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**Optimum harvesting scenarios for the management of the bigeye tuna *Thunnus obesus* at the Eastern Pacific Ocean**

Escenarios óptimos de manejo para el atún patudo *Thunnus obesus* en el Océano Pacífico oriental

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**INTRODUCTION**

In the Eastern tropical Pacific Ocean (EPO), tuna and other highly migratory fish are caught by various fishing gears, from long lines to purse-seine (Lennert-Cody *et al.* 2007, FSR 2010). This area of the ocean contributes with about 14% of the world tuna production (ISSF 2010). Among its key commercial species are yellowfin (*Thunnus albacares*), skipjack (*Katsuwonus pelamis*), bigeye (*Thunnus obesus*) and albacore (*Thunnus alalunga*) (FSR 2010). The bigeye is amongst the most important tuna species landed. Bigeye tuna are distributed throughout the Pacific Ocean and it is believed there is only a single stock with a local exchange of individuals (FSR 2010). Numerous stock assessments of bigeye have been made from analysis of old data, capture, length, etc. (Maunder...
& Watters 2003, Harley & Maunder 2005, Maunder & Hoyle 2006) published in reports of the Interamerican Tropical Tuna Commission, IATTC (Maunder & Harley 2007), indicating that the bigeye stock in the Eastern Pacific was over-exploited (IATTC 2008). However, according to the most recent stock assessment conducted in 2009 (Aires-da Silva & Maunder 2010), fishing mortality rates are estimated to be below the level corresponding to the Maximum Sustainable Yield (MSY), and the recent levels of spawning biomass are estimated to be above that level (IATTC 2010).

For the management of this fishery, several international agreements have been established e.g., United Nations Convention on the Law of the Sea and the United Nations Fish Stocks, whose main objective is to maintain populations at levels of MSY. However, this goal has the problem of no taking into account the economic aspects or the social ones (Maunder & Harley 2007). Therefore, the aim of this paper is to evaluate the fishery and simulate different management scenarios taking into account not just the MSY, but also the Maximum Economic Yield (MEY), and the social aspect in terms of the dependence of the number of direct jobs from changes in $F$, including a brief analysis of the cost of fishing and the elasticity of prices and demand.

**MATERIALS AND METHODS**

The assessment of the stock was made using 15 years of catch data of the bigeye tuna at the EPO (FSR 2010). The population parameter values were taken from published sources and are indicated in Table 1. Changes in abundance over time were determined using the catch data in metric tons (mt). Trends in fishing mortality ($F$) over time and the estimates of the stock biomass were examined. The criteria used for evaluation of fishing scenarios were based on the $F$ and the age of first catch ($t_c$). Other reference point examined was the $F$ at the maximum-economic-yield ($F_{MEY}$) and the $F$ at the maximum sustainable yield ($F_{MSY}$). The number of direct jobs is a consequence of the former 2 variables and therefore it could not be set as a target in simulations, it is a variable directly depending on the fishing intensity and its maximum value is found at the highest $F$ used as input.

**Table 1. Population parameter values used for the evaluation of the big eye tuna exploited in the EPO (FSR 2010).** $K$, $t_c$, $L_\infty$ and $W_\infty$ are from the von Bertalanffy growth model. $a$ and $b$ are from the length-weight allometric equation; $tc$ = age of first catch; $M$ = Natural mortality; $tm$ = Maturity age; $t_\lambda$ = Longevity and $E_{max}$ = Exploitation rate at the MSY level. The $W_\infty$ value was obtained by using the length - weight equation, as $W = 0.0178L^{2.902}$ / Parámetros de la población usados para la evaluación del atún patudo explotado en el OPO (FSR 2010). $K$, $t_c$, $L_\infty$ y $W_\infty$ fueron obtenidos del modelo de crecimiento de Von Bertalanffy. $a$ y $b$ de la relación longitud-peso; $tc$ = edad de primera captura; $M$ = mortalidad natural; $tm$ = edad de madurez; $t_\lambda$ = longevidad y $E_{max}$ = tasa de explotación al nivel de RMS. El valor de $W_\infty$ fue obtenido usando la relación peso - longitud, como $W = 0,0178L^{2.902}$.

