.4	49 In 114.8	50 Sn 118.7	51 Sb 121.8	52 Te 127.6	53 I 126.9	54 Xe 131.3														
55 Cs 132.9	56 Ba 137.3	57 La 138.9	58 Ce 140.1	59 Pr 140.9	60 Nd 144.2	61 Pm (145)	62 Sm 150.4	63 Eu 151.9	64 Gd 157.3	65 Tb 158.9	66 Dy 162.5	67 Ho 164.9	68 Er 167.3	69 Tm 168.9	70 Yb 173.0	71 Lu 174.9	72 Hf 178.5	73 Ta 180.9	74 W 183.8	75 Re 186.2	76 Os 190.2	77 Ir 192.2	78 Pt 195.1	79 Au 197.0	80 Hg 200.6	81 Tl 204.4	82 Pb 207.2	83 Bi 209.0	84 Po (209)	85 At (210)	86 Rn (222)
87 Fr (223)	88 Ra (226)	89 Ac (227)	104 Rf (261)	105 Db (262)	106 Sg (263)	107 Bh (264)	108 Hs (265)	109 Mt (266)	110 Ds (271)	111 Uuu (272)	112 Uub (285)	113 Uut (284)	114 Uuq (289)	115 Uup (288)																	

Insoluble

Soluble

Fig. 7. Partitioning of soluble/insoluble concentrations of the determined elements.

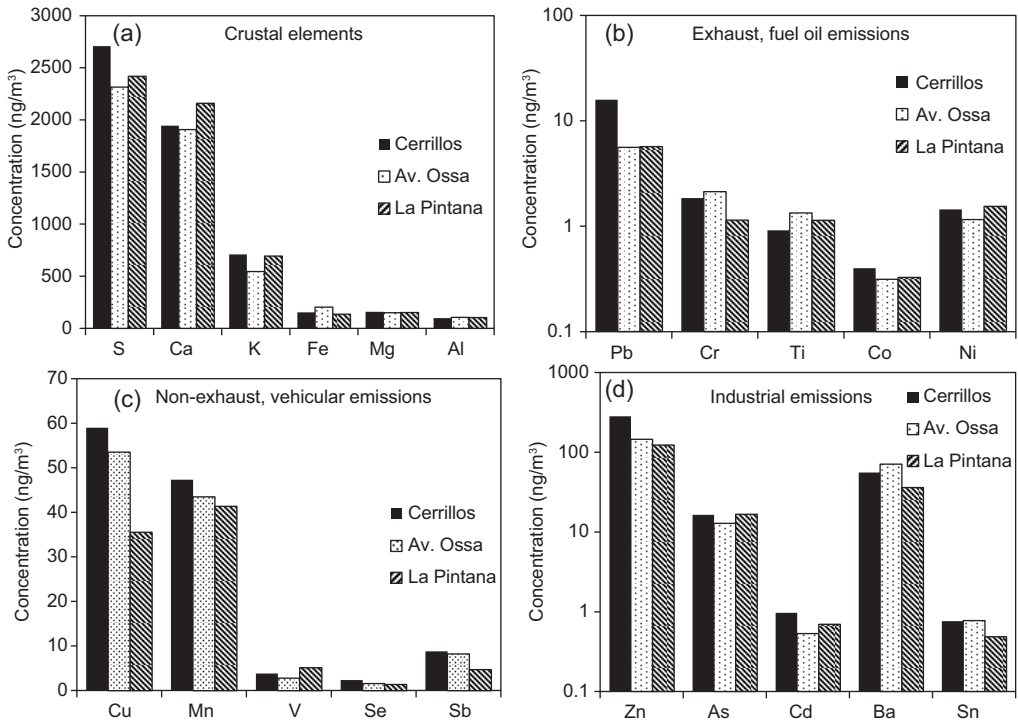


Fig. 8. Soluble components of PM₁₀ associated with different sources in the three sites. (a) Crustal elements, (b) exhaust and fuel oil emissions, (c) non-exhaust emissions, (d) industrial emissions.

concentrations from this source. This figure indicates that exhaust emissions are well distributed over the city, which can be explained because mobile sources are spread out. Non-exhaust vehicular emissions seem to be higher in Cerrillos although Cu and Mn could also be generated in industrial processes. Industrial emissions are shown in Figure 8d, which also seem to be higher in Cerrillos. This can be explained

because in this sector there are several large industries and many small auto repair and paint shops.

