

Anais da Academia Brasileira de Ciências

ISSN: 0001-3765 aabc@abc.org.br Academia Brasileira de Ciências Brasil

MARIA, TATIANA F.; VANAVERBEKE, JAN; VANREUSEL, ANN; ESTEVES, ANDRÉ M.
Sandy beaches: state of the art of nematode ecology
Anais da Academia Brasileira de Ciências, vol. 88, núm. 3, 2016, pp. 1635-1653
Academia Brasileira de Ciências
Rio de Janeiro, Brasil

Available in: http://www.redalyc.org/articulo.oa?id=32748476005



Complete issue

More information about this article

Journal's homepage in redalyc.org





Anais da Academia Brasileira de Ciências (2016) 88(3 Suppl.): 1635-1653 (Annals of the Brazilian Academy of Sciences)
Printed version ISSN 0001-3765 / Online version ISSN 1678-2690
http://dx.doi.org/10.1590/0001-3765201620150282
www.scielo.br/aabc

Sandy beaches: state of the art of nematode ecology

TATIANA F. MARIA¹, JAN VANAVERBEKE², ANN VANREUSEL² and ANDRÉ M. ESTEVES³

¹Universidade Federal do Estado do Rio de Janeiro/UNIRIO, Departamento de Ecologia e Recursos Marinhos, Av. Pasteur, 458, 22290-240 Rio de Janeiro, RJ, Brasil ²Ghent University, Biology Department, Marine Biology, Krijgslaan 281- S8, B-9000 Ghent, Belgium ³Universidade Federal de Pernambuco, Departamento de Zoologia, Av. Prof. Moraes Rêgo, s/n, Cidade Universitária, 50670-901 Recife, PE, Brasil

Manuscript received on May 25, 2015; accepted for publication on September 2, 2015

ABSTRACT

In this review, we summarize existing knowledge of the ecology of sandy-beach nematodes, in relation to spatial distribution, food webs, pollution and climate change. We attempt to discuss spatial scale patterns (macro-, meso- and microscale) according to their degree of importance in structuring sandy-beach nematode assemblages. This review will provide a substantial background on current knowledge of sandybeach nematodes, and can be used as a starting point to delineate further investigations in this field. Over decades, sandy beaches have been the scene of studies focusing on community and population ecology, both related to morphodynamic models. The combination of physical factors (e.g. grain size, tidal exposure) and biological interactions (e.g. trophic relationships) is responsible for the spatial distribution of nematodes. In other words, the physical factors are more important in structuring nematodes communities over large scale of distribution while biological interactions are largely important in finer-scale distributions. It has been accepted that biological interactions are assumed to be of minor importance because physical factors overshadow the biological interactions in sandy beach sediments; however, the most recent results from in-situ and ex-situ experimental investigations on behavior and biological factors on a microscale have shown promise for understanding the mechanisms underlying larger-scale patterns and processes. Besides nematodes are very promising organisms used to understand the effects of pollution and climate changes although these subjects are less studied in sandy beaches than distribution patterns.

Key words: biodiversity, benthos, distribution patterns, food webs, climatic changes.

INTRODUCTION

Sandy beaches are dynamic ecosystems driven by prominent physical processes that shape the habitat for different functional and taxonomic groups. The term "sandy beach" can be used to describe a wide range of environments, from high-energy open-

Correspondence to: Tatiana F. Maria E-mail: tatiana fabricio@yahoo.com.br

ocean beaches to sheltered estuarine sand flats (McLachlan 1983). Sandy beaches are, in general, dynamic environments occurring worldwide along ice-free coastlines, and located at the transition between the land and a waterbody such as oceans, seas or lakes. The beach sediment may be supplied by rivers or by the erosion of highlands adjacent to the coast, and the sea may also contribute to the sediment supply through input of biogenic

structures (animal skeletons, coral and shell fragments) (adapted from McLachlan and Brown 2006). Here, the term sandy beach was used in a narrower context, applying only for high-energy open-ocean or pocket beaches ranging from the reflective to the dissipative extremes.

The sediment is mainly composed of quartz and/or carbonate sands of terrestrial and marine origin, respectively. Small amounts of feldspar, basalt and heavy minerals can also contribute to the sediment composition. Important features to characterize sediments are the grain size, the sorting coefficient and the angularity; all of them influence the porosity of sediments. The combination of permeability and penetrability determines the volume of water percolation, its drainage, and the oxygen penetration in sandy-beach sediments (McLachlan and Brown 2006).

The sandy beach ecosystem can be divided horizontally into different zones consisting of the foredune, backshore, swash/shoreline, surf zone and nearshore (Defeo and McLachlan 2005); vertically, it is represented by the pelagic and benthic systems. Sandy-beach endobenthic communities were largely neglected by ecologists until Remane (1933) initiated the first survey in Germany, with a focus on the benthic community. Today, sandy beaches are still less studied than most other coastal systems (Defeo and McLachlan 2011), and most research on intertidal sandy beaches has been concentrated on macrofauna and on birds (see Cornelius et al. 2001, Defeo and McLachlan 2005, for a review). The less-prominent sandybeach meiofauna has received considerably less attention. For several years, meiofauna research on sandy beaches was dedicated to general surveys at higher taxonomic levels, and the first investigations focusing on the composition of sandy-beach nematode communities began only in the 1970s (Heip et al. 1985). More recent investigations have dealt with patterns of macro- (10³m), meso-(m) and microscale (10⁻²m) distributions and are more often framed in a context of coastal zone management. The macroscale pattern consists of differences between beach types and latitudes, the mesoscale pattern refers to distribution and community structure alongshore and across shore transects and finally the microscale pattern is related to distance between millimeters and few meters (McLachlan and Brown 2006). These three patterns of distribution are used in the context of this review and we chose to focus on nematodes because they are the most dominant group of the meiofauna in soft sediments and are supposed to act as a link for the flux of energy since they can be the food source for macrofaunal organisms and fishes (Esteves and Genevois 2006).

Although nematodes act as important nutritional resources for macroscopic organisms, the functional role of these organisms in sandy beach ecosystem is not well established since this environment is supposed to support three partially unconnected food-webs: a discrete food web consisting of interstitial organisms, a microbial loop and a macroscopic food web (McLachlan and Brown 2006). The relative importance of these three food webs is conceptually different among beach types; for instance, on cold-temperate sandy beaches, a diverse interstitial food web dominates on high wave energy while nematodes and the macroscopic food web are more important in dissipative shores (Menn 2002).

Beaches are squeezed between rising sea level on the marine side and expanding human population and development on the landward side (Schlacher et al. 2008). Therefore, this ecosystem faces numerous threats coming from both directions. Pollution, mining, disruption of sand transport and tourism development are threats mainly originated in the terrestrial side (Brown and McLachlan 2002) while the vulnerability to the impacts of the climate change is more related to the processes occurring in the sea side, such as sea level increase, shore erosion, acidification and increase of sea water temperature (McGlone and Vuille 2012).

The aim of this review is to summarize what is known for sandy-beach nematodes in terms of spatial distribution patterns, temporal variability, food webs, pollution, and climate change. We believe that this review will open a new niche for future researchers to fill gaps in the understanding of nematode ecology.

MATERIALS AND METHODS

A bibliographic survey was done using Web of Science®, SCOPUS and Google Scholar considering works published until early 2015. Solely papers published in scientific journals and those with an ecological purpose related to meiofauna and nematode community from the intertidal region of sandy beaches were selected. Taxonomical articles and those ecological related to subtidal sampling design were largely excluded from the analysis. The selected studies were classified according to 1) type of benthic association (meiofauna or nematode), 2) pattern of distribution (macro-, meso- and microscale), 3) any other ecological approach including temporal variation, pollution, coastal management, colonization, natural impact, recreational activity, climate change and food web, 4) number of sampled beaches (Table I).

RESULTS AND DISCUSSION

In order to provide an overview of world investigations for meiofauna and nematode sandy-beach ecology, we have summarized this information in a world map showing where the studies were conducted (Fig.1) followed by the adopted approach (Table I). Of the 87 meiofauna studies, approximately 50% deal with nematodes (Table I). Our discussion is based on the main topics included in the built table and it was subdivided in the following sections.

PATTERNS OF NEMATODE DISTRIBUTION

Since the first scheme of horizontal faunal zonation was published by Mortensen (1921), several

attempts have been made to define well-demarcated zones in sandy-beach sediments, by using physical factors (Salvat 1964, Pollock and Hummon 1971, McLachlan 1980), macroinvertebrates (Dahl 1952, Salvat 1964, 1967, McLachlan and Jaramillo 1995, Defeo and McLachlan 2005, for a recent review) or meiofauna (Blome 1983, Rodriguez et al. 2001, Gheskiere et al. 2004, 2005a, Kotwicki et al. 2005a, Gingold et al. 2010, Maria et al. 2013b, c).