<table>
<thead>
<tr>
<th>Parameter</th>
<th>Value</th>
<th>Units-Model</th>
<th>Source</th>
</tr>
</thead>
<tbody>
<tr>
<td>$K$</td>
<td>0.23</td>
<td>Bertalanffy</td>
<td>Zhu et al. (2009)</td>
</tr>
<tr>
<td>$L_\infty$</td>
<td>207.4</td>
<td>cm</td>
<td>Zhu et al. (2009)</td>
</tr>
<tr>
<td>$W_\infty$</td>
<td>94,145</td>
<td>Weight, g</td>
<td>Nakamura &amp; Uchirama (1996)</td>
</tr>
<tr>
<td>$t_c$</td>
<td>-0.4</td>
<td>Years-Bertalanffy</td>
<td>Zhu et al. (2009)</td>
</tr>
<tr>
<td>$a$</td>
<td>0.0178</td>
<td>Length-weight</td>
<td>Nakamura &amp;Uchirama (1996)</td>
</tr>
<tr>
<td>$b$</td>
<td>2.92</td>
<td>Length-weight</td>
<td>Nakamura &amp;Uchirama (1996)</td>
</tr>
<tr>
<td>$tc$</td>
<td>4</td>
<td>Years</td>
<td>Compean (2010)</td>
</tr>
<tr>
<td>$tm$</td>
<td>5</td>
<td>Years</td>
<td>FISMO</td>
</tr>
<tr>
<td>$t_\lambda$</td>
<td>13</td>
<td>Years</td>
<td>Pauly (1983)</td>
</tr>
<tr>
<td>$\beta$</td>
<td>2.25</td>
<td></td>
<td>Beverton &amp; Holt (1957)</td>
</tr>
<tr>
<td>$M$</td>
<td>0.345</td>
<td>Instantaneous rate</td>
<td>Jensen (1996, 1997)</td>
</tr>
<tr>
<td>$E_{max}$</td>
<td>0.25</td>
<td>$F_{MSY}(M+ F_{MSY})$</td>
<td>This paper</td>
</tr>
<tr>
<td>$phi^*$</td>
<td>4</td>
<td>$Log K + 2Log L$</td>
<td>Vakily (1992); Sparre &amp;Venema (1997)</td>
</tr>
<tr>
<td>Value/Kg</td>
<td>1.5</td>
<td>USD</td>
<td>FIS (2013)*</td>
</tr>
<tr>
<td>Cost/day/trip</td>
<td>879</td>
<td>USD</td>
<td>Mora (1997)</td>
</tr>
<tr>
<td>Cost/Catch</td>
<td>755</td>
<td>USD/mt</td>
<td>This paper</td>
</tr>
</tbody>
</table>

The population parameters plus the catch data were analyzed with the aid of a simulation model, implemented in the age-structured simulation model FISMO (Chávez 2005). The age of first capture was 3 years and was maintained constant in the fitting process, but for the simulation all the $F$ and $tc$ values feasible to apply were tested in the optimizing process and for designing exploitation policies.

Additionally, the maximum social value was determined in 2 ways. The first one is the maximum level of employment or in other words, the maximum number of fishermen. The second way of the approach consists in evaluating the maximum profit per fisherman. Economic and social values were the value per kg landed, the number of fishermen per vessel, the number of vessels and the number of fishing days, all from the last fishing season and later estimated for each year of catch data. Costs were obtained by adding the cost/vessel/day times the total number of fishing days of the fleet over the fishing season. Profits were substracted from the total value of the catch minus the total costs. Costs and value were linked to the catch in the FISMO simulation model (Chávez 1996, 2005), which allowed testing all of the possible scenarios of exploitation. The model is based on general principles of fish stock assessment (Hilborn & Walters 1992).