4. Conclusions

The inhabitants of Santiago are exposed to high PM₁₀ pollution, which ranged from 71 µg m⁻³ (Cerrillos) to 128 µg m⁻³ (La Pintana), associated mainly with

the vehicular traffic present in the city. In Santiago the central and peripheral urban zones are subjected to the same levels of PM₁₀ and its heavy metal load.

The levels of As and Cd, which are reported to be toxic to human health, had high solubility in PM₁₀, showing that breathable PM₁₀ is very noxious to the inhabitants of Santiago. The As levels exceed the environmental quality standard. One of the sites (Cerrillos) was clearly influenced from industrial sources because it had higher levels of Zn, As and Cd.

The soluble or labile fraction, which is the really important one regarding breathable pollution, had high levels of the species in the form of nitrates, chlorides, and sulfates of Cd, Zn, Ba, As, and Sb. The soluble fraction of Ba, Sb, Cd, As, and Zn, which are generated mainly from industrial sources, ranged from 50% to 73%. The chemical species associated with Zn and with NO₃⁻ were the most soluble.

These results call for a long-term study of PM composition and sources in Santiago which will serve a double purpose: (1) surveillance of the government air quality mitigation plan and its effectiveness, and (2) creation of a database for further epidemiological studies including sources and/or chemical species and their contributions.

Acknowledgments

Funding from different projects supported this study: Fondecyt 1170456, 1151117 and CONICYT by the doctoral scholarship of Karen Sánchez grant 24129966. Most of the analytical costs were covered by different research projects from the Spanish Ministry of Science and Innovation (CARIATI- CGL2008-06294/CLI, CGL2007-62505/CLI, GRACCIE-CSD2007-00067). J. Pey currently benefits from a Ramón y Cajal Research Grant (RYC-2013-14159) from the Spanish Ministry of Economy and Competitiveness.

References

- Amato F., Pandolfi M., Escrig A., Querol X., Alastuey A., Pey J., Pérez N. and Hopke P., 2009. Quantifying road dust resuspension in urban environment by Multilinear Engine: a comparison with PMF2. *Atmos. Environ.* 43, 2770-2780.
DOI: 10.1016/j.atmosenv.2009.02.039
- Amato F., Viana M., Richard A., Furger M., Prévôt A.S.H., Nava S., Lucarelli F., Bukowiecki N., Alastuey A., Reche C., Moreno T., Pandolfi M., Pey J. and Querol X., 2011. Size and time-resolved roadside enrichment of atmospheric particulate pollutants. *Atmos. Chem. Phys.* 6, 2917-2931.
DOI: 10.5194/acp-11-2917-2011
- Artaxo P., Oyola P. and Martínez R. 1999. Aerosol composition and source apportionment in Santiago de Chile. *Nucl. Instrum. Methods* 150, 409-416.
DOI: 10.1016/S0168-583X(98)01078-7
- Bell M.L., Ebisu K., Peng R., Samet J.M. and Dominici F., 2009. Hospital admissions and chemical composition of fine particle air pollution. *Am. J. Resp. Crit. Care Med.* 179, 1115-1120.
DOI: 10.1164/rccm.200808-1240OC
- Cakmak S., Dales R. and Blanco C., 2007. Air pollution and mortality in Chile: Susceptibility among the elderly. *Environ. Health Persp.* 115, 524-527.
DOI: 10.1289/ehp.9567
- Cakmak S., Dales R., Blanco C., Fernández M., Rubio M.A. and Oyola P., 2009. Components of particulate air pollution and emergency department visits in Chile. *Arch. Environ. Occup. H.* 64, N3, 148-155.
DOI: 10.1080/19338240903240228
- Cakmak S., Dales R., Rubio M.A. and Blanco C., 2011. The risk of dying on days of higher air pollution among the socially disadvantaged elderly. *Environ. Res.* 111, 388-93.
DOI: 10.1016/j.envres.2011.01.003
- Cheng Y., Engling G., He K.-B., Duan F.-K., Ma Y.-L., Du Z.-Y., Liu J.-M., Zheng M. and Weber R.J., 2013. Biomass burning contribution to Beijing aerosol. *Atmos. Chem. Phys.* 13, 7765-7781.
DOI: /10.5194/acp-13-7765-2013
- Didyk B., Simoneit B.R., Pezoa L., Riveros M. and Flores A., 2000. Urban aerosol particles of Santiago, Chile. *Atmos. Environ.* 34, 1167-1179.
DOI: 10.1016/S1352-2310(99)00403-3
- Dos Santos M., Gómez D., Dawidowski L., Goutier E. and Smichowski P., 2009. Determination of water-soluble and insoluble compounds in size classified airborne particulate matter. *Microchem. J.* 91, 133-139.
DOI: 0.1016/j.microc.2008.09.001
- Figuro E., Gómez-Lobos A., Jorquera P. and Labrín F., 2017. Los impactos del Transantiago en la contaminación atmosférica de Santiago. Una estimación econométrica. Sociedad Chilena de Políticas Públicas.