The pattern of nematode distribution in sandy beaches is explained in detail in the following subsections. The variables are discussed below in relation to sandy-beach nematode communities.

Macroscale distribution of nematode communities

This topic has received relatively little attention. Two studies concerning the macroscale latitudinal distribution of sandy-beach nematodes and their diversity suggest that the present pattern follows the trend of increasing diversity toward the tropics (Nicholas and Trueman 2005, Lee and Riveros 2012). Macroscale nematode studies exploring the full range of beach types have not yet been undertaken, but the limited information available shows that nematode diversity is highest in coarsegrained, intermediate sandy beaches (Gheskiere et al. 2005a) or in sheltered conditions (Hourston et al. 2005, Urban-Malinga et al. 2005). It indicates that the a high diversity can be found in the full range of beach types, but the above mentioned references do not use the same methodological strategy what limited the interpretation of the results. A comprehensive understanding of the nematode diversity pattern among different beach types will be only possible when an equal sampling design would be adopted. For instance, the sampling design used to understand the distribution patterns of polychaetes in different beach types could be adopted (Di Domenico et al. 2008); in this study, it is indicated that reflective sandy beaches are more diverse than intermediate and dissipative ones.

TABLEI

Nematodes, CC: Climate change, Co: colonization, CM: Coastal management; DT: Patterns of Distribution, FW: Food webs P: pollution, TP: Temporal variation. Global overview of meiofauna and nematodes studies after 1970, with indication of the focus of the study and the number of beaches included. M: Meiofauna, N:

Code for		,		N of studied	
fig. 1	Keterence	Fauna	Focus	beach	Main Findings
	Boaden and Platt 1971	M and N	DT		Daily tidal variations and biotic interactions affect the vertical distribution of meiofauna
2	Gray and Rieger 1971	M and N	DT	_	Highest meiofauna richness and diversity in less exposed (finer grain) beach
3	Harris 1972	M	DT		High temperatures in the upper beach is a limiting factor for meiofauna zonation
4	Ott 1972	Z	DT		Presence of a typical nematofauna in sulfide systems
S	Hulings and Gray 1976	\mathbb{Z}	DT	68	Physical factors control the meiofauna abundance (e.g. wave, tide and currents) in tidal beaches while biological interactions (e.g. predation and competition) may control abundance in atidal beaches
9	McLachlan et al. 1977	M	DT and P	1	Meiofauna can occur down to 35 cm. Sewage can cause an increase of the meiofauna abundance, specially the nematode one
7	Platt 1977	Z	DT	1	Horizontal distribution is driven by physical factors (e.g. grain size) and availability of food while vertical distribution is limited by oxygen availability
∞	Munro et al. 1978	\mathbb{Z}	DT	7	Tropical beach shows higher respiration rates, less microbial production and meiofauna biomass when compared to a temperate one
6	Fricke et al. 1981	M and N	Ь	7	Harpacticoid copepods are more sensitive to oil pollution than nematodes; meiofauna can be recovered six months after an oil spill
10	McLachlan and Harty 1982	\boxtimes	Ъ	1	Oligochaetes are more sensitive to oil pollution than nematodes; meiofauna can be recovered five months after an oil spill
111	Ansari and Ingole 1983	\mathbb{Z}	DT	9	Meiofauna is more abundant in the upper 4 centimeters
12	Blome 1983	Z	DT	1	Nematodes are more abundant in the oxic layers and shows a horizontal zonation; vertical migration is attributed to temperature variations
13	Sharma and Webster 1983	Z	DT	7	Highest nematode density in well sorted medium grain size: most nematodes occur in the top 4 cm, but those that occur deeper seems to be tolerant to anoxic conditions
14	Ansari et al. 1984	\boxtimes	DT	7	Highest density in the upper tide level for sheltered beach and highest density in the midtide level for exposed beach
15	Ansari et al. 1990	M	DT	9	Meiofauna abundance is reduced in zones of the beach exposed to desiccation
16	Pattnaik and Rao 1990	M	DT	1	Harpacticoid copepods are the dominant group; meiofauna is abundant in the midtide level occuring up to 20 cm deeper; their spatial distribution is controlled by physical factors.
17	Alongi 1990	Z	TP		Highest nematode densities in austral autumn and winter and lowest densities in spring and summer
18	Ólafsson 1991	\boxtimes	DT	4	Meiofauna is more abundant in the top 10 cm and nematodes show the most uniform vertical distribution in 95 cm deeper
19	Szymelfenig 1995	M	DT	10	The occurrence of meiofauna taxa is highly correlated to climatic characteristics of the habitats
20	Gourbault et al. 1995	M and N	DT	13	Harpacticoid copepods are more abundant than nematodes; this latter taxa shows higher diversity in coarse sediment
21	Long and Ross 1999	Z	DT	1	Nematodes occur at 30cm deep and their vertical distribution is related to food availability

TABLE I (continuation)

Code for fig. 1	Reference	Fauna	Focus	N of studied beach	Main Findings
22	Nicholas and Hodda 1999	Z	DT		Higher nematode abundance in the mid and high tide level of the beach; persistence of nematode zonation during a short period (24h and under calm conditions
23	Armonies and Reise 2000	M	DT	1	Turbellarian is the dominant taxa; highest meiofauna richness at midtide level
24	Nicholas 2001	Z	DT	-	Nematodes occur at 60 cm deep; highest densities in the warmer months of the year
25	Rodriguez et al. 2001	M	DT	3	Highest density of the meiofauna in reflective beach and at upper and midtide level
26	Gheskiere et al. 2002	M and N	DT	-	Increase in meiobenthic density towards low tide level; presence of nematode horizontal zonation
27	Menn 2002	M	DT and FW	7	Nematodes are an important food sources for higher trophic level in intermediate beaches while macrofauna plus nematodes are the food sources in dissipative beaches
28	De Oliveira and Soares-Gomes 2003	M	Ь	П	No relationship between sewage disposal and meiofauna
29	Moellmann and Corbisier 2003	M	Ь	7	Nematodes migrate to the deeper layers of the sediment in intense trampling areas
30	Souza-Santos et al. 2003	M	DT	1	Physical factors (e.g. grain size, chla, phaeopigments, sediment skewness) are the main factors structuring the meiofauna community seasonally
31	Rodríguez et al. 2003	M	DT	10	Richness and biomass of meiofauna increases toward very exposed and coarse beaches
32	Gheskiere et al. 2004	Z	DT	1	Higher nematode density in the low tide level of the beach; high richness in the midtide level; presence of nematode zonation
33	Rodriguez 2004	M	DT	-	Highest meiofauna density in the midtide level
34	Urban-Malinga et al. 2004	M and N	DT		Turbellarian is the dominant meiofauna taxa; sheltered beach showed a higher meiofauna density and respiration rates than exposed beach
35	Gheskiere et al. 2005a	Z	DT	7	Highest nematode density and diversity in coarse-grained sandy beach; presence of isocommunities only in the upper beach
36	Kotwicki et al. 2005a	M	DT	5	Meiofauna show high density in the upper 5 cm; high densities in the midtide level or in the lower tide level; presence of horizontal zonation
37	Calles et al. 2005	M and N	DT	7	Turbellarian is the dominant taxa in the sheltered beach; high meiofauna density in the exposed beach; no difference between meiofauna diversity of sheltered and exposed beaches
38	Kotwicki et al. 2005b	M	DT	13	Nematodes are dominant in warmer regions of the globe while turbellarian are dominant in cold water regions; no latitudinal trend was found for meiofauna density or diversity
39	Lee and Correa 2005	M	Ь	12	Reduction of density and diversity in sites impacted by mine tailing.
40	Gheskiere et al. 2005b M and N DT	M and N	DT and P	4	Lowest species diversity in the upper tide level, especially in touristic beaches
41	Hourston et al. 2005	Z	DT	ю	Highest density in the sheltered habitat, but declining during winter; weak zonation reflecting the presence of a small tidal range
42	Nicholas and Trueman 2005	Z	DT	8	Richness increases with decreasing latitude

TABLE I (continuation)