Population parameter values were known and were taken from the literature (Zhu et al. 2009, Nakamura & Uchiyama 1966, Table 1), and estimates of the age composition of the catch from the FAO records. In each case total mortality ($Z$) was determined with the exponential decay model as

$$N_{a+1} = N_a e^{-\lambda t}$$

where, $N_{a+1}$ is the number of bigeye fish of age $a+1$ and $N_a$ is the number of bigeye fish of age $a$ in reconstructed age-groups. With the use of the von Bertalanffy growth equation describing growth as a function of age, it was possible to determine the numbers per age, using the catch as reference, allowing the determination to their corresponding individual lengths and weights. These lengths were transformed into their respective weights by using the equation

$$W = a L^b$$

where $W$ = total weight (g) and $L$ = total length (cm).

The time units were years. The age structures for each year were estimated assuming a constant natural mortality. For setting the variables of the initial state, the abundance per age class ($N_a$) was defined using the age-specific abundance $N_a/\Sigma N_a$ obtained from first equation. In subsequent years, the age structure was defined after the estimation of the number of one-year-old recruits. These values were used to calculate catch-at-age as proposed by Sparre & Venema (1997) and were integrated into the FISMO simulation model (Chávez 2005) as:

$$Y_{a,y} = N_{a,y} W_{a,y} \frac{F_t}{(F_t + M)} (1 - e^{-(F_t-M)})$$

where, $Y_{a,y}$ is the catch-at-age $a$ of each year $y$, $N_{a,y}$ is the number of bigeye tunas at age $a$ in year $y$, $W_{a,y}$ is the fish weight equivalent to $N_{a,y}$, $F_t$ is described, and $M$ is the natural mortality. Given the established initial conditions, the values of $Y_{a,y}$ were adjusted by varying the initial number of recruits and linked to the equations described above until the condition of the following equation was fulfilled:

$$\sum_a Y_{a,y} = Y_{y(REC)}$$

where $Y_{y(REC)}$ is the yield recorded during the year $y$, $a = 2$ years, $K$ is the growth constant of the von Bertalanffy growth equation, and $t = 3/K$ or longevity ($= 25$ years). This value was found by assuming that a reasonable life expectancy ($L_{max}$) is when 95% of the population reaches 95% of $L_{max}$, the asymptotic length (Chávez 1995). The longevity was found by making $L_{max} = 0.95L_{\infty}$ in the von Bertalanffy growth equation and finding the respective value of $t$ in the equation. Use of the catch equation was made for each year in the time-series analyzed. For the estimation of the natural mortality ($M$), the criterion proposed by Jensen (1996, 1997) was adopted, where $M = 1.5*K = 0.1793$ (see Table 1); the approach by Gislason et al. (2010) is not too different. Estimations of the stock biomass and the exploitation rate $E = [F/(M+F)]$ were made for each age-class in every fishing year analyzed by the model. These values were compared to the $E$ value at the $F_{MSY}$ level, corresponding to the maximum exploitation rate that a fishery attains before the stock is over exploited. A diagnosis of which years of the series the stock was under or over exploited was then made, providing an easy way to recommend either a further increase or decrease of $F$.

The annual cohort abundance ($N_{a,y}$) coming from ages older than age-at-maturity ($tm = 5$ years) were used to estimate the annual abundance of adults ($S$) over the years, whereas the abundance of the one-year-old group
was used as the number of recruits \((R)\). The stock-recruitment relationship was evaluated by using a slightly modified version of the Beverton & Holt (1957) model in the form:

\[
R_{y+1} = \frac{\alpha S_o S_y}{S_y + \beta S_o}
\]

where \(R_{y+1}\) is the number of one-year-old recruits in year \(y+1\), \(S_y\) is the number of adults in year \(y\), \(S_o\) is the maximum number of adults in the population, and \(\alpha\) and \(\beta\) are: parameters modified from the original model where \(\alpha\) is the maximum number of recruits and \(\beta\) is the initial slope of the recruitment line, which was constant through the simulation. Estimates of economic data were made after the report of a commercial trip for 1997 (Mora 1997). The values of the parameters used as input are shown in Table 1.

