



Ambiente & Água - An Interdisciplinary
Journal of Applied Science

ISSN: 1980-993X

ambi-agua@agro.unitau.br

Universidade de Taubaté
Brasil

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Effects of global climate change on chlorophyll-a concentrations in a tropical aquatic
system during a cyanobacterial bloom: a microcosm study
Ambiente & Água - An Interdisciplinary Journal of Applied Science, vol. 12, núm. 3, mayo-
junio, 2017, pp. 390-404
Universidade de Taubaté
Taubaté, Brasil

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Effects of global climate change on chlorophyll-a concentrations in a tropical aquatic system during a cyanobacterial bloom: a microcosm study

doi:10.4136/ambi-agua.2014

Received: 11 Oct. 2016; Accepted: 23 Feb. 2017

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ABSTRACT

Recent studies have investigated the impact of climate change on aquatic environments, and Chlorophyll-*a* (Chl-*a*) concentration is a quick and reliable variable for monitoring such changes. This study evaluated the impact of rainfall frequency as a diluting agent and the effect of increased temperature on Chl-*a* concentrations in eutrophic environments during a bloom of cyanobacteria. This was based on the hypothesis that the concentration of Chl-*a* will be higher in treatments in which the rainfall frequency is not homogeneous and that warmer temperatures predicted due to climate change should favor higher concentrations of Chl-*a*. The experiment was designed to investigate three factors: temperature, precipitation and time. Temperature was tested with two treatment levels (22°C and the future temperature of 25°C). Precipitation was tested with four treatments (no precipitation, a homogeneous precipitation pattern, and two types of concentrated precipitation patterns). Experiments were run for 15 days, and Chl-*a* concentration was measured every five days in each of the temperature and precipitation treatments. The water used in the microcosms was collected from a eutrophic lake located in Central Brazil during a bloom of filamentous cyanobacteria (*Geitlerinema amphibium*). Chl-*a* levels were high in all treatments. The higher temperature treatment showed increased Chl-*a* concentration ($F=10.343$; $P=0.002$); however, the extreme precipitation events did not significantly influence Chl-*a* concentrations ($F=1.198$; $P=0.326$). Therefore, the study demonstrates that future climatic conditions (projected to 2100), such as elevated temperatures, may affect the primary productivity of aquatic environments in tropical aquatic systems.

Keywords: extreme events, *Geitlerinema amphibium*, primary productivity, temperature.

Efeito das mudanças climáticas globais na concentração de Chlorofila-a em um Sistema aquático tropical durante uma floração de cianobactéria: Um estudo em microcosmo

RESUMO

Estudos recentes têm investigado o impacto das mudanças climáticas em ambientes aquáticos, além disso a Clorofila-a (Clo-a) é uma variável de rápida avaliação e confiável para

o monitoramento de ambientes aquáticos. O objetivo do presente estudo foi avaliar o impacto da frequência da precipitação e do aumento da temperatura na concentração de Clo-a em um ambiente aquático eutrófico durante um período de floração de cianobactéria. As hipóteses para o presente trabalho: i) A concentração de Clo-a será maior em tratamentos em que a frequência de precipitação não é homogênea, e ii) Temperaturas mais quentes devem promover aumento na concentração de Clo-a. Foi utilizado um desenho experimental com três fatores: Precipitação, temperatura e tempo. A temperatura foi avaliada em dois tratamentos (22°C e a temperatura futura de 25°C). A precipitação foi avaliada em quatro tratamentos (ausência de precipitação, precipitação homogênea, e dois tipos precipitação concentrada). O experimento foi desenvolvido por 15 dias e a concentração de Clo-a foi mensurada a cada cinco dias para cada tratamento. Os níveis de Clo-a foram elevados em todos tratamentos, além disso, a concentração de Clo-a foi maior em tratamentos mais aquecidos (simulando cenário futuro) ($F=10.343$; $P=0.002$); entretanto, os eventos extremos de precipitação não demonstraram influência na concentração de Clo-a ($F=1.198$; $P=0.326$). Portanto, o presente trabalho demonstrou que as condições climáticas futuras (projetadas para 2100), como o aumento da temperatura, devem afetar a produtividade primária de ambientes aquáticos.