- Available at: https://www.sociedadpoliticaspublicas.cl/archivos/BLOQUEM/Vivienda_Transporte_Desafios_Sectoriales/Los_impactos_del_Transantiago_en_la_contaminacion_de_Santiago.pdf (last accessed on August 2018).
- Franklin M., Koutrakis P. and Schwartz J., 2008. The role of particle composition on the association between $PM_{2.5}$ and mortality. *Epidemiology* 19, 680-689. DOI: 10.1097/EDE.0b013e3181812bb7
- Gil L., King L. and Adonis M., 2000. Trends of polycyclic aromatic hydrocarbon levels and mutagenic in Santiago's inhalable airborne particles in the period 1992-1996. *Inhal. Toxicol.* 12, 1185-1204. DOI: 10.1080/08958370050198539
- Gramsch E., Cereceda-Balic F., Oyola P. and von Baer D., 2006. Examination of pollution trends in Santiago de Chile with cluster analysis of PM_{10} and ozone data. *Atmos. Environ.* 40, 5464-5475. DOI: 10.1016/j.atmosenv.2006.03.062
- Gramsch E., Gidhagen L., Wahlin P., Oyola P., Moreno F., 2009. Predominance of soot-mode ultrafine particles in Santiago de Chile: Possible sources. *Atmos. Environ.* 43, 2260-2267. DOI: 10.1016/j.atmosenv.2009.01.047
- Gramsch E., Le Nir, G., Araya M., Rubio M.A., Moreno F. and Oyola P., 2013. Influence of large changes in public transportation (Transantiago) on the black carbon pollution near streets. *Atmos. Environ.* 65, 153-163. DOI: 10.1016/j.atmosenv.2012.10.006
- Hughes M., 2002. Arsenic toxicity and potential mechanisms of action. *Toxicol. Lett.* 133, 1-16. DOI: 10.1016/S0378-4274(02)00084-X
- INE, 2011, Annual report of motorized vehicles in Chile. Available at: historico.ine.cl/canales/chile_estadistico/estadisticas_economicas/transporte_y_comunicaciones/parquevehiculos.php (last accessed on May 2018).
- INE, 2017. Censo 2017. Instituto Nacional de Estadística, Chile. Available at: <http://www.censo2017.cl/>.
- Jorquera H., 2002. Air quality at Santiago, Chile a box modeling approach II. $PM_{2.5}$, coarse and PM_{10} particulate matter fractions. *Atmos. Environ.* 36, 331-344. DOI: 10.1016/S1352-2310(01)00418-6
- Jorquera H. and Barraza F., 2012. Source apportionment of ambient $PM_{2.5}$ in Santiago, Chile: 1999 and 2004 results. *Sci. Total Environ.* 435-436, 418-429. DOI: 10.1016/j.scitotenv.2012.07.049