The bio-economic analysis of the fishery was carried out by following the simplistic approach of assigning a mean cost to each fishing day by a mean-sized vessel, and an account of the days of a typical fishing season, multiplied by the number of vessels; with this information it was possible to have an estimate of total costs. The number of direct jobs, the fishermen, was determined by multiplying the number of fishermen per vessel and then multiplying this number by the total number of vessels. The catch value is usually obtained from statistics, associated to current price of the fish landed and the profits are known by subtracting total value minus total costs in a given fishing season. To relate the costs with \(F\), it was necessary to estimate a catchability coefficient based on the \(F\) and the fishing effort \((f)\) applied to the last fishing season, when this data were obtained, allowing to estimate \(f\) for the 15 years of catch data and to simulate costs and profits for the 30 following years. This way, to a certain \(F\) value, there is a number of fishing days, costs, profits, vessels, and a number of fishermen associated, such that in a given fishing season,

- Total costs \(= Y_{av} \times \text{cost per fishing day}\)
- Total profits \(= (Y_{av} \times \text{value per kg}) - \text{Total Costs}\)
- Number of vessels \(= F/(qf)\)
- Number of fishermen = Number of vessels \(\times\) Number of fishermen per vessel
- Number of fishing days = Number of vessels \(\times\) Number of fishing days in the season

The bio-economic model was developed in Excel, where with the aid of the some macros, it was possible to obtain a wide number of output variables that allowed to evaluate many numerical and graphical harvesting scenarios, where the most important ones are the trend lines of catch, profits, stock biomass, number of direct jobs (the fishermen), the number of vessels, and the benefit/cost ratio. In most cases, for a given \(tc\), the graphic output describes asymmetrical curve lines, where there is a maximum value at a given \(F\); from these outputs, it is particularly important to mention that the Yield and the profits attain a maximum value at a certain \(F\) value, corresponding to the maximum yield at \(F_{\text{MSY}}\) or maximum sustainable yield, and the maximum profits at \(F_{\text{MEY}}\) or the maximum economic yield. In most cases, the last one \((F_{\text{MEY}})\) is attained at an \(F\) value that is lower than the one required for the \(F_{\text{MSY}}\). In some cases, particularly in valuable fisheries, the \(F_{\text{MEY}}\) is at a higher value than the \(F_{\text{MSY}}\) or at the same \(F\) value, but never higher.

The bioeconomic analysis of the fishery was made as far as the landing time, not afterwards, so no added value was considered. Cost of catch was determined simply as a division of total costs, divide by total catch, both in a given time. Correspondence between \(f\) and \(F\) was made, to estimate the cost per fishing day in reference to the last year of catch records.

Elasticity of prices and demand was taken into account by adding 5 and 10% to current fishing cost, and adding the same proportions to catch value; impact of these inputs was determined on total profits, on the cost of catch, and on the Benefit/Cost ratio.

Results of simulation are based only in the long-term output, not on the transition period immediately after the historical records; it was done this way because in cases of over exploitation of long-lived species, restoration of stocks may take more than 10 years. For this reason, any exploitation scenario indicated in the section of results should consider a gradual application of any changes of \(F\), not shown here, to avoid unexpected results inconsistent with the simulated outputs.