Palavras-chaves: eventos extremos, *Geitlerinema amphibium*, produtividade primária, temperatura.

1. INTRODUCTION

The study of global climate change has attracted interest from the scientific community since the 1970s (e.g., Kopec, 1971). However, it was not until the beginning of the twenty-first century that scientific publications concerning climate change increased significantly (see Nabout et al., 2012a). Various scenarios have been proposed concerning the negative effects of climate change on human and natural systems, including warming and ocean acidification (Christensen et al., 2006; Doney et al., 2009), biodiversity loss (Hoegh-Guldberg et al., 2007), loss of arable land (Nabout et al., 2012b), and changes to disease vector distributions (Paz, 2015).

In aquatic environments, the major factors impacted by increased air temperatures are changes in precipitation and wind (Nickus et al., 2010; Roland et al., 2012). These factors can cause physical changes to the environment through stratification (Jeznach and Tobiasson, 2015) or turbidity (Meerhoff et al., 2007), and chemical changes such as oxygen concentration (Gordon et al., 2004; Jeppesen et al., 2013) or nutrient cycling (Lecerf et al., 2007), and biological changes in the phenology of species (Daufresne et al., 2009). Experimental approaches have been used to evaluate the effect of climate change in aquatic environments. These experiments typically focus on the effects of the increased temperature expected in future scenarios (Jeppesen et al., 2010; Yvon-Durocher et al., 2010; Roland et al., 2012). However, recent studies have investigated other elements affecting aquatic ecosystems as well as their combined effects, such as the interaction between warming, drying, and acidification on consumers and planktonic producers (Christensen et al., 2006), the effects of climate change on water levels compounded with stratification (Berger et al., 2010), the effect of light and CO₂ enrichment on nutrient concentrations (Andersen et al., 2005), and the effect of temperature and predation on phytoplankton community (He et al. 2015).

In addition to changes in temperature, the Intergovernmental Panel on Climate Change (IPCC, 2014) predicts that tropical ecosystems will experience altered hydrological cycles with an increased frequency of extreme events such as concentrated precipitation (IPCC, 2014). Extreme rainfall in some river basins can lead to increased flooding risk (Rockström et al., 2014), changes in runoff (Roland et al., 2012), and changes in Chl-*a* concentrations (Belnap et al., 2005). Thus, extreme rainfall events can change dilutive effects and reduce zooplankton

predation, promoting an increase in Chl-*a* concentrations (see Dolan et al., 2000 for dilution effects). In addition to altered aquatic and peripheral communities, climate change can also compromise water resource quality (Codd, 2000; IPCC, 2014). Certain types of climate change can cause eutrophication that stimulates the growth of cyanobacteria facilitating the concentration of toxins (Reichwaldt and Ghadouani, 2012). Thus, studies that investigate the combined effect of climate change phenomena, such as elevated temperature and precipitation, are necessary in order to understand the changing conditions of tropical aquatic environments.

Experimental studies are therefore important tools to test the effects of climate change in aquatic ecosystem (see for example Li et al., 2016; 2017; Short et al., 2016). In fact, the majority of these experimental studies have been developed in temperate regions, and thus offer little insight about the impact of climate change in tropical aquatic ecosystems (see Roland et al., 2012). Nonetheless, experimental and whole lakes studies have shown that climate change (e.g. extreme events and warming) can affect aquatic tropical environments by modifying their geochemistry (Roland et al., 2012), altering species composition and functional groups (e.g. Costa et al., 2015), or increasing cyanobacterial dominance (Kosten et al., 2012). In fact, toxic Cyanophyceae has occurred in tropical environments (Mowe et al., 2015a).

Some studies have not found any relationship between temperature change and Chl-*a* concentrations (see Moss et al., 2003; Feuchtmayr et al., 2009). Due to the diversity of phytoplankton species, there are many optimum temperature ranges for growth (Seip and Reynolds, 1995), thus elevated temperatures may not change the size of the community, but rather species composition (Moss et al., 2003; Van de Bund et al., 2004). Other studies, however, have found associations between increased temperature and frequency of cyanobacterial bloom phenomena (Reynolds, 2006; Jeppesen et al., 2009; Tundisi et al., 2015; Mowe et al., 2015b) and increased gross primary production (Yvon-Durocher et al., 2015).