- Koutrakis P., Sax S., Sarnat J., Coull B., Demokritou P., Oyola P., García J. and Gramsch E., 2005. Analysis of PM_{10} , $PM_{2.5}$ and $PM_{2.5-10}$ concentrations in Santiago, Chile from 1989 to 2001. *J. Air Waste Man.* 55, 342-351. DOI: 10.1080/10473289.2005.10464627
- Lafuente A., 2013. The hypothalamic-pituitary-gonadal axis in target of cadmium toxicity. An update of recent studies and potential therapeutic approaches. *Food Chem. Toxicol.* 59, 395-404. DOI: 10.1016/j.fct.2013.06.024
- Mena-Carrasco M., Saide P., Delgado R., Hernández P., Spak S., Molina L., Carmichael G. and Jiang X., 2014. Regional climate feedbacks in central Chile and their effect on air quality episodes and meteorology. *Urban Climate* 10, 771-781. DOI: 10.1016/j.uclim.2014.06.006
- MMA, 2014. Actualización y sistematización del inventario de emisiones de contaminantes atmosféricos en la Región Metropolitana, 2014. Ministerio del Medio Ambiente, Chile. Available at: <http://ambiente.usach.cl/estudios/estudios.html> (last accessed on May 2018).
- Molina M. and Molina L., 2004. Megacities and atmospheric pollution. *J. Air Waste Manage.* 54, 644-680. DOI: 10.1080/10473289.2004.10470936
- Moreno F., Gramsch E., Oyola P. and Rubio M.A., 2010. Modification in the soil and traffic related sources of particle matter between 1998 and 2007 in Santiago de Chile. *J. Air Waste Manage.* 60, 1410-1421. DOI: 10.3155/1047-3289.60.12.1410
- Ostro B., Tobías A., Querol X., Alastuey A., Amato F., Pey J., Pérez N., Borrell C. and Sunyer J., 2011. The effects of particulate matter sources on daily mortality: A case-crossover study of Barcelona, Spain. *Environ. Health Persp.* 119, 1781-1787. DOI: 10.1289/ehp.1103618
- Pacyna J.M., 1998. Source inventories for atmospheric trace metals in atmospheric particles. In: *Atmospheric particles* (Harrison R.M. and van Grieken R.E., Eds.). John Wiley and Sons, 385-424 (IUPAC Series on Analytical and Physical Chemistry Environmental Systems, 5).
- Paris R., Desboeufs K.V., Formenti P., Nava S. and Chou C., 2010. Chemical characterisation of iron in dust and biomass burning aerosols during AMMA-SOPO/DABEX: Implication for iron solubility. *Atmos. Chem. Phys.* 10, 4273-4282. DOI: 10.5194/acp-10-4273-2010
- Peng R., Bell M., Geyh A., McDermott A., Zeger S., Samet J. and Dominici F., 2009. Emergency admissions for