**Results**

From the generalized simulation model of a fishery (FISMO), age structure and catch for the last 15 years of data (1995-2009) were reconstructed. The catches range between 25% and 40% of the stock biomass (Fig. 1), which shows that the stock is under-exploited and close to the MSY level. The rate of exploitation \((F/Z)\) also confirms this trend. The last years of the simulation reveal a relative stability in the stock biomass with some fluctuations at the beginning of the time series (Fig. 1). The trend of \(F\)
and the exploitation rate in the 15 years of the analysis (Fig. 1b) indicates that the population is under exploited, i.e., in the case of $F$, in only 3 years the values were exceeded, particularly during the 2000 season; in the seasons 1996 and 1997, a moderate over-exploitation was observed (should not be $F > 0.29$). The trend of the exploitation rate values ($E$) are more consistent and only in 2000 exceed the limit of $E = 0.46$. The profits of this activity (Fig. 1c) indicate that this is a very productive fishery, for even considering the bigeye as the target species, the fleet generates profits that in half of the study period are over 40 M USD, with a benefit/cost ratio (B/C) > 1.9, indicating that it is a healthy economic activity.

The simulation of different management scenarios proposed ($MSY$, $MEY$ and direct jobs) to ensure a full biological and economic exploitation, shows that the fishery ($F = 0.249$) shows higher profits at the current state than at the $MSY$ level (Table 2). The $MSY$ is attained at $F = 0.29$ and the $MEY$ at $F = 0.16$ maintaining the same age of first catch of 4 years in all cases (Fig. 2). To achieve the level of $MEY$, the $F$ should be reduced from the current value of 0.249 to 0.16. In principle, this would be the best management scenario because it would generate the highest profits of the fishery and the highest income per fisherman; it would imply a reduction of the fishing effort by about 30% which would result in a reduction of the catch in nearly 8,000 mt (Table 2).

The elasticity of prices and value were examined by adding 5 and 10% to catch value and to the cost of fishing, as was stated before. The effect of these changes was evaluated in the profits (Fig. 3a), in the cost of catch (Fig. 3b), and in the B/C ratio (Fig. 3c), as a function of $F$. In the 3 cases the 5 and 10% additions are represented as lines thinner than the current trend, which is shown as a thicker line in each case.

### Table 2. Management scenarios for the big eye tuna ($Thunnus obesus$) fishery at the EPO. The age of first catch is 4 years, the same as in the fishery / Escenarios de manejo para la pesquería del atún patudo ($Thunnus obesus$) en el OPO. La edad de primera captura es 4 años, igual que en la pesquería

<table>
<thead>
<tr>
<th>Indicators</th>
<th>Current</th>
<th>$F_{MEY}$</th>
<th>$F_{MSY}$</th>
</tr>
</thead>
<tbody>
<tr>
<td>Stock biomass, mt</td>
<td>736,130</td>
<td>675,546</td>
<td>903,724</td>
</tr>
<tr>
<td>$F$ (yr)</td>
<td>0.249</td>
<td>0.29</td>
<td>0.16</td>
</tr>
<tr>
<td>Age first catch</td>
<td>4</td>
<td>4</td>
<td>4</td>
</tr>
<tr>
<td>Yield (mt)</td>
<td>77,605</td>
<td>78,206</td>
<td>69,870</td>
</tr>
<tr>
<td>Fishing days/yr</td>
<td>66,085</td>
<td>77,665</td>
<td>42,850</td>
</tr>
<tr>
<td>Days/vessel/yr</td>
<td>304</td>
<td>304</td>
<td>304</td>
</tr>
<tr>
<td>Vessels</td>
<td>219</td>
<td>274</td>
<td>219</td>
</tr>
<tr>
<td>Direct jobs</td>
<td>3,517</td>
<td>3,572</td>
<td>2,852</td>
</tr>
<tr>
<td>Catch value</td>
<td>116,407,500</td>
<td>117,309,279</td>
<td>104,805,550</td>
</tr>
<tr>
<td>Total Costs</td>
<td>58,616,233</td>
<td>68,267,902</td>
<td>37,665,049</td>
</tr>
<tr>
<td>Costs/Catch, USD/mt</td>
<td>755</td>
<td>873</td>
<td>539</td>
</tr>
<tr>
<td>B/C</td>
<td>1.99</td>
<td>1.72</td>
<td>2.78</td>
</tr>
<tr>
<td>Profits</td>
<td>57,791,267</td>
<td>49,041,377</td>
<td>67,140,501</td>
</tr>
<tr>
<td>Profits/vessel</td>
<td>213,907</td>
<td>178,740</td>
<td>306,578</td>
</tr>
<tr>
<td>Profits/Fisherman</td>
<td>16,431</td>
<td>13,730</td>
<td>23,542</td>
</tr>
</tbody>
</table>