This study evaluated the impact of rainfall frequency and increased temperature on chlorophyll-*a* concentrations in eutrophic environments during a bloom of cyanobacteria. This was based on the expectations that: i) the intensity of rainfall in a short period of time would increase chlorophyll-*a* concentrations in the treatments with the most homogeneous precipitation conditions, and ii) the warmer temperatures predicted due to climate change should favor higher concentrations of chlorophyll-*a*.

2. MATERIAL AND METHODS

2.1. Experimental design

The experimental design investigated three factors: temperature, precipitation, and time. There were two temperature levels (current and future) and four levels of precipitation (no precipitation, a homogeneous precipitation pattern, and two types of concentrated precipitation patterns). Experiments were run for 15 days, and Chl-*a* concentration was measured every five days in each temperature and precipitation treatment.

Experiments were performed in microcosms. We prepared 40 beaker microcosms, each with a total volume of 2 L. Each beaker received 1 L of water initially from an artificial-landscape with a eutrophic lake located at the Universidade Estadual de Goiás (City of Anápolis, Central Brazil). Samples were gathered from the lake during a bloom of the filamentous cyanobacteria, *Geitlerinema amphibium* (Figure 1 in supplementary material S1), in May 2015. Other organisms, such as other species of algae or zooplankton adhering to the mass of *Geitlerinema amphibium* were not removed. However, the analysis of algae species presented in this lake indicated that 99% of the total biomass was represented by *G. amphibium*. Thus, the Chl-*a* analysis used in this study was represented exclusively by this species.

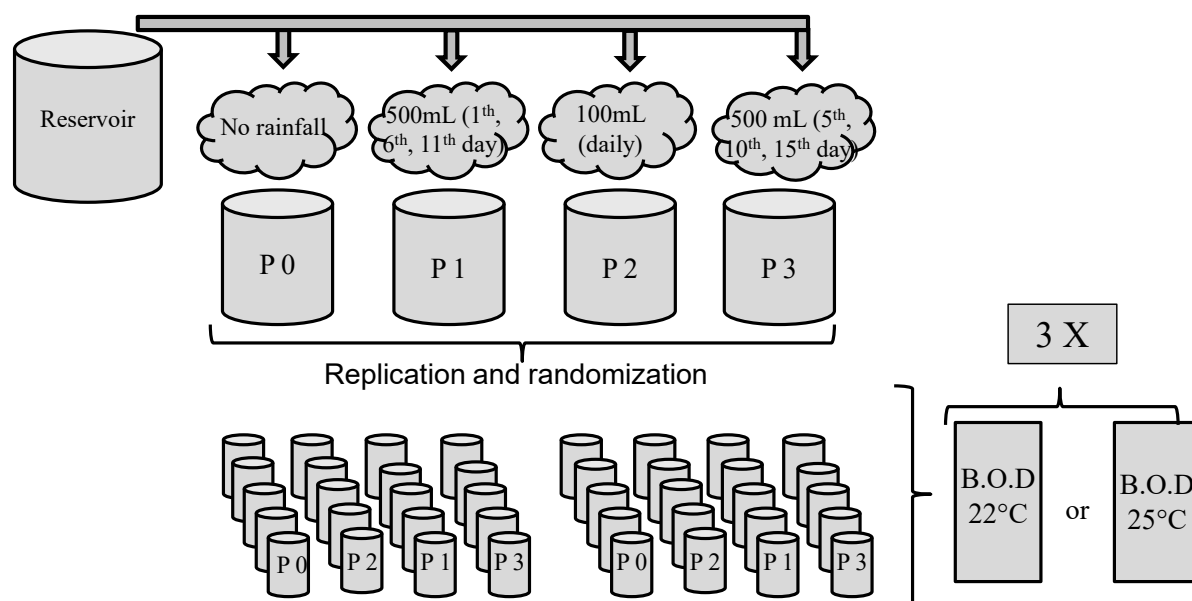


Figure 1. Schematic representation of the experiment. Water from a lake was used to fill the microcosms. The microcosm P0 received no distilled water, P1 received 500 mL of distilled water on the first day and every five days following, P2 received 100 mL of distilled water daily, and P3 received 500 mL of distilled water on the fifth day and every five days following. The experiments were placed in BOD incubators, simulating current (22°C) and future (25°C) temperatures. There were 20 replicates of each treatment of temperature and 10 replicates of each treatment of precipitation. The experiment was conducted for 15 days with samples collected every five days for Chlorophyll-*a* measurements.