Code for	Reference	Fauna	Focus	N of studied	Main Findings
43	Urban-Malinga et al. 2005	Z	DT	2	Oligochaete is the dominant taxa; highest diversity in sheltered beaches; density increases towards the low water level
44	Moreno et al. 2006	M and N	DT	1	Highest density and diversity in the swash level
45	Harriague et al. 2006	M	DT	_	Higher density in the surfzone than in the swash level
46	Gheskiere et al. 2006	M	$_{\rm CM}$	-	There is no reduction in density or diversity of mechanical cleaning areas.
47	Urban-Malinga and Moens 2006	M	FW	2	Fastest carbon remineralization in exposed coarse sandy beach
48	Nicholas 2006	M and N	DT	-	Nematodes occur at 60 cm deep
49	Albuquerque et al. 2007	M	DT	П	Tardigrades are the most dominant taxa; highest density in the midtide level
50	Papageorgiou et al. 2007	M	FW	-	Bacteria are a potential food source for meiofaunal organisms
51	De Jesús Navarette 2007	Z	DT	2	Highest richness in coarse-grained beach
52	Mundo Ocampo 2007	Z	DT	7	Highest density and diversity in the coarse-grained beach
53	Rodríguez et al. 2008	M	DT	10	Eh and water content are the best environmental variables that explains meiofaunal densities
54	Harriague et al. 2008	M	DT	9	Food availability is a key role in structuring meiofauna community
55	Liu et al. 2008	Z	DT	П	Physical factors (e.g. temperature, $chla$ and grain size are responsible for the nematode community structure
99	Maria et al. 2008	Z	DT	33	Highest richness in very coarse sandy beach
57	Urban-Malinga et al. 2008	Z	DT	П	Fastest organic matter weight loss in the highest diverse habitat of the beach - midtide level
58	Delgado et al. 2009	M	DT	23	Gravel and organic matter are best correlated to meiofauna composition and abundance; absence of pattern in meiofaunal abundance along a degree of exposure
65	Grzelak et al. 2009	M	Z	3	Fast recover of meiofauna community after a Tsunami event
09	Gomes and Rosa Filho 2009	M	DT	П	Highest density and diversity in the midtide level; presence of zonation
61	Urban-Malinga and Burska 2009	M	DT	П	Wrack-associated meiofauna community and density from upper and lower shores are similar
62	Veiga et al. 2010a	M	Ь	6	Low densities in exposed sandy beaches contaminated by polycyclic aromatic hydrocarbons
63	Veiga et al. 2010b	M	DT	7	Highest density in the top 5 cm; oxygen is the limiting factor especially for copepods and ostracods
64	Yamanaka et al. 2010	M	DT	3	Positive correlation between meiofauna abundance and exposure index
65	Gingold et al. 2010	Z	DT	-	Highest diversity in the midtide level; presence of similar zonation patterns in topographical heterogeneous habitat

TABLE I (continuation)

Code for fig. 1	Reference	Fauna	Focus	N of studied beach	Main Findings
99	Gingold et al. 2011	Z	DT	1	Patch distribution is more prone to occur in the runnel habitat than in the sandbar
<i>L</i> 9	Nanajkar and Ingole 2010	Z	Ь	_	Highest density in organically polluted sites
89	Barnes et al. 2011	\mathbb{Z}	DT	9	Highest density in the mid and lower tide level; diversity decreases towards the tropics
69	Maria et al. 2011a	M and N	FW	1	Benthic diatoms are the preferential food source of meiofaunal organisms
70	Maria et al. 2011b	Z	Co		Polychaetes seems to facilitate the colonization of azoic sediments by nematodes
71	Riera et al. 2011	M	TV		Total abundance and community structure are similar along a year of study
72	Lee and Riveros 2012	Z	DT	99	Richness decreases with increasing latitude
73	Mantha et al. 2012	\mathbb{Z}	DT	5	Harpacticoid copepods are thedominanttaxa
74	Maria et al. 2012	M and N	DT and FW		Biological factors are important in controlling the vertical distribution of nematodes; interstitial and macrofaunal food webs seems do not compete for food sources
75	Riera et al. 2012	M and N	DT	2	Copepods are dominant in medium sand while nematodes are dominant in fine sand, i.e. grain size influences in the meiofauna composition
92	Gingold et al. 2013	Z	CC	1	Temperature increase shows a negative impact on dominant predator species
77	Maria et al. 2013a	Z	Co	1	Polychaetes did not affect the colonization of azoic sediments by nematodes
78	Maria et al. 2013b	Z	DT	1	Runnels and sandbar harbors different communities showing different zonation patterns
79	Maria et al. 2013c	Z	DT	2	Highest density in the upper tide level; gravel is the most important variable that explains the nematode community
80	Schlacher and Hartwig 2013	M	FW	_	Meiofauna community is strongly influenced by sediment grain size; the density of nematodes and ostracod is controlled by bottom up processes
81	Kang et al. 2014	M and N	Ь		Reduction of density in oil polluted sites; meiofauna can be recovered one month after an oil spill
82	Kotwicki et al. 2014	\mathbb{Z}	DT	7	Mean grain size and sorting coefficient showed great influence on meiofauna variation
83	Cooke et al. 2014	M	DT	ю	Patch distribution of the meiofauna in the upper tide level, especially harpacticoid copepods and gastrotrichs
84	Netto and Meneghel 2014	M	FW	_	Surfzone diatoms are an important food source for meiofaunal organism of high-energy sandy beaches
85	Sun et al. 2014	\mathbb{M}	Ь	5	Anthropogenic disturbed beaches showed high meiofauna abundance reflecting high chl a and dissolved oxygen levels
98	Venekey et al. 2014a	Z	DT	1	Lowest density in dry months; no influence of the tidal cycle in the nematode density
87	Venekey et al. 2014b	Z	DT		Seasonal and spatial variation of nematode communities in a tropical sandy beach

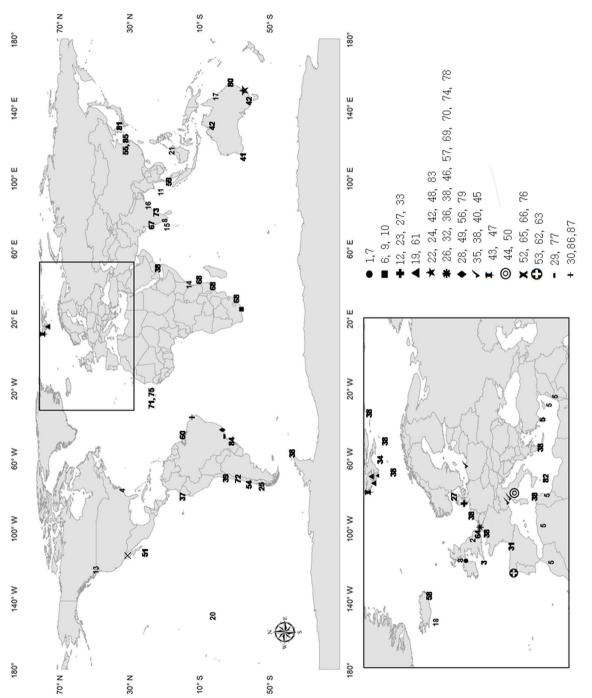


Figure 1 - Global distribution of sandy beach meiofauna studies after 1970 (for studies before 70' see Heip et al. 1985). See Table I for authors and a detailed information of the studies.

Mesoscale distribution of nematode communities

Concerning the horizontal zonation in across-shore transect on a single beach, granulometry, organic input, temperature and salinity are the key factors governing the horizontal nematode distribution (Platt 1977, Gheskiere et al. 2004, Urban-Maligna et al. 2005, Moreno et al. 2006, Maria et al. 2012), whereas biological interactions between macro- and meiofauna and/or among meiofauna organisms have only an indirect effect (Maria et al. 2011b, 2012, 2013a). High diversity is often found where there is an optimal balance of these variables, i.e., physical and chemical conditions are intermediate (Gheskiere et al. 2004, Hourston et al. 2005, Gingold et al. 2010). However, some, more heterogeneous beaches display intertidal sandbars intercalated by depressions, which retain water when tides recedes, across their wide intertidal zone (Masselink et al. 2006); this habitat is therein called runnel and those beaches that bear it are so-called macrotidal ridge-and-runnel beach. On these beaches, for example De Panne Beach in Belgium, studied by Maria et al. (2013b), the two microhabitats (runnels and sandbars) contain different nematode communities, which are reflected in dissimilar across-shore zonation. Three nematode associations (upper, middle and lower beach) are much more evident in the sandbars than in the runnels. These three biological zones on the sandbars were in accordance with the tidal zonation previously observed by Gheskiere et al. (2004) in the same study are contradicting the previous knowledge that the horizontal nematode zonation may just persist for short intervals of time and under calm conditions (Nicholas and Hodda 1999). The main difference between the two most recent studies is that the latter excluded the runnel microhabitat. The inclusion of the runnels by Maria et al. (2013b) showed that the shift over three horizontal nematode communities found in the sandbar was interrupted by the runnel

communities. These results contrast somewhat with the results of Gingold et al. (2010) for a macrotidal ridge-and-runnel beach in the Gulf of California, where both sandbars and runnels showed the same pattern of across-shore zonation. We can conclude that next to the horizontal mesoscale, the presence of runnels may influence the nematode zonation. Although both habitats (sandbars and runnels) are horizontally distributed over the sandy beach interface and did not show consistent differences in grain size or chlorophyll *a* content, they were not under the same horizontal gradient of air exposure during low tide.

Based on these results, it is not so much the combination of a physical characteristic that determines the nematode community structure, but rather the degree of variation of these physical factors that affect the nematode horizontal distribution in the intertidal zone.