Figure 1. a) Estimated stock biomass (area, right scale) and catch (solid line, left scale), from data after the IATTC; b) exploitation rate ($E$) and fishing mortality ($F$) of the Thunnus obesus fishery; c) Profits (bars, left scale, USD) and Benefit/Cost ratio (line, right scale) of the big eye tuna fishery at the EPO, estimated after the simulation model / a) Biomasa (área, eje derecho) y captura (línea, eje izquierdo) estimados a partir de datos de captura de la IATTC; b) Tasa de explotación ($E$) y mortalidad por pesca ($F$) en la pesquería de Thunnus obesus; c) Utilidades (barras, eje izquierdo, Dólares) y tasa Beneficio/Costo (línea, eje derecho) de la pesquería del atún patudo en el OPO, estimado con el modelo de simulación.
The potential profits as a function of $F$ describe a curve concave downward (Fig. 3a); the $F_{MEY}$ is found at $F = 0.16$, and evidences a stronger effect of catch value than similar increases of the cost. The differences respect to current trend are bigger in $F \geq F_{MEY}$. On examining the cost of fishing, this variable increases nearly 4 times as a function of $F$; at $F$ values ranging from 0.05 to 0.5. At the current fishing intensity, the cost of fishing is $755$ USD per mt, and $873$ at the MSY; it is not sensible to increases in catch value (Fig. 3b). When the B/C ratio was examined, the response of this variable describes a decreasing curve respecting to $F$ ranging from 4.5 at $F = 0.5$, to 0.8 at $F = 0.5$. In this case, the 5 and 10% increase of cost and catch value display a symmetric range respecting to the mean trend (Fig. 3c).

DISCUSSION

Results show that the current status of the resource is evidence of recent efforts to maintain an adequate status of the stock of bigeye tuna in the EPO, according to the most recent assessment (Maunder & Hoyle 2006, IATTC 2010), indicating that fishing mortality is still appropriate to maintain a sustainable fishery despite increasing captures of young fish in fish aggregating devices (FSR 2010). Continuation of these regimes and forms of exploitation may be causing an evolutionary change in the size and age-structure of populations, decreasing the size and average age of capture (Shin et al. 2005), and may modify the predator-prey interactions (Froese et al. 2008) affecting the stability of the community and pelagic ecosystem where the bigeye tuna belongs (Bascompte et al. 2005). This could aggravate the potential reduction of the $MSY$ level affecting the economic value of this fishery over the long term (Trexler & Travis 2000). However, at
current levels, this fishery is economically inefficient and a significant reduction of $F$ should be applied to achieve the $MEY$. Thus, for a management regime to ensure that juveniles reach maturity to maintain the population, the current fishing condition need to be changed, because it has been shown that the size of sexual maturity is strongly correlated with growth, the maximum size, and longevity (Froese & Binohlan 2000). The simulation analysis also suggests that with reasonably low levels of fishing mortality below the $MSY$ level, the population can support a moderate exploitation of juveniles without being exhausted. Care should be taken to avoid overfishing of recruits, which could have adverse effects on the population by reducing the spawning stock (Pauly et al. 1989).