We obtained climate data from the EcoClimate database (Lima-Ribeiro et al., 2015) for neotropical regions, which includes current and predicted future climate conditions under different scenarios and global circulation models. We defined temperature and precipitation values using the CCSM model (Community Climate System Model) RCP 4.5 (obtained through the EcoClimate database; Lima-Ribeiro et al., 2015). The EcoClimate database has 19 climatic variables (derived from temperature and precipitation) for worldwide, containing current and future climate data. We used a climate scenario considered to have an intermediate level of greenhouse gas emissions, projected for the year 2100. This climate scenario has been used often in global climate change research (Pendergrass and Hartmann, 2012). The annual mean temperature for the Cerrado Biome is 23.7°C, and the future climate scenarios estimate an increase of 3°C (according to the CCSM model, available in the EcoClimate database). According CCSM model the annual rainfall for the Cerrado Biome is 1170L for the current climate scenario, and 1096L in future scenarios. Although this is only a small change in rainfall, the climate models suggest an increase of extreme events, such as, concentrated rainfall in a short period of time, and long periods of drought (see Marengo et al., 2009). Thus, based on this expectation, we simulated the concentrated precipitation as an extreme event. For this, the experiments received the same amount of water, but the variance was different. Thus, the concentrated treatments presented higher variance. The variance was estimated based on the change of precipitation in future scenarios, according to the CCSM model.

We used the temperature and precipitation values of the region where the lake samples were collected (City of Anápolis, Central Brazil), and selected two climatic variables: annual mean temperature and the precipitation of the wettest quarter corresponding to our study period. These two variables were obtained for the current and future climate scenarios. The current temperature was set at 22 °C and the future temperature at 25°C. For precipitation, we simulated three conditions of rainfall (two concentrated and one homogeneous) and one control, according

to projected future scenarios (Marengo et al., 2009) Thus, for the same amount of precipitation we simulate extreme and homogeneous conditions.

Considering the climatic data of the study region, we developed the following experimental design. We simulated two temperature conditions in the microcosms; one treatment level corresponded to the current temperature of 22°C and the other corresponded to the future temperature of 25°C predicted according to CCSM RCP4.5, projected to 2100. To simulate the effect of different rainfall patterns, distilled water was added to the microcosms. There were three precipitation scenarios, and a control (P0) with no precipitation. In the first treatment (P1), 500 mL of distilled water was added to the microcosms every five days, beginning on the first day of the experiment, in order to simulate an extreme precipitation event. In the second treatment (P2), 100 mL of distilled water was added daily to the microcosms in order to simulate frequent rainfall. In the third treatment (P3), 500 mL of distilled water was added every five days beginning on the fifth day of the experiment. Treatments P1 and P3 simulated similar extreme events but with different start dates during the experiment (Figure 1).

We distributed the precipitation treatments in two BOD incubators with a photoperiod of 12 h/12 h set at the two different temperatures, 22°C and 25°C. The distribution of beakers in BOD incubators was determined randomly. Thus, each temperature treatment had 20 replicas, and each precipitation treatment had 10 replicas.

Before the experiment, we developed two series of pilot experiments, in which it was possible to determine the temporal interval of the experiment (degradation of chlorophyll-*a* occurred at more than 15 days), time among water adding, the method for adding water, time of light exposure, and other factors.

2.2. Chlorophyll-*a* analysis

The water from each microcosm was filtered through 47 mm glass fiber filters (Milipore 0.45 m) prior to chlorophyll-*a* analysis. Filters were placed in foil envelopes, stored in the dark, and kept in a refrigerator until processing. Quantitative analysis of chlorophyll-*a* was performed after acetone extraction according to the method in Golterman et al. (1978). Chl-*a* concentrations were determined once for each beaker using a volume of 190 mL.