Another uncommon way to analyze the mesoscale distribution of nematodes is to examine an along-shore transect in the intertidal zone. In this case, samples that are far from each other are more heterogeneous (Nicholas and Hodda 1999, Gingold et al. 2011); however, if the beach has different microhabitats along the across-shore transect, as in the case of a macrotidal ridge-and-runnel beach, the degree of patchiness is more accentuated in the microhabitat with calmer conditions (Gingold et al. 2011).

Microscale distribution

In relation to microscale vertical distribution, desiccation and oxygen availability are considered to be ultimate factors controlling the vertical distribution of the meiofauna (McLachlan 1978, Coull 1988, Maria et al. 2012). Oxygen is a significant limiting factor only in sheltered beaches. In this habitat, marine nematodes are restricted to the first centimeters of the sediment (McLachlan 1978), while in more exposed conditions, nematodes can be distributed more deeply (Urban-

Malinga et al. 2004). This phenomenon has led to the use of vertically undivided cores in meiofauna sandy-beach sampling, or the use of cores divided into large intervals of cm (i.e., larger than 1cm) (Blome 1983, Sharma and Webster 1983, Long and Ross 1999, Urban-Malinga et al. 2005). Moreover, samples are usually taken during low tide. This is a sampling strategy initially adopted for the macrofauna to avoid contamination by tidal migrants. It is however, also known that nematodes can migrate downward (Boaden and Platt 1971) or upward (McLachlan et al. 1977, Maria et al. 2012) in the sediment column at high tide. Therefore, to better understand the vertical distribution of nematodes, a finer-scale sampling scheme should be adopted, using cores subdivided into thin layers of a few centimeters, when dealing with protected and dissipative sandy beaches. Joint et al. (1982) has already demonstrated that nematodes showed different vertical distribution patterns at mm scale in an intertidal sandflat.

The relationship between the vertical distribution of sandy-beach nematodes and the tidal cycle was demonstrated by Boaden and Platt (1971) and McLachlan et al. (1977). These two studies were rather contradictory. The former authors showed that nematodes move down in the sediment, escaping from the upper layers and thus avoiding being washed away by the turbulent conditions. McLachlan et al. (1977) demonstrated that meiofauna living above the depth of the permanent water table undergoes vertical movements coupled with the tidal cycle, i.e., move up during submersion and vice-versa. Since the depth fluctuation of the groundwater level is directly linked to the different tidal stages (Urish and McKenna 2004), the upward movements of the meiofauna are closely linked to the tidal cycle. Indeed, the vertical distribution of many nematode species in the upper sediment layers of De Panne Beach during submersion was mainly explained by upward passive transport from deeper layers. Downward movements during emersion were explained by a combination of passive and active transport (Maria et al. 2012). Passive transport could be attributed to the drop in the underground water level, while active migration would occur to avoid harsh conditions created by drying of surface layers (McLachlan et al. 1977).

Although the investigation of the importance of the tidal cycle for the vertical distribution of the nematodes by Maria et al. (2012) was restricted to the upper five centimeters of the middle beach level, these results could be extrapolated to the other intertidal beach levels by combining the observations of the groundwater discharge in the beach sediments during the different tidal stages. The water table is very close to the sediment surface toward the low-water line during low tide. Therefore, here the vertical rise of the water table during the incoming tide would be more discrete than toward the high-water line. Consequently, upward passive transport of nematodes in the low intertidal beach would be less likely, so that nematodes would not dramatically modify their vertical distribution over the tidal cycle in this part of the beach. On the other hand, changes in the nematode vertical distribution over the tidal cycle would be more evident in the upper beach.

Biological interactions may also affect the vertical distribution of nematodes in the sediment, and may occur among different sizes of organisms. The role of biological interactions, such as competition and predation, in regulating macrofauna zonation remains little known (McLachlan and Jaramillo 1995). From the point of view of the meiofauna, biological interactions were only suggested to be important for atidal beaches (Hulings and Gray 1976), but Maria et al. (2012) showed that the species-specific vertical distribution of nematodes over a tidal cycle is clearly driven by biological interactions. The importance of predation was evidenced by the predacious nematode *Enoplolaimus litoralis*. This

species was recorded in the layers just above the deposit-feeder Daptonema normandicum, which inhabited the subsurface 2-5-cm layers during submersion (Maria et al. 2012). The same pattern of predator-prey segregation was also observed in the coarsest sediments (<170µm and less than 5% silt) of a mudflat under high hydrodynamic stress, and was explained by food preferences (Stevaert et al. 2003). During exposure in De Panne Beach, E. litoralis was no longer restricted to the 1-3-cm layers as during submersion, but was found deeper in the sediment. This change may be the consequence of active migration, since another thoracostomopsid nematode, Enoploides longispiculosus, has shown high motility to actively catch deeper-living prey when the surface sediment was not water-saturated and the groundwater level was lowered (Steyaert et al. 2001). This migration took place during ebb tide, and also coincided with a low abundance of suitable prev species within the upper 5cm.

The importance of competition in driving the vertical distribution of species is confirmed to some extent by the vertical segregation of nematode species in the sediment. Species with a more ¹³C-enriched diet, such as Sigmophoranema rufum, were abundant in the upper two centimeters of the sediment, whereas species belonging to epistrate (2A) and non-selective feeder (1B) groups that exploit a more ¹³C-depleted diet showed higher densities in the subsurface layers - 2 to 3 cm - (Maria et al. 2012). The importance of food competition between S. rufum and 1B/2A nematodes was supported by the results of an enrichment experiment, since the diatom uptake of epistrate nematodes was highest when the same food sources (diatoms) were offered to the nematode community. This may indicate that 2A nematodes are better competitors (Maria et al. 2011a).

NEMATODE COMMUNITIES IN BEACH FOOD WEBS

Food webs in sandy beaches differ with the beach type, and particularly with the degree of coupling between the beach and the surf zone (McLachlan and Brown 2006). Two distinct ecosystems are recognized: (1) beaches with little or no surf zone (reflective beaches), which are dependent on food input from the sea; and (2) beaches with extensive surf zones (dissipative beaches), which have high primary production from the surf diatoms that shows a vertical migration from the sediment deep in the surf zone to the water (McLachlan and Brown 2006). This ecosystem may support three partially unconnected food webs: the interstitial food web, the microbial loop (which is restricted to the surfzone and has been little investigated), and the macroscopic food web (Heymans and McLachlan 1996).

There are two studies that investigated the importance of nematodes in sandy beach food webs. Menn (2002) showed that a high wave energy beach has an impoverished food web due to a more diverse meiofauna while a less dynamic sandy beach showing a high dominance of nematodes in the meiofauna community can support a richer food web.

The second study assessed the integrative food web of the middle part of De Panne Beach by means of stable isotope analyses (Maria et al. 2012). Vascular plants were not included as a possible food source, since they are important primary producers of the supralittoral zone of Belgian sandy beaches (Speybroeck et al. 2008) and were assumed to be of less importance for the middle beach of De Panne, where a concrete dike separates the dunes from the beach (Gheskiere et al. 2004). Most of the meiobenthic species (nematodes and copepods) utilize more carbon-enriched food sources than the macrofaunal organisms, indicating that there was no competition for food between these two classes of organisms. Phytoplankton and suspended detritus seem to be more important food sources for macrobenthic organisms on dissipative and reflective sandy beaches (Bergamino et al. 2011, Maria et al. 2012), while microphytobenthos

(MPB) is utilized as a carbon source by nematodes, copepods, turbellarians and a few macrofaunal species, such as two species of the amphipod Bathyporeia (Maria et al. 2012). Laboratory experiments also indicated a higher uptake of benthic diatoms by these organisms (Maria et al. 2011a). Nonetheless, MPB is considered a limiting food source for sandy-beach organisms due to its low primary productivity, probably because of the strong hydrodynamic forces acting on the beach interface (McLachlan and Brown 2006); the preference for benthic over planktonic diatoms, as demonstrated by Maria et al. (2011a), leads to us to speculate that the absence of a diatom biofilm in sandy beaches, a condition extremely common on tidal flats, can be attributed to the rapid consumption of diatoms by benthic sandy-beach organisms.

Although some field investigations and experimental approaches demonstrated that nematodes may serve as food source for large organisms, such as shrimps and fishes (see Coull 1990, for a revision), there was no explicit trophic link between macrofaunal and nematode species evidenced by isotopes analysis (Maria et al. 2012). It can be easily explained by the few numbers of species investigated in the study.

POLLUTION

A large set of features, including ease of sampling, omnipresence, high diversity, short generation time, and absence of a planktonic stage, make nematodes an excellent tool to evaluate the ecological condition of different environments (Giere 2009); however, very few studies have dealt with nematodes at lower taxonomic resolution (e.g. Fricke et al. 1981, Gheskiere et al. 2005b, Nanajkar and Ingole 2010). This may be related to the difficulty and tedium of species identification; therefore, pollution studies tend to use broader taxonomic categories and the two numerically dominant groups, nematodes and copepods, are the main focus of such studies.