Exploitation of fisheries evidences the complexity of translating the management regulation in the traditional context of single-species view to the ecosystem-based management, because of the multiple factors involved in the problem of conservation of the integrity of ecosystem trying to achieve social and economic benefits (García et al. 2011), even if evolutionary pressure is ignored. This consideration evidences the inadequacy of the single species-based based theory to deal with complex ecosystems. Despite all of these constrains, this approach must be in use until other more effective management tools are developed.

Fluctuation in the stock biomass is probably caused by environmental changes and their effect on catchability. Because the longevity of bigeye, the numbers in the population are not likely to change dramatically over a short time as often happens with short-lived species like the sardine. The IATTC indicates that there is a recent trend of increased biomass of bigeye tuna in the EPO, which could be caused by conservation measures implemented since 2004 (FSR 2010). The implementation of some area closures has been considered feasible to reduce the catch of bigeye and thus ensure sustainable fisheries (Harley & Suter 2007), and the inclusion of amendments in the art (Maunder 2007), but according to the results of this study, it would be premature to implement those changes, and for the moment it is not perceived that this stock is threatened by over fishing.

To obtain the $MSY$, is necessary to increase the $F$ from the current value of 0.249 to 0.29, and to exploit the stock at the $MEY$, it should be necessary to reduce $F$ from 0.249 to 0.16. This reduction might be interpreted as an extreme reduction of fishing effort, but the economic returns would justify the change, because the profits would increase significantly respecting to the current $F$ or the $F_{MSY}$. Current profits are almost $58$ M USD, and at the $F_{MSY}$ they would decrease to almost $47$ M USD; however, at the $F_{MEY}$, a spectacular increase in profits of over $67$ M USD would be obtained, which could be used to invest outside this activity to provide jobs to the 665 fishermen that could have to leave this fishery. These results are consistent with Grafton et al. (2010, 2012), Kompas et al. (2010), and Kompas & Che (2012), who stated that the biomass at the $F_{MEY}$ level exceeds the biomass at the $F_{MSY}$. In the case of the big eye tuna fishery, the biomass at the $F_{MEY}$ is about 20% above the $F_{MSY}$ despite that some warnings and assumptions should be considered (Dichmont 2010, Kompas & Che 2012). These results and other trials made with the model applied to other artisanal fisheries provide similar outcome. The Christensen (2010) statement that the surplus economic benefits that can be obtained by reducing fishing intensity from the $MSY$ level to the $MEY$, can’t be enough to fulfill the need of new jobs for the fishermen left out of the fishery, would be acceptable if the stock biomass would be not threatened by overfishing. In addition, should not be forgotten that fisheries exploited at the $MSY$ level are very close of being overexploited, which is an undesirable risk and for this reason has been identified as a limit reference point (Garcia 1994, Caddy & Mahon 1995).

The IATTC has suggested the application of lower $F$ for the exploitation of bigeye tuna (FSR 2010). The results of our simulations suggest that the best harvesting strategy to adopt would be the $MEY$, which implies several benefits and a few inconvenient social costs, which after a comparison with the current condition, seem to be quite acceptable. However, if this option is adopted, it would be convenient to apply it gradually to avoid undesirable social impact of a sudden change, because it would imply a reduction of nearly 665 fishermen from the current 3,517. The economic benefit would mean a significant increase from 57.8 M USD at the current condition to 67.1 M USD at the $MEY$, showing that this surplus income could be used to finance other economic activities to provide employment for the fishermen leaving this activity. To minimize the social impact, the reduction of fishing effort should be made over several years. Taking the fishery to these levels, it is more likely to capture individuals at good sizes and there would be major economic gains in biomass (Froese 2004). An appropriate harvest regime would decrease the mortality of juveniles (Froese et al. 2008), which would reduce the impact of fishing on the population, generate higher stability in landings (Hsieh et al. 2006), and benefits to the ecosystem (Pikitch et al. 2004).
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