2.3. Data analysis

The difference between treatments was determined using a factorial analysis of variance (ANOVA, $P < 0.05$) for repeated measures (Zar, 2010). Chl-*a* concentration was the dependent variable, with precipitation (at 3 levels) temperature (at 2 levels) and time as factors. Time was the total number of days in the experiment. The variables were $\log x + 1$ transformed to meet the assumptions of normality and homogeneity of variances. To verify the assumptions, we used the Kolmogorov-Smirnov (KS) test for normality and the Levene homogeneity tests for homogeneity of variance.

3. RESULTS AND DISCUSSION

Chlorophyll-*a* concentrations presented high values in all treatments (see descriptive statistic in Table 1); moreover, the value varied between treatments. Thus, after transforming the data, the assumption of normality required for ANOVA was supported, for Chl-*a* on the 5th day ($d = 0.18$; $P = 0.15$), 10th day ($d = 0.15$; $P = 0.20$) and 15th day ($d = 0.14$; $P = 0.20$). The assumption of homogeneity of variance was also supported for Chl-*a* on the 5th day ($F = 3.8$; $P = 0.06$), 10th day ($F = 1.1$; $P = 0.29$) and 15th day ($F = 0.07$; $P = 0.79$).

Table 1. Mean and standard deviation (SD) of Chlorophyll-*a* concentration (μgL^{-1}) registered in each treatment and levels in microcosm experiment.

Treatment	Levels	5th day		10th day		15th day	
		Mean	SD	Mean	SD	Mean	SD
Temperature	Current	170.8	169.8	191.8	122.0	157.3	164.3
	Future	179.9	117.6	223.1	161.2	706.5	1235.9
	P1	195.4	174.0	330.5	175.8	727.9	957.5
Precipitation	P2	152.5	48.2	144.0	86.3	235.8	308.8
	P3	240.4	213.4	178.9	70.7	624.4	1524.4
	P4	113.1	44.4	176.5	144.0	139.6	66.1

Chl-*a* concentrations varied significantly with temperature (Table 2; Figure 2). Increased temperature caused a significant increase in Chl-*a* concentrations, but precipitation did not cause any significant difference in Chl-*a* concentrations (Figure 3).

Table 2. Statistical significance of the treatment effects on Chlorophyll-*a* concentration, considering the interaction of precipitation, temperature, and time, through repeated analysis of variance. Significant values are in bold ($P < 0.05$).

	Degrees Freedom	F	P
Temperature (Temp)	1	10.343	0.002
Precipitation (Prec)	3	1.198	0.326
Temp X Prec	3	0.705	0.556
Time	2	0.241	0.786
Time X Temp	2	3.256	0.044
Time X Prec	6	0.634	0.702
Time X Temp X Prec	6	1.065	0.393

The interaction between time and temperature was also significant. The third week of the experiment (15 days), there was an increase in the concentration of Chl-*a* in treatments simulating future climate change scenarios (Figure 4). There was no significant interaction between time and precipitation. However, at elevated temperatures, time has an important role in determining concentrations of Chl-*a*.

We used experimentally manipulated microcosms to evaluate the effects of precipitation (i.e. rainfall) and temperature on the Chlorophyll-*a* concentration of a sampled phytoplankton community collected from a eutrophic environment with an ongoing cyanobacterial bloom (*G. amphibium*). Although previous multi-factorial studies have investigated the impacts of climate change, this is the first study to assess the combined effects of temperature and precipitation in tropical environments. We found that temperature was the factor responsible for increased Chl-*a* concentrations in future climate change scenarios.