As seen in Table I, the four pollution-related studies focusing on sandy-beach nematodes dealt with different subjects: oil, recreational activities, and sewage discharge. In terms of oil pollution, the nematode density was still similar to reference sites when the sediment of the polluted site was not mechanically removed (Fricke et al. 1981). This finding contradicts the expected pattern of reduction in density at sites where an oil spill occurred (Wormald 1976, Giere 1979, Danovaro et al. 1995, Kang et al. 2014), but other field and experimental studies have indicated that nematodes are insensitive to oil pollution (Boucher 1980, Warwick et al. 1988, respectively). On the other hand, a mechanical treatment of the oil-polluted sediment had more impact in the nematode density than the pollution by oil per se, and more individuals were found in the deeper layers (Fricke et al. 1981).

There is little controversy around the time required for meiofauna recovery in areas subjected to oil spills. Fricke et al. (1981) found that meiofauna had recovered six months after an oil spill on the coast of South Africa (SA), while this time period was not enough for meiofauna recovery in three beaches located on the coast of Galicia (GC) (Veiga et al. 2010a). The lack of a similar response here is probably related to the difference in the magnitude of the oil spill (31,000 tons in SA x 50,000 tons in GC), and the local hydrodynamics (exposed in SA and under intermediate conditions in GC). In situ experiments have shown that sandybeach meiofauna can recover their density and community composition one month after an oil spill when very small quantities of oil are spilled (Kang et al. 2014).

The discharge of domestic sewage onto sandy beaches is a common problem in coastal areas, especially in developing countries, where a large part of the population may develop coastal areas in an unregulated manner. Nanajkar and Ingole (2010), studying the impact of chronic sewage discharge on a tropical sandy beach, observed that

nematodes responded to the organic enrichment with an increase in density and dominance of one genus (*Daptonema*) in the areas closest to the sewage outflow. These authors believed that the non-selective deposit-feeding activity of *Daptonema* aided the bioremediation of the site.

From the point of view of beach management, this ecosystem has great socio-economic value as a recreational area and is therefore considered a key tourist destination during holiday periods. The recreational use of sandy beaches has a significant impact in terms of reduction of meioand nematofauna diversity, which responds by shifting to a large number of species with small size, rapid growth and high rates of reproduction (r-strategists). Nevertheless, this reduction may only be associated with the upper beach, which can be mechanically cleaned daily (Gheskiere et al. 2005b). On the other hand, in intense trampling activity in the intertidal area leads the meiofauna to migrate downward to layers deeper than 5cm (Moellman and Corbisier 2003).

CLIMATE CHANGE

Concerns about global warming have dramatically increased and it is widely understood that sandy beaches will be drastically affected, especially by sea-level rise. To date, very few studies have dealt with this subject (e.g. Kont et al. 2003, Lock et al. 2011, Mead et al. 2013). In terms of nematodes, only the study of Gingold et al. (2013) has concentrated on this group in sandy-beach environments. Laboratory microcosm experiments have shown that high temperatures lead to loss of important predacious and omnivorous nematodes that are important for the top-down control of the community, consequently leading to a change in the food web.

Concerning sea-level rise, the nematode intertidal community will be immersed more often or will become permanently submersed; therefore, we can speculate that many intertidal species will be unable to withstand long periods submerged.

This may eliminate many of these species and consequently reduce biodiversity in areas where the nematode biodiversity is presently very high. Besides, the rapid sea level rise will move beaches towards a more reflective morphodynamic state which is characterized by low meiofauna abundances (Yamanaka et al. 2010). In both cases, the reduction in nematode abundance may lead to a lost in ecosystem functioning. However, this topic requires further investigation.

CONCLUSIONS

Sandy beach ecosystems are one of the less studies coastal ecosystems; it can be easily exemplified by a simple search in the Web of Science which the number of studies in this environment correspond to 10% and 25% of those realized in estuaries and mangroves, respectively. The current state of knowledge shows us that over the decades, sandy beaches have been the scene of studies focusing on community and population ecology, both in relation to morphodynamic models. The combination of physical factors (e.g. grain size, tidal exposure and degree of drainage) and biological interactions (e.g. trophic relationships and competition) is responsible for the spatial distribution of nematodes. The degree of importance of these factors is related to the kind of distribution pattern analyzed; physical factors are more important in structuring nematodes communities over large scale of distribution (e.g. macro- and mesoscale) while biological interactions are largely important in finer-scale distributions.

Although nematodes are the dominant meiofauna group and are largely used as indicator of environmental conditions it is of primordial importance to understand the natural relationship between the environment and these organisms. Therefore, to understand better the ecological processes driving the sandy-beach nematode community is necessary to increase the efforts in understanding latitudinal and beach type patterns also taking into account the environmental heterogeneity. In terms of energy, nematodes are considered an important link between micro- and macrofaunal organisms; however, isotopic studies showed that there is a clear separation between the interstitial and macrofaunal food webs at least for temperate sandy beaches.

Even though the nematode diversity is related to physical and biological interactions, it is also related to the degree of impact occurring in the environment; under disturbed conditions (i.e., organic pollution, oil pollution, beach cleaning or trampling) the diversity decreases and the density of some dominant species (especially deposit feeders) increases, the natural recovery depends on the amount and duration of the impact.

And last but not the least, the relationship between nematodes and climate change is a promising research area, which many aspects must still be evaluated for sandy-beach organisms, but a common believe is that the nematode density must decrease.

This review elucidates the current knowledge of sandy-beach nematodes and as this ecosystem is easily reached and occurs worldwide we should stimulate more and more studies that can provide a more comprehensive approach toward a better understanding of the physical and biological interactions among sandy-beach organisms. This is essential in order to understand the possible effects of key human pressures on the beach ecosystem.

ACKNOWLEDGMENTS

Renata Alves and Alex Silva are acknowledged for their help with GIS Mapping Software and the map quality, respectively. The authors thank Jannet W. Reid, JWR Associates, for the critical review of the English. The data was obtained from the PhD thesis of the first author that was sponsored by Vlaamse Interuniversitaire Raad—VLIR-UOS. Additional funding was provided by the Special Research Fund of Ghent University

(BOF-GOA 01GA1911W). André Esteves (CNPq 312143/2013-3) wishes to express his gratitude for the research fellowship granted by the Conselho Nacional de Desenvolvimento Científico e Tecnológico (CNPq). Authors are also grateful to the two reviewers (Maickel Armenteiros and the anonymous one) by their comments.

RESUMO

Nessa revisão, reunimos o conhecimento existente sobre a ecologia dos nematódeos de praias arenosas, em relação à distribuição espacial, teias tróficas, poluição e mudanças climáticas. Tentamos discutir os padrões em escala especial (macro-, meso- e microescala) conforme o seu grau de importância na estruturação das associações de nematódeos em praia sarenosas. Esta revisão fornece uma base substancial sobre o conhecimento atual dos nematódeos de praias arenosas e poderá ser usada como um ponto de partida para futuras investigações nesse campo da Ciência. Por décadas, as praias arenosas têm sido objeto de estudos enfocando a ecologia de comunidade e populacional, ambas relacionadas aos modelos morfodinâmicos. A combinação dos fatores físicos (p. ex., tamanho do grão, nível de exposição às marés) e os fatores biológicos (p. ex., relações tróficas) é responsável pela distribuição espacial dos nematódeos. Em outras palavras, os fatores físicos são mais importantes na estruturação das comunidades de nematódeos em grandes escalas, enquanto que as interações biológicas são mais relevantes em escalas menores de distribuição. Tem sido aceito que as interações biológicas seriam de menor importância porque os fatores físicos ocultam as interações biológicas nos sedimentos de praias arenosas; no entanto, estudos experimentais mais recentes (insitu e ex-situ) sobre o comportamento e os fatores biológicos (em microescala) têm mostrado resultados promissores para a compreensão dos mecanismos subjacentes aos processos e padrões em maior escala. Além disso, os nematódeos são organismos muito promissores para serem utilizados a fim de compreender os efeitos da poluição e das mudanças climáticas, ainda que esses dois tópicos sejam menos estudados em praias arenosas do que os padrões de distribuição.

Palavras-chave: biodiversidade, bentos, padrões de distribuição, teias tróficas, mudanças climáticas.