- cardiovascular and respiratory diseases and the chemical composition of fine particle air pollution. *Environ. Health Persp.* 117, 957-963.
DOI: 10.1289/ehp.0800185
- Pey J., Querol X. and Alastuey A., 2010. Discriminating the regional and urban contributions in the north-western Mediterranean: PM levels and composition. *Atmos. Environ.* 44, 1587-1596.
DOI: 10.1016/j.atmosenv.2010.02.005
- Pey J., Pérez N., Cortez J., Alastuey A. and Querol X., 2013. Chemical fingerprint and impact of shipping emission over a western Mediterranean metropolis: primary and aged contributions. *Sci. Total Environ.* 463-464, 497-507. DOI: 10.1016/j.scitotenv.2013.06.061
- Pope C.A. III and Dockery D.W., 2006. Health effects of fine particulate air pollution: lines that connect. *J. Air Waste Mana.* 56, 709-742.
DOI: 10.1080/10473289.2006.10464485
- Prendez M., Ortiz J., Cortez E. and Cassorla V., 1984. Elemental composition of airborne particulate matter from Santiago city, Chile. *JAPCA J. Air Poll. Control Ass.* 4, 54-56. DOI: 10.1080/00022470.1984.10465726
- Querol X., Viana M., Alastuey A., Amato F., Moreno T., Castillo S., Pey J., de la Rosa J., Sánchez de la Campa A., Artiñano B. and Salvador P., 2007. Source origin of trace elements in PM from regional background urban and industrial sites of Spain. *Atmos. Environ.* 41, 7219-7231. DOI: 10.1016/j.atmosenv.2007.05.022
- Richter P., Griño P., Ahumada I. and Giordano A., 2007. Total element concentration and chemical fractionation in airborne particulate matter from Santiago, Chile. *Atmos. Environ.* 41, 6729-6738.
DOI: 10.1016/j.atmosenv.2007.04.053
- Sarnat J.A., Marmur A., Klein M., Kim M., Russell A.G., Sarnat S.E., Mulholland J.A., Hopke P.K. and Tolbert P.E., 2008. Fine particle sources and cardiorespiratory morbidity: an application of chemical mass balance and factor analytical source-apportionment methods. *Environ. Health Persp.* 116, 459-466.
DOI: 10.1289/ehp.10873
- Sax S., Koutrakis P., Ruiz P., Cereceda-Balic F. and Gramsch E., 2007. Trend in elemental composition of fine particulate matter in Santiago, Chile from 1998-2003. *J. Air Waste Manage.* 57, 845-855.
DOI: 10.3155/1047-3289.57.7.845
- Seguel R., Morales R.G.E. and Leiva M., 2009. Estimation of primary and secondary organic carbon formation in PM_{2.5} aerosols of Santiago City, Chile. *Atmos. Environ.* 43, 2125-2131.
DOI: 10.1016/j.atmosenv.2009.01.029
- SINCA, 2018. Sistema de Información Nacional de Calidad del Aire. Ministerio del Medio Ambiente, Chile. Available at: <https://sinca.mma.gob.cl/>.
- Toro R., Morales R.G., Canales S., González-Rojas C. and Leiva M., 2014. Inhaled and inspired particulate in Metropolitan Santiago, Chile exceed air quality standards. *Build. Environ.* 79, 115-123.
DOI: 10.1016/j.buildenv.2014.05.004
- Troncoso R., de Grange L. and Cifuentes L., 2012. Effects of environmental alerts and pre-emergencies on pollutant concentrations in Santiago, Chile. *Atmos. Environ.* 61, 550-557.
DOI: 10.1016/j.atmosenv.2012.07.077
- Ulrich A., Wilchser A., Hess A., Heeb N., Emmenegger L. and Czerwinski J., 2012. Particle and metal emissions of diesel and gasoline engines. Are particle filters appropriate measures. In 16th conference on combustion generated nanoparticles. Zurich, Switzerland.
- Valdés A., Zanobetti A., Halonen J.I., Cifuentes L., Morata D. and Schwartz J., 2012. Elemental concentrations of ambient particles and cause specific mortality in Santiago, Chile: A time series study. *Environ. Health* 11, 82.
DOI: 10.1186/1476-069x-11-82
- Villalobos A.M., Barraza F., Jorquera H. and Schauer J.J., 2015. Chemical speciation and source apportionment of fine particulate matter in Santiago, Chile. *Sci. Total. Environ.* 512-513, 133-142.
DOI: 10.1016/j.scitotenv.2015.01.006
- WHO, 2013. Public health, environmental and social determinants of health (PHE). World Health Organization. Available at: http://www.who.int/phe/health_topics/outdoorair/databases/en/index.html.
- Wang R., Balkanski Y., Boucher O., Ciais P., Peñuelas J. and Tao S., 2015. Significant contribution of combustion related emissions to the atmospheric phosphorus budget. *Nat. Geosci.* 8, 48-54.
DOI: 10.1038/ngeo2324