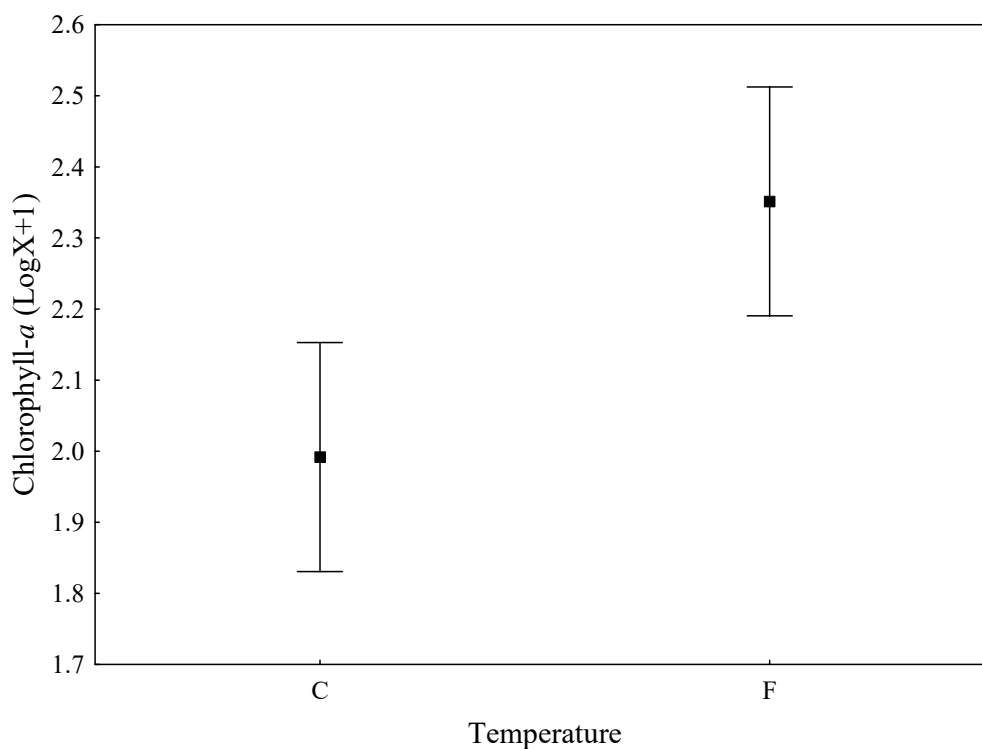


Figure 2. Boxplot (mean \pm 95% confidence intervals) representing the concentration of Chlorophyll-*a* in both simulated temperature conditions, where C is the current temperature and F the future temperature.

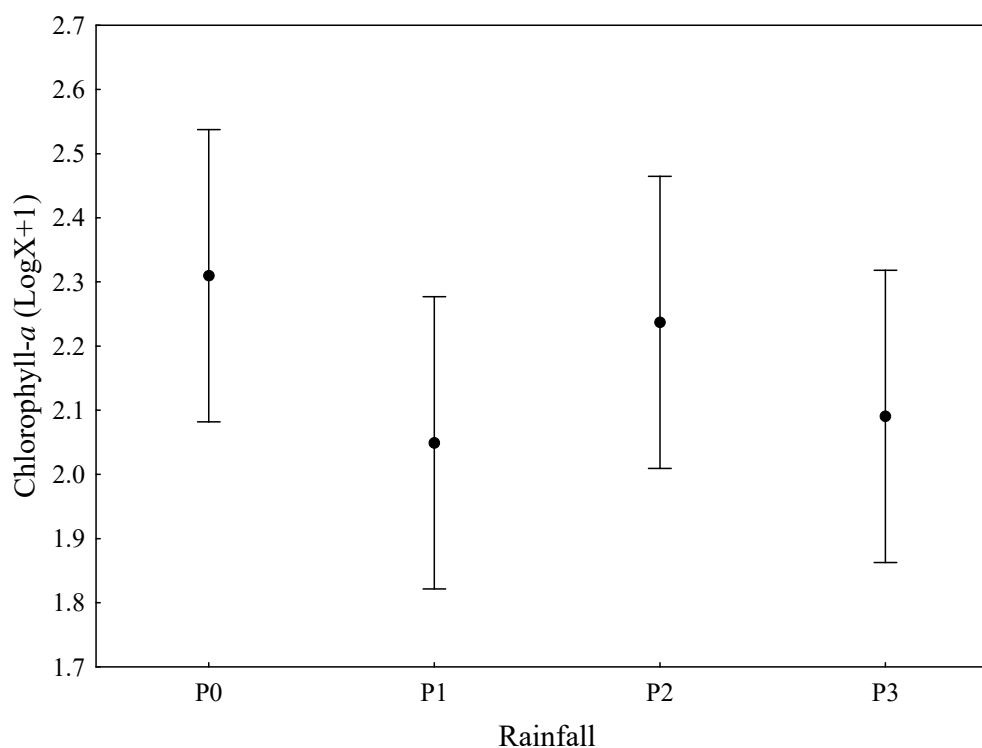


Figure 3. Boxplot (mean \pm 95% confidence interval) representing the concentration of Chlorophyll-*a* in the four simulated rainfall conditions (see details of P0, P1, P2, and P3 in methods).