REFERENCES

- ALBUQUERQUE EFP, PINTO APB, PEREZ AAQ AND VELOSO VG. 2007. Spatial and temporal changes in interstitial meiofauna on a sandy ocean beach of South America. Braz J Oceanogr 55: 121-131.
- ALONGI DM. 1990. Community dynamics of free-living nematodes in some tropical mangrove and sandflat habitats. B Mar Sci 46: 358-373.
- ANSARI ZA AND INGOLE BS. 1983. Meiofauna of Some Sandy Beaches of Andama Islands. Indian J Mar Sci 12: 245-246.
- ANSARI ZA, INGOLE BS AND PARULEKAR AH. 1984. Macrofauna and meiofauna of two sandy beaches at Mombasa, Kenya. Indian J Mar Sci 13: 187-189.
- ANSARI ZA, RAMANI P, RIVONKER CU AND PARULEKAR AH. 1990. Macro and meiofaunal abundance in six sandy beaches of Lakshadweep islands. Indian J Mar Sci 19: 159-164.
- ARMONIES W AND REISE K. 2000. Faunal diversity across a sandy shore Mar Ecol-Prog Ser 196: 49-57.
- BARNES N ET AL. 2011. Assessment of regional and local biodiversity in tropical and subtropical coastal habitats in the East African Marine Ecoregion. Biodivers Conserv 20: 2075-2109.
- BERGAMINO L, LERCARI D AND DEFEO O. 2011. Food web structure of sandy beaches: temporal and spatial variation using stable isotope analysis. Estuar Coast Shelf S 91: 536-543.
- BLOME D. 1983. Okologie der Nematoda eines Sandstrandes der Nord see in sel Sylt. Mikrof Meeresb 88: 517-590.
- BOADEN PJS AND PLATT HM. 1971. Daily migration patterns in an intertidal meiobenthic community. Thalassia Jugosl 7: 1-12
- BOUCHER G. 1980. Impact of Amoco Cadiz oil spill on intertidal and subtidal meiofauna. Mar Pollut Bull 11: 95-101.
- BROWN AC and MCLACHLAN A. 2002. Sandy shores ecosystems and threats facing them: some predictions for the year 2025. Environ Conserv 29: 62-77.
- CALLES A, VINCX M, CORNEJO P AND CALDERON J. 2005. Patterns of meiofauna (especially nematodes) in physical disturbed Ecuadorian sandy beaches. Meiofauna Mar 14: 121-129.
- COOKE BC, GOODWIN ID AND BISHOP MJ. 2014. Small-scale spatial distribution of interstitial invertebrates on three embayed beaches, Sydney, Australia. Estuar Coast Shelf S 160: 92-101.
- CORNELIUS C, NAVARRETE SA AND MARQUET P. 2001. Effects of human activity on the structure of coastal marine birds assemblages in Central Chile. Conserv Biol 15: 1394-1404.
- COULL BC. 1988. Ecology of the marine meiofauna. In: Higgins RP and Thiel H (Eds), Introduction to the study

- of meiofauna. Smithsonian Institution Press Washington, D.C., USA, p. 18-38.
- COULL BC. 1990. Are Members of the Meiofauna Food for Higher Trophic Levels? Trans Am Microsc Soc 109: 233-246
- DAHL E. 1952. Some aspects of the ecology and zonation of the fauna on sandy beaches. Oikos 4: 1-27.
- DANOVARO R, FABIANO M AND VINCX M. 1995. Meiofauna Response to the Agip Abruzzo Oil Spill in Subtidal Sediments of the Ligurian Sea. Mar Pollut Bull 30: 133-145.
- DE JESÚS NAVARETTE A. 2007. Nematodos de los arrecifes de Isla Muyeres y Banco Chincorro Quintana Roo, México. Rev Biol Mar Ocean 42: 193-200.
- DE OLIVEIRA EB AND SOARES-GOMES A. 2003. Impact of a point source domestic sewage on intertidal meiofauna at Charitas Beach, Niterói, Rio de Janeiro, Brazil. J Coast Res 35: 573-579.
- DEFEO O AND MCLACHLAN A. 2005. Patterns, processes and regulatory mechanisms in sandy beach macrofauna: a multi-scale analysis. Mar Ecol Prog Ser 295: 1-20.
- DEFEO O AND MCLACHLAN A. 2011. Coupling between macrofauna community structure and beach type: a deconstructive meta-analysis. Mar Ecol Prog Ser 433: 29-41.
- DELGADO J, RIERA R, MONTERROSO Ó AND NÚÑEZ J. 2009. Distribution and abundance of meiofauna in intertidal sand substrata around Iceland. Aquat Ecol 43: 221-233.
- DI DOMENICO M, LANA PC AND GARRAFFONI ARS. 2008. Distribution patterns of interstitial polychaetes in sandy beaches of southern Brazil. Mar Ecol 30: 47-62.
- ESTEVES AM AND GENEVOIS V. 2006. Os nematódeos e sua importância nos ecossistemas marinhos. Floresta Ambient 13: 113-120.
- FRICKE AH, HENNIG H AND ORREN MJ. 1981. Relationship between oil pollution and psammolittoral meiofauna density of two South-african beaches. Mar Environ Res 5: 59-77.
- GHESKIERE T, HOSTE E, KOTWICKI L, DEGRAER S, VANAVERBEKE J AND VINCX M. 2002. The sandy beach meiofauna and free-living nematodes from De Panne (Belgium). Bull Inst R Sci Nat Belg 72: 43-49.
- GHESKIERE T, HOSTE E, VANAVERBEKE J, VINCX M AND DEGRAER S. 2004. Horizontal zonation patterns and feeding structure of marine nematode assemblages on a macrotidal, ultra-dissipative sandy beach (De Panne, Belgium). J Sea Res 52: 211-226.
- GHESKIERE T, VINCX M, GREET P AND DEGRAER S. 2006. Are strandline meiofaunal assemblages affected by a onceonly mechanical beach cleaning? Experimental findings. Mar Environ Res 61: 245-264.
- GHESKIERE T, VINCX M, URBAN-MALINGA B, ROSSANO C, SCAPINI F AND DEGRAER S. 2005a. Nematode from wave-dominated sandy beaches: diversity, zonation, patterns and

- testing iso-communities concept. Estuar Coast Shelf Si 62: 365-375
- GHESKIERE T, VINCX M, WESLAWSKI JM, SCAPINI F AND DEGRAER S. 2005b. Meiofauna as descriptor of tourism-induced changes at sandy beaches. Mar Environ Res 60: 245-265
- GIERE O. 1979. The impact of oil pollution on intertidal meiofauna. Field studies after the la Coruna-spill, May 1976. Cah Biol Mar 20: 231-251.
- GIERE O. 2009. Meiobenthology. The microscopic motile fauna of aquatic sediments, 2nd ed., Springer, Berlin Heidelberg, 527 p.
- GINGOLD R, IBARRA-OBANDO S AND ROCHA-OLIVARES A. 2011. Spatial aggregation patterns of free-living marine nematodes in contrasting sandy beach micro-habitats. J Mar Biol Assoc UK 91: 615-622.
- GINGOLD R, MOENS T AND ROCHA-OLIVARES A. 2013.
 Assessing the Response of Nematode Communities to Climate Change-Driven Warming: A Microcosm Experiment. PLoS ONE 8: e66653.
- GINGOLD R, MUNDO-OCAMPO M, HOLOVACHOV O AND ROCHA-OLIVARES A. 2010. The role of habitat heterogeneity in structuring the community of intertidal free-living marine nematodes. Mar Biol 157: 1741-1753.
- GOMES TP AND ROSA FILHO JS. 2009. Composição e variabilidade espaço-temporal da meiofauna de uma praia arenosa na região amazônica (Ajuruteua, Pará). Iheringia 99: 210-216.
- GOURBAULT N, WARWICK RM AND HELLÉOUET MN. 1995. A survey of intertidal meiobenthos (especially nematodes) in coral sandy beaches of Moorea (French Polynesia). B Mar Sci 57: 476-488.
- GRAY JS AND RIEGER RM. 1971. A quantitative study of the meiofauna of an exposed sandy beach, at Robin Hood's Bay, Yorkshire. J Mar Biol Assoc UK 51: 1-19.
- GRZELAK K, KOTWICKI L AND SZCZUCINSKI W. 2009. Monitoring of Sandy Beach Meiofaunal Assemblages and Sediments after the 2004 Tsunami in Thailand. Pol J Environ Stud 18: 43-51.
- HARRIAGUE AC, GAOZZA L, MONTELLA A AND MISIC C. 2006. Benthic communities on a sandy Ligurian beach (NW Mediterranean). Hydrobiologia 571: 383-394.
- HARRIAGUE AC, PETRILLO M, DELLA CROCE N, PANCIROLI H, PUSCEDDU A, PELOSI G, DANOVARO R AND ALBERTELI G. 2008. Intertidal benthic communities of two Chilean coastal islands (Santa Maria and Mocha, Southeastern Pacific). Chem Ecol 24: 119-128.
- HARRIS RP. 1972. The distribution and ecology of the interstitial meiofauna of a sandy beach at Whitsand Bay, East Cornwall. J Mar Biol Assoc UK 52: 1-18.
- HEIP C, VINCX M AND VRANKEN G. 1985. The ecology of Marine Nematodes. Oceanogr Mar Biol 23: 399-489.