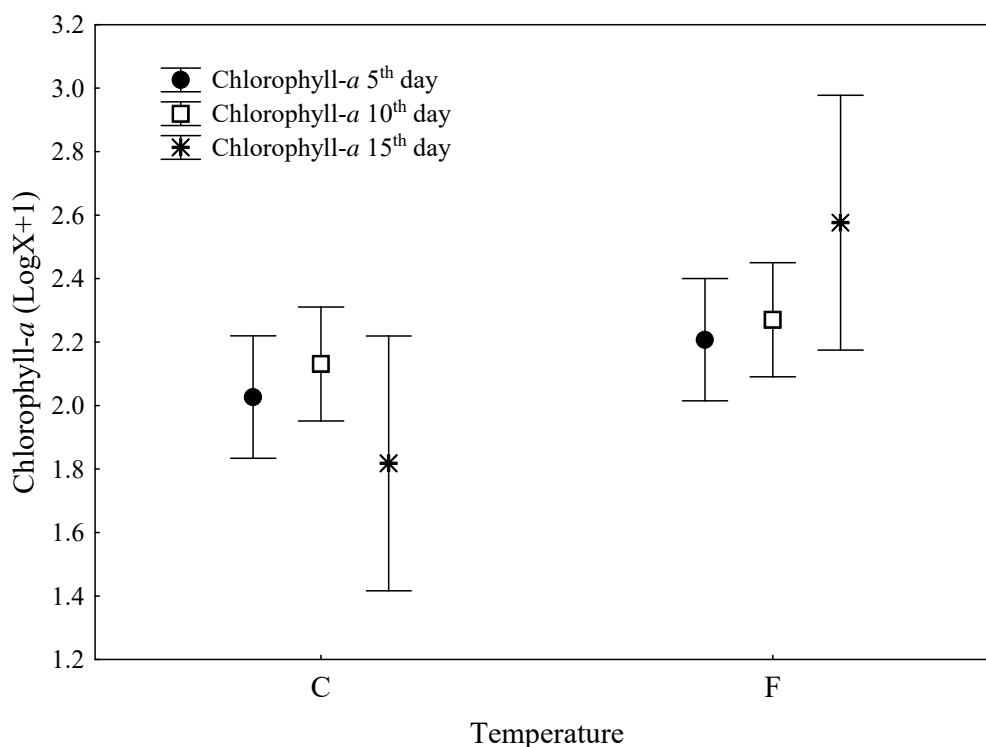


Figure 4. Boxplot (mean \pm 95% confidence interval) representing the chlorophyll-*a* concentration and the interaction effect among Temperature (C is current and F is Future) and time.

In fact, both thermal and hydrological changes can have serious consequences for aquatic ecosystems (Floury et al., 2012). Water temperature influences many processes in aquatic environments, including dissolved oxygen concentration (Gordon et al., 2004; Jeppesen et al., 2013), conductivity (Esteves et al., 2011), nutrient concentration (Jeppesen et al., 2009; 2011), and rates of primary productivity and decomposition (Lecerf et al., 2007).

Primary producers in aquatic environments are primarily controlled by nutrient availability, light (Huszar et al., 2006; Roland et al., 2012), and temperature (Roland et al., 2012). Temperature changes directly affect photosynthesis because higher temperatures accelerate the enzymatic reactions during photosynthesis and result in higher rates of primary productivity (Fernandes et al., 2005). Thus, algae growth tends to be linked with temperature, being faster at higher temperatures (Fernandes et al., 2005). Many cellular processes of algae depend on this variable (Reynolds, 1984). Moreover, many species of cyanobacteria are favored by an increase of temperature, promoting bloom events (see Moss et al., 2011; Jeppesen et al., 2013; 2014).