- HEYMANS JJ AND MCLACHLAN A. 1996. Carbon budget and network analysis of a high energy beach/surf-zone ecosystem. Estuar Coast Shelf S 43: 485-505.
- HOURSTON M, WARWICK RM, VALESINI FJ AND POTTER IC. 2005. To what extent are the characteristics of nematode assemblages in nearshore sediments on the west Australian coast related to habitat type, season and zone? Estuar Coast Shelf S 64: 601-612.
- HULINGS NC AND GRAY JS. 1976. Physical Factors Controlling Abundance of Meiofauna on Tidal and Atidal Beaches. Mar Biol 74: 77-83.
- JOINT IR, GEE JM AND WARWICK RM. 1982. Determination of Fine-Scale Vertical Distribution of Microbes and Meiofauna in an Intertidal Sediment. Mar Biol 72: 157-164
- KANG T, MIN W, RHO HS, PARK H AND KIM D. 2014. Differential responses of a benthic meiofaunal community to an artificial oil spill in the intertidal zone. J Mar Biol Assoc UK 94: 219-231.
- KONT A, JAAGUS J AND AUNAP R. 2003. Climate change scenarios and the effect of sea-level rise for Estonia. Glob Planet Chang 36: 1-15.
- KOTWICKI L, DE TROCH M, URBAN-MALIGNA B, GHESKIERE T AND WESLAWSKI JM. 2005a. Horizontal and vertical distribution of meiofauna on sandy beaches of the North Sea (The Netherlands, Belgium, France). Helgoland Mar Res 59: 255-264.
- KOTWICKI L, DEIDUN A, GRZELAK K AND GIANNI F. 2014. A preliminary comparative assessment of the meiofaunal communities of Maltese pocket sandy beaches. Estuar Coast Shelf Si 150: 111-119.
- KOTWICKI L, SZYMELFENIG M, DE TROCH M, URBAN-MALINGA B AND WESLAWSKI J. 2005b. Latitudinal biodiversity patterns of meiofauna from sandy littoral beaches. Biodivers Conserv 14: 461-474.
- LEE MR AND CORREA JA. 2005. Effects of copper mine tailings disposal on littoral meiofaunal assemblages in the Atacama region of northern Chile. Mar Environ Res 59: 1-18.
- LEE MR AND RIVEROS M. 2012. Latitudinal trends in the species richness of free-living marine nematode assemblages from exposed sandy beaches along the coast of Chile (18-42°S). Mar Ecol Evol Persp 33: 317-325.
- LIU HB, ZHANG ZN, FAN SL, HUA E AND DENG K. 2008. Seasonal variability in free-living marine nematode community structure in a sandy beach of the Taiping Bay of Qingdao, China. Acta Oceanol Sin 27: 102-115.
- LOCK K, MEES J, VINCX M AND GOETHALS PLM. 2011. Did global warming and alien invasions affect surf zone hyperbenthic communities on sandy beaches in Belgium? Hydrobiologia 664: 173-181.
- LONG SM AND ROSS OBH. 1999. Vertical distribution of nematodes (Nematoda) and harpacticoid copepods

- (Copepoda: Harpacticoidea) in muddy and sandy bottom of intertidal zone at LokSawi, Sabah, Malaysia. Raffles B Zool 43: 349-363.
- MANTHA G, MOORTHY MSN, ALTAFF K, DAHMS HU, LEE WO, SIVAKUMAR K AND HWANG JS. 2012. Seasonal shifts of meiofauna community structures on sandy beaches along Chennai Coast, India. Crustaceana 85: 27-53.
- MARIA TF, DE TROCH M, VANAVERBEKE J, ESTEVES AM AND VANREUSEL A. 2011a. Use of benthic vs planktonic organic matter by sandy-beach organisms: A food tracing experiment with ¹³C labelled diatoms. J Exp Mar Biol Ecol 407: 309-314.
- MARIA TF, DE TROCH M, VANAVERBEKE J, ESTEVES AM AND VANREUSEL A. 2012. The importance of biological interactions for the vertical distribution of nematodes in a temperate ultra-dissipative sandy beach. Estuar Coast Shelf S 97: 114-126.
- MARIA TF, ESTEVES AM, SMOL N, VANREUSEL A AND DECRAEMER W. 2008. Nematodes from sandy beaches of Guanabara Bay, Rio de Janeiro, Brazil. Biociências 16: 92-103.
- MARIA TF, ESTEVES AM, VANAVERBEKE J AND VANREUSEL A. 2011b. The effect of the dominant polychaete *Scolelepis squamata*on nematode colonisation in sandy beach sediments: An experimental approach. Estuar Coast Shelf S 94: 272-280.
- MARIA TF, ESTEVES AM, VANAVERBEKE J AND VANREUSEL A. 2013a. Is nematode colonisation in the presence of *Scolelepis* in tropical sandy-beach sediment similar to the colonisation process in temperate sandy beaches? Braz J Biol 73: 19-28.
- MARIA TF, GINGOLD R, VANAVERBEKE J, ESTEVES AM AND VANREUSEL A. 2013b. Tidal exposure or microhabitats: what determines sandy-beach nematode zonation? A case study of a macrotidal ridge-and-runnel sandy beach in Belgium. Mar Ecol Evol Persp 34: 207-217.
- MARIA TF, PAIVA P, VANREUSEL A AND ESTEVES AM. 2013c. The relationship between sandy beach nematodes and environmental characteristics in two Brazilian sandy beaches (Guanabara Bay, Rio de Janeiro). An Acad Bras Cienc 85: 257-270.
- MASSELINK G, KROON A AND DAVIDSON-ARNOTT RGD. 2006. Morphodynamics of intertidal bars in wave-dominated coastal settings A review. Geomorphology 73: 33-49.
- MCGLONE D AND VUILLE M. 2012. The associations between El Niño-Southern Oscillation and tropical South American climate in a regional climate model. J Geophysic R 117: 1-15.
- MCLACHLAN A. 1978. A quantitative analysis of the meiofauna and the chemistry of the redox potential discontinuity zone in a sheltered sandy beach. Estuar Coast Shelf S 7: 275-290.

- MCLACHLAN A. 1980. The Definition of Sandy Beaches in Relation to Exposure: A Simple Rating System. S Afr J Sci 76: 137-138.
- MCLACHLAN A. 1983. Sandy beach ecology a review. In: McLachlan A and Anderasmus T (Eds), Sandy beaches as Ecosystems. Dr.W. Junk, p. 321-380.
- MCLACHLAN A AND BROWN A. 2006. The ecology of sandy shores. Elsevier, USA, 373 p.
- MCLACHLAN A AND HARTY B. 1982. Effects of crude-oil on the supralittoral meiofauna of a sandy beach. Mar Environ Res 7: 71-79.
- MCLACHLAN A AND JARAMILLO E. 1995. Zonation of sandy beaches. Oceanogr Mar Biol 33: 305-335.
- MCLACHLAN A, WINTER PED AND BOTHA L. 1977. Vertical and Horizontal Distribution of Sub-Littoral Meiofauna in Algoa Bay, South Africa. Mar Biol 40: 355-364.
- MEAD A ET AL. 2013. Human-mediated drivers of change impacts on coastal ecosystems and marine biota of South Africa. Afr J Mar Sci 35: 403-425.
- MENN I. 2002. Beach morphology and food web structure: comparison of an eroding and accreting sandy shore in the North Sea. Helgol Mar Res 56: 177-189.
- MOELLMANN AM AND CORBISIER TN. 2003. Does tourist flow affect the meiofauna of sandy beaches? Preliminary results. J Coast Res 35: 590-598.
- MORENO M, FERRERO TJ, GRANELLI V, MARIN V, ALBERTELLI G AND FABIANO M. 2006. Across shore variability and trophodynamic features of meiofauna in a microtidal beach of the NW Mediterranean. Estuar Coast Shelf Sci 66: 357-367.
- MORTENSENT. 1921. Biologisk Studier over Sandstrandfaunen, saerlig ved de danske Kyster. Vidensk Medd Dan Nat Hist Foren 74: 23-56.
- MUNDO-OCAMPO M ET AL. 2007. Biodiversity of littoral nematodes from two sites in the Gulf of California. Hydrobiologia 586: 179-189.
- MUNRO ALS, WELL JBL AND MCINTYRE AD. 1978. Energy flow in the flora and meiofauna of sandy beaches. P Roy Soc Edinb B 76: 297-315.
- NANAJKAR M AND INGOLE B. 2010. Impact of sewage disposal on a nematode community of a tropical sandy beach. J Environ Biol 31: 816-826.
- NETTO SA AND MENEGHEL A. 2014. Pulse of marine subsidies: the role of surf diatom *Asterionellopsis glacialis* accumulations in structuring the meiofauna of sandy beaches. Mar Biodiv 44(3): 455-457.
- NICHOLAS WL. 2001. Seasonal variations in nematode assemblages on an Australian temperate ocean beach; the effect of heavy seas and unusually high tides. Hydrobiologia 464: 17-26.
- NICHOLAS WL. 2006. The meiofauna of a New South Wales sandy beach: an introduction to the meiofauna of Australian ocean sandy beaches. Wetlands 23: 14-31.