An increase of water temperatures in the future may favor not only higher growth rates but may also increase the activity (Paerl and Huisman, 2009) and geographical distribution of cyanobacteria (Briand et al., 2004; Paerl and Huisman, 2009). Cyanobacterial blooms will likely be more frequent in future climatic conditions (Romo et al., 2005; Huber et al., 2012). This might be problematic because some species of cyanobacteria can produce toxic substances (Stewart et al., 2007; Paerl and Huisman, 2009) that can directly or indirectly contaminate other aquatic organisms through the food chain (Smith et al., 2008).

These blooms can also compromise the integrity of aquatic ecosystems by promoting massive oxygen consumption by algae and bacterial respiration during decomposition (Fernandes et al., 2005). In addition, cyanobacteria may be able to raise the local water

temperature through light absorption, increasing their ability to dominate aquatic ecosystems (Paerl and Huisman, 2008).

In this study, precipitation does not affect the concentration of Chl-*a*; however, it was expected that concentrated rainfall can result in the dilution of nutrients and the washing away of algae blooms (Reichwaldt and Ghadouani, 2012) due to high discharge rates (Bouvy et al., 2003). In fact, the treatments (P1, P2, P3) presented higher Chl-*a* concentrations than P0 (no precipitation) (see Figure 3). However, the possible nutrient dilution effect was compensated by the increase in temperature, promoting the growth of Chl-*a* concentrations. Moreover, in microcosm the rainfall not remove the algae bloom (because of the physical limitations of the experiment); thus, even with the precipitation the original mass of the species was allowed to remain in the experiment, not being carried away. Finally, the addition of a large amount of water may interrupt the bloom and cause a lag in the resurgence of the algae bloom (Ahn et al., 2002), explaining the increase of Chl-*a* in some treatments during the third week of the experiment.

However, it is still a challenge to provide precise answers regarding the consequences of climate change in aquatic ecosystems because there are a variety of Atmosphere-Ocean General Circulation Models (AOGCM), as well as climate scenarios (e.g. RCP 2.6, RCP 4.5, RCP6 and RCP 8.5). Current climate projections have a high degree of uncertainty, and this impedes the study of potential future impacts of warming. Microcosm experiments are one way to overcome this problem. This type of experiment provides important information about large-scale processes while only requiring minimal time to replicate different systems and scales (Benton et al., 2007). The use of microcosms in research is also supported by the need to tightly control conditions in order to correlate community dynamics with environmental variability. Future studies should consider the indirect effects of climate change, such as runoff, that are strongly influenced by the intensity of rain. Additionally, future studies investigating water temperature effects should also evaluate the uncertainty involved in choosing different climate change scenarios.

4. CONCLUSIONS

Based on our results, we reject our first hypothesis that extreme precipitation events increase Chl-*a* concentrations relative to homogeneous rainfall. However, it is important to consider that in this study we evaluated the effect of precipitation on the dilution of the Chl-*a*. Our results supported our second hypothesis; higher temperatures favor higher Chl-*a* concentrations. The future temperature used in this study (RCP 4.5) is based on intermediate greenhouse gas emissions, which would cause changes in the primary productivity of eutrophic environments.

This study also adds more information to a number of recent studies that investigate the impact of climate change on aquatic environments. In addition, the use of microcosm has proven to be satisfactory. Therefore, new studies using a similar approach may be undertaken, and different factors may be explored, such as new climatic scenarios and climatic variables.

5. ACKNOWLEDGEMENTS

We thank L.C.G. Vieira and F.B. Teresa for suggestions on early versions of the experimental design and manuscript. We thank the two anonymous reviewers whose suggestions helped to improve the manuscript. We thank the CNPq (Process 473730/2013-8), CAPES (Process 2036/2013) and FAPEG (Process 201210267001071) for their financial support. Meirielle Euripa Pádua de Moura and Lorraine dos Santos Rocha thank Universidade

Estadual de Goiás (UEG) and CNPq, respectively, for their scholarships. João Carlos Nabout thanks CNPq (Process 309700/2013-2) for the “*Produtividade em Pesquisa*” Scholarship.

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