- NICHOLAS WL AND HODDA M. 1999. The free-living nematodes of a temperate, high energy, sandy beach: faunal composition and variation over space and time. Hydrobiologia 394: 113-127.
- NICHOLAS WL AND TRUEMAN JWH. 2005. Biodiversity of marine nematodes in Australian sandy beaches from tropical and temperate regions. Biodivers Conserv 14: 823-839.
- ÓLAFSSON E. 1991.Intertidal meiofauna of four sandy beaches in Iceland. Ophelia 33: 55-65.
- OTT JA. 1972. Determination of Fauna Boundaries of Nematodes in an Intertidal Sand Flat. Int Rev Gesamten Hydrobiol 57: 645-663.
- PAPAGEORGIOU N, MORENO M, MARIN V, BAIARDO S, AR-VANITIDIS C, FABIANO M AND ELEFTHERIOU A. 2007. Interrelationships of bacteria, meiofauna and macrofauna in a Mediterranean sedimentary beach (Maremma Park, NW Italy). Helgoland Mar Res 61: 31-42.
- PATTNAIK A AND RAO MVL. 1990. Composition and distribution of interstitial meiofauna of the sandy beach at Gopalpur, South Orissa Coast. Indian J Mar Sci 19: 165-170.
- PLATT HM. 1977. Vertical and horizontal distribution of freeliving marine nematodes from Strangford Lough, Nothern Ireland. Cah Biol Mar 18: 261-273.
- POLLOCK LW AND HUMMON WD. 1971. Cyclic changes in interstitial water content, atmospheric exposure, and temperature in a marine beach. Limnol Oceanogr 16: 522-535.
- REMANE A. 1933. Verteilung und Organisation der benthonischen Mikrofauna der Kieler Bucht Wiss Meer Abt Kiel 21: 161-221.
- RIERA R, NÚÑEZ J, BRITO MC AND TUYA F. 2012. Differences in diversity, structure, and variability between intertidal and subtidal meiofaunal assemblages. Cienc Mar 38: 677-693.
- RIERA R, NÚÑEZ J, DEL CARMEN BRITO M AND TUYA F. 2011. Temporal variability of a subtropical intertidal meiofaunal assemblage: contrasting effects at the species and assemblage-level. Vie Milieu 61: 129-137.
- RODRIGUEZ JG. 2004. Community structure of intertidal meiofauna along a gradient of morphodynamic state on exposed North Sea beach. Sarsia 89: 22-32.
- RODRÍGUEZ JG, INCERA M AND LÓPEZ J. 2008. Ecophysiological response of meiofauna to physicochemical gradients in intertidal sandy sediments. Mar Ecol 29: 60-69.
- RODRÍGUEZ JG, LASTRA M AND LÓPEZ J. 2003. Meiofauna distribution along a gradient of sandy beaches in northern Spain. Estuar Coast Shelf S 58: 63-69.
- RODRIGUEZ JG, LÓPEZ J AND JARAMILLO E. 2001. Community structure of the intertidal meiofauna along a gradient of morphodynamic sandy beach types in southern Chile. Rev Chil Hist Nat 74: 885-897.
- SALVAT B. 1964. Les conditions hydrodynamiques interstitielles des sediment meubles intertidaux et larepartition verticale de la jemmeendogee. Cr Hebd Acad Sci 259: 1576-1579.

- SALVAT B. 1967. La macrofaune carinologique endogue des sediments meubles intertidaux (Tanadaices, Isopodes et Amphipodes): ethologie, binomie et cycle biologique. Mem Mus Hist Nat 45: 1-275.
- SCHLACHER TA AND HARTWIG J. 2013. Bottom-up control in the benthos of ocean-exposed sandy beaches? Austral Ecol 38: 177-189.
- SCHLACHER TA, SCHOEMAN D, DUGAN J, LASTRA M, JONES A, SCAPINI F AND MCLACHLAN A. 2008. Sandy beach ecosystems: key features, sampling issues, management challenges and climate changes impacts. Mar Ecol 29(S): 70-89.
- SHARMA J AND WEBSTER JM. 1983. The abundance and distribution of free-living nematodes from two Canadian Pacific beaches. Estuar Coast Shelf S 16: 217-227.
- SOUZA-SANTOS LP, RIBEIRO VSS, SANTOS PJP AND FONSECA-GENEVOIS V. 2003. Seasonality of intertidal meiofauna on a tropical sandy beach in Tamandaré Bay (Northeast Brazil). J Coastal Res 35: 369-377.
- SPEYBROECK J ET AL. 2008. The Belgian sandy beach ecosystem: a review. Mar Ecol 29: 171-185.
- STEYAERT M, HERMAN PMJ, MOENS T, WIDDOWS J AND VINCX M. 2001. Tidal migration of nematodes on an estuarine tidal flat (the Molenplaat, Schelde Estuary, SW Netherlands). Mar Ecol Progr S 224: 299-304.
- STEYAERT M, VANAVERBEKE J, VANREUSEL A, BARRANGUET C, LUCAS C AND VINCX M. 2003. The importance of fine-scale, vertical profiles in characterising nematode community structure. Estuar Coast Shelf S 58: 353-366.
- SUN X, ZHOU H, HUA E, XU S, CONG B AND ZHANG Z. 2014. Meiofauna and its sedimentary environment as an integrated indication of anthropogenic disturbance to sandy beach ecosystems. Mar Pollut Bul 88: 260-267.
- SZYMELFENIG M, KWASNIEWSKI S AND WESLAWSKI JM. 1995. Intertidal zone of Svalbard II. Meiobenthos density and occurrence. Polar Biol 15: 137-141.
- URBAN-MALINGA B AND BURSKA D. 2009. The colonization of macroalgal wrack by the meiofauna in the Arctic intertidal. Estuar Coast Shelf S 85: 666-670.
- URBAN-MALINGA B, GHESKIERE T, DEGRAER S, DERYCKE S, OPALINSKI KW AND MOENS T. 2008. Gradients in biodiversity and macroalgal wrack decomposition rate across a macrotidal, ultradissipative sandy beach. Mar Biol 155: 79-90.
- URBAN-MALINGA B, KOTWICKI L, GHESKIERE T, JANKOWSKA K, OPALIÑSKI K AND MALINGA M. 2004. Composition and distribution of meiofauna, including nematode genera, in two contrasting Arctic beaches. Polar Biol 27: 447-457.
- URBAN-MALINGA B AND MOENS T. 2006. Fate of organic matter in Arctic intertidal sediments: Is utilisation by meiofauna important? J Sea Res 56: 239-248.

- URBAN-MALINGA B, WIKTOR J, JABLONSKA A AND MOENS T. 2005. Intertidal meiofauna of a high-latitude glacial Arctic fjord (Kongsfjorden, Svalbard) with emphasis on the structure of free-living nematode communities. Polar Biol 28: 940-950.
- URISH DW AND MCKENNA TE. 2004. Tidal effects on ground water discharge through a sandy marine beach. Ground Water 42: 971-982.
- VEIGA P, BESTEIRO C AND RUBAL M. 2010a. Meiofauna communities in exposed sandy beaches on the Galician coast (NW Spain), six months after the *Prestige* oil spill: the role of polycyclic aromatic hydrocarbons (PAHs). Sci Mar 74: 385-394.
- VEIGA P, BESTEIRO C AND RUBAL M. 2010b. The role of sediment type on the vertical distribution of meiofauna at two Galician rias (NW Iberian Peninsula). Cah Biol Mar 51: 249-263.

- VENEKEY V, DOS SANTOS PJP AND FONSECA-GENEVOIS VG. 2014a. The influence of tidal and rainfall cycles o intertidal nematodes: a case study in a tropical sandy beach. Braz J Oceanogr 62: 247-256.
- VENEKEY V, DOS SANTOS PJP AND FONSECA-GENEVOIS VG. 2014b. Effect of environmental factors on intertidal nematode in a tropical sandy beach (Tamandaré, Pernambuco, Brazil). J Coastal Res 30: 785-794.
- WARWICK RM. 1988. The level of taxonomic discrimination required to detect pollution effects on marine benthic communities. Mar Pollut Bull 19: 259-268.
- WORMALD AP. 1976. Effects of a spill of marine diesel oil on meiofauna of a sandy beach at Picnic Bay, Hong Kong. Environ Pollut 11: 117-130.
- YAMANAKA T, RAFFAELLI D AND WHITE PCL. 2010. Physical determinants of intertidal communities on dissipative beaches: Implications of sea-level rise. Estuar Coast Shelf S 88: 267